



**Cherry Creek Basin Water Quality Authority
Technical Advisory Committee Meeting Agenda
Thursday, January 4, 2024, 9:00 a.m.**

In-person attendance is encouraged due to audio limitations in the meeting room.

**In-Person: SEMSWA
7437 S. Fairplay St.
Centennial, CO 80112**

**Virtual: Zoom
<https://us06web.zoom.us/j/87425775963> Passcode: CCBWQA
Phone (646)931-3860 Mtg ID: 874 2577 5963# Passcode: 815374**

TAC Meeting Documents can be found online at the link below.

<https://drive.google.com/drive/folders/12BoEhmFbnnMCxivnpjY2l7T5TzP8Azlq?usp=sharing>

1. Call to Order (9:00) (5 minutes)
2. Meeting Minutes from December 7, 2023 (enclosed)
3. Action Items (none)
4. Discussion Items (9:05) (90 minutes)
 - a. TAC Vision for 2024 (Knerr, 30 minutes)
 - i. TAC Overview and Role
 - ii. Subcommittees
 - iii. Open Discussion and Input
 - b. Draft Timeline and Activities for 2024 (Clary, 15 minutes)
 - c. WY 2023 Monitoring Report (Provisional Draft) (Stewart, 30 minutes, [enclosed](#))
 - d. Modeling Subcommittee Update (Alan Leak, 15 minutes, enclosed)
5. Presentations (none)
6. Updates (10:35) (25 minutes)
 - a. Manager (Clary)
 - i. Support Letter for USACE Project (enclosed)
 - ii. Regulation 72 Update ([Responsive Prehearing Statements - Google Drive](#))
 - iii. Lake Loop Shoreline Stabilization
 - iv. Lakeview Drive Road Repairs
 - v. Peoria Pond O&M Plan
 - vi. PAPM Recommendation
 - vii. Governor Appointees to the Board
 - viii. July TAC Meeting (Currently on July 4th)
 - b. Cherry Creek Stewardship Partners (Davenhill)
 - c. TAC Members (As Needed)
 - d. TAC Subcommittees (As Needed)
 - i. Modeling Subcommittee
 - ii. Watershed Plan Subcommittee
 - iii. Cherry Creek Reservoir to Lakeview Drive Alternatives Analysis Subcommittee
 - iv. Lone Tree, Windmill, and Cottonwood Creek Subcommittee
 - e. Contractors (As Needed)
 - i. [Water Quality Update](#) (Stewart)
 - ii. Pollution Abatement Projects (see Manager update)
 - iii. In-Park PRF and RDS Maintenance and Operations (Goncalves)
 - iv. Regulatory (DiToro)
 - v. [Land Use Referral Tracking](#) (Endyk)
7. Adjournment

[Board Binder](#)



**Cherry Creek Basin Water Quality Authority
Minutes of the Technical Advisory Committee Meeting
Thursday, December 7, 2023, 9:00 a.m.**

TAC Members Present

Alex Mestdagh, Town of Parker
Ashley Byerley, SEMSWA
Caitlin Gappa, Board Appointee, Douglas County Health Department
Casey Davenhill, Board Appointee, Cherry Creek Stewardship Partners
Cayla Cappello, City of Greenwood Village
David Van Dellen, Town of Castle Rock
Diana Rashash, Board Appointee, Arapahoe County Public Health (zoom)
Jacob James, City of Lone Tree
James Linden, SEMSWA - Alternate (zoom)
Jim Watt, Board Appointee, Mile High Flood District
Joseph Marencik, City of Castle Pines (zoom)
Jon Erickson, TAC Chair, Board Appointee, Colorado Parks and Wildlife
Lisa Knerr, TAC Vice Chair, Arapahoe County (zoom)
Rick Goncalves, Board Appointee
Ryan Adrian, Douglas County (zoom)

Board Members Present

Bill Ruzzo, Assistant Secretary, Governor's Appointee (zoom)
Tom Downing, Governor's Appointee (zoom)

Others Present

Alan Leak, RESPEC
Chris Olson, Wright Water Engineers (zoom)
Erin Stewart, LRE Water
Jane Clary, Wright Water Engineers, CCBWQA Technical Manager
Jessica DiToro, LRE Water
Richard Borchardt, R2R Engineers
Val Endyk, CCBWQA

1. Call to Order

Jon Erickson called the meeting to order at 9:00 am. Laura Kindt with Castle Rock Water has taken a new position and will not replace David VanDellen on the TAC.

2. Meeting Minutes from November 2, 2023

David VanDellen moved to approve the Nov 2, 2023 meeting minutes. Seconded by Jacob James. The motion carried.

3. Highlights from the November 16, 2023 Board Meeting

Jane Clary provided an update on actions taken at the November 16, 2023 Board meeting. Minutes from the meeting can be found [here](#).

4. Action Items

a. Acceptance of Project Summary Reports

- i. [Happy Canyon Creek upstream of I25](#)
- ii. [Dove Creek from Otero Avenue to Chambers Road](#)

Rich Borchardt provided the TAC with a [memo](#) explaining that CCBWQA and its partners completed stream reclamation on Happy Canyon Creek upstream of I-25 and Dove Creek Phase 1 from Otero Avenue to Chambers Road projects in 2023. Rich presented the project summaries and [drone footage](#) of the completed Happy Canyon Creek project. The project summaries describe the background and purpose, existing conditions, design approach, construction, funding, and water quality benefits of each project. Once CCBWQA's TAC and Board accept the project summaries, they will be included in CCBWQA's 2023 Annual Report.

Lisa Knerr moved to accept the project summaries for stream reclamation on Happy Canyon Creek upstream of I-25 and Dove Creek Phase 1 from Otero Avenue to Chambers Road. Seconded by Diana Rashash. The motion carried.

The TAC acknowledged Rich Borchardt's service to the CCBWQA and everyone had an opportunity to provide personal comments and thanks.

5. Discussion Items

a. 2024 TAC Chairman and Vice Chairman Positions

Jon Erickson led a discussion with the TAC regarding the Chair and Vice Chair positions for 2024.

Jacob James moved to amend the agenda to move this to an action item to vote on 2024 TAC positions. Seconded by Alex Mestdagh. The motion carried.

Rick Govcalvez moved to nominate Lisa Knerr as the Chair for 2024 and Ashley Byerly as the Vice Chair. Seconded by Cayla Cappello. The motion carried.

b. 2024 TAC Appointments

Jon Erickson explained that every year the TAC reviews its current members and discusses recommendations for Board-appointed TAC members for the upcoming year. Val Endyk provided a current [list](#) of TAC members for review and discussion purposes.

The TAC recommended that the Authority's administrative assistant reach out to current Board-appointed TAC members to gauge interest in continuation on the TAC.

c. Lone Tree Creek Master Drainage Plan Update

Jane Clary provided the TAC with a [memo](#) from Maggie Lewis and Andrew Earles with Wright Water Engineers summarizing the results of the November 30, 2023 Subcommittee meeting regarding the Major Drainageway Plan for the Lone Tree, Windmill, and Cottonwood Creek watersheds. The Subcommittee discussed hydrologic model methods and results, a public and stakeholder outreach strategy, and the selection of project alternatives to recommend to the TAC. From that discussion, five alternatives were selected for inclusion as recommendations in the Master Drainageway Plan corresponding to the locations on the figure in [Attachment 1](#).

The TAC was also given a copy of the [PowerPoint](#) from the November 30th meeting.

Next steps:

- A draft MDP report will be provided to the TAC by December 31, 2023.
- Comments on the MDP will be due by January 31, 2024.

- The final report will be completed by March 31, 2024.

d. Annual WQ Monitoring Report Status Update

Erin Stewart provided the TAC with a [memo](#) identifying factors that will impact the monitoring report schedule. Due to factors beyond our control, some of the biological monitoring information used to evaluate plankton dynamics analyses will not be available for a few months. In addition some of the flow information that is used for water balance calculations was lost due to flooding. USACE storage information to estimate flows, but this information will not be available until late 2023 or early 2024, which will restrict the completion of certain portions of the draft of the WY 2023 Monitoring Report by the typical deadline. Reg 72 reporting requirements will still be met on schedule. An amended report with the additional information will be provided as soon as it is available, estimated April 2024.

e. USACE Reservoir Release Proposal

Katie Seefus provided the TAC with a [memo](#) regarding the proposal for the Sustainable Rivers Program from the Omaha District of the United States Army Corps of Engineers.

Erin Stewart [presented](#) the slides provided by the USACE regarding the proposal for a low-level release strategy for improving reservoir water quality.

Discussion included:

- CPW's feedback from a water rights perspective.
- Evaluation of downstream water quality impacts, if any, should be addressed as part of project monitoring.

Jane Clary will coordinate with the Executive Committee and staff to provide a letter to USACE regarding the proposal.

f. Regulation 38 Site Specific Standards Letter to CDPHE and Updated Hydros Technical Memorandum (Clary/Hawley, enclosed)

Jane Clary provided the TAC with the [letter](#) to the WQCC stating that the Authority may propose site-specific total phosphorus and total nitrogen standards for Cherry Creek Reservoir at the June 2025 Regulation 38 Rulemaking Hearing with a delayed effective date after 12/31/2027.

Jane also provided the updated Hydros [memo](#) to the TAC, providing the most up-to-date version of Hydros' analysis.

6. Presentations

a. Runoff Reduction Study Update

Chris Olson with Wright Water Engineers provided an [update](#) on the runoff reduction monitoring for receiving pervious areas study.

Highlights from the project update included an overview of the site selection and monitoring plan development, data collection, and data analysis and reporting.

Chris provided photos from the monitoring sites within the State Park and graphs detailing the grass swale runoff reduction results. Also highlighted were the 17-Mile House infiltration basin preliminary results and the two Simulated Runoff Tests (SRTs) completed at Parker Town Hall. A draft report will be provided to the TAC in early 2024.

7. Updates

a. Cherry Creek Stewardship Partners (Davenhill)

Casey Davenhill provided an update to the TAC on the plans for the mural at the new educational building at Cherry Creek State Park and discussed interpretative signage opportunities for water quality at the State Park. Casey would like to request funding from the Cherry Creek Basin Water Quality for the new mural, which is estimated to cost \$15,000.

b. TAC Members

c. TAC Subcommittees

i. Modeling Subcommittee

Alan will have updated HSPF watershed model runs by the end of the year.

ii. Watershed Plan Subcommittee

Ongoing progress is being made on the watershed plan update.

iii. Cherry Creek Reservoir to Lakeview Drive Alternatives Analysis Subcommittee

Muller is working on the alternatives analysis.

d. Contractors

i. [Water Quality Update](#) (Stewart)

ii. Pollution Abatement Projects

a. [CIP Status Report](#) (Borchardt, enclosed)

Staff met with USACE (along with CPW) to discuss requirements for their 408 review. These requirements impact contractors who work within the State Park. Staff is working on a memo detailing the requirements and will provide that to the TAC when completed.

CPW has nearly completed the repairs to Lakeview Drive and implemented improvements in water quality during construction.

b. Wetland Harvesting (Stewart)

Erin will provide a full report in early 2024.

iii. In-Park PRF and RDS Maintenance and Operations (Goncalves)

iv. Regulatory (DiToro)

v. [Land Use Referral Tracking](#) (Endyk)

e. Manager (Clary)

i. CU-Boulder Landscape Transformation Proposal—Jane Clary provided a letter of support on behalf of CCBWQA related to a grant application prepared by Dr. Aditi Bhaskar to the Colorado Water Conservation Board. This study may help to fill data gaps related to water quality benefits associated with turf conversion to alternative landscape types (e.g., Coloradoscape).

ii. Peoria Pond Shared Maintenance Plan—Jane Clary reported that a revised draft shared maintenance plan for Peoria Pond will be circulated later in December to SEMSWA, Greenwood Village and MHFD. Tim Flynn will be preparing a draft IGA.

iii. PAPM RFQ ([New! RFQ for Pollution Abatement Project Manager - Cherry Creek Basin Water Quality Authority](#))--The RFQ for PAPM has been posted on the project website, with proposals due by December 13 and a selection process led by the Executive Committee to follow.

iv. July TAC Meeting (Currently on July 4th)

Additionally, Jane noted that the Reg 72 hearing documents and PWSD's and Castle Rock's proponent's prehearing statement and supporting exhibits can be viewed [here](#).

8. Adjournment

Jon Erickson adjourned the meeting at 11:00 am.

[Board Binder](#)



TECHNICAL MEMORANDUM

Date: December 23, 2023
To: Cherry Creek Basin Water Quality Authority TAC
Jane Clary, CCBWQA Technical Manager
From: Erin Stewart, LRE Water
Subject: WY 2023 Monitoring Report Provisional Draft

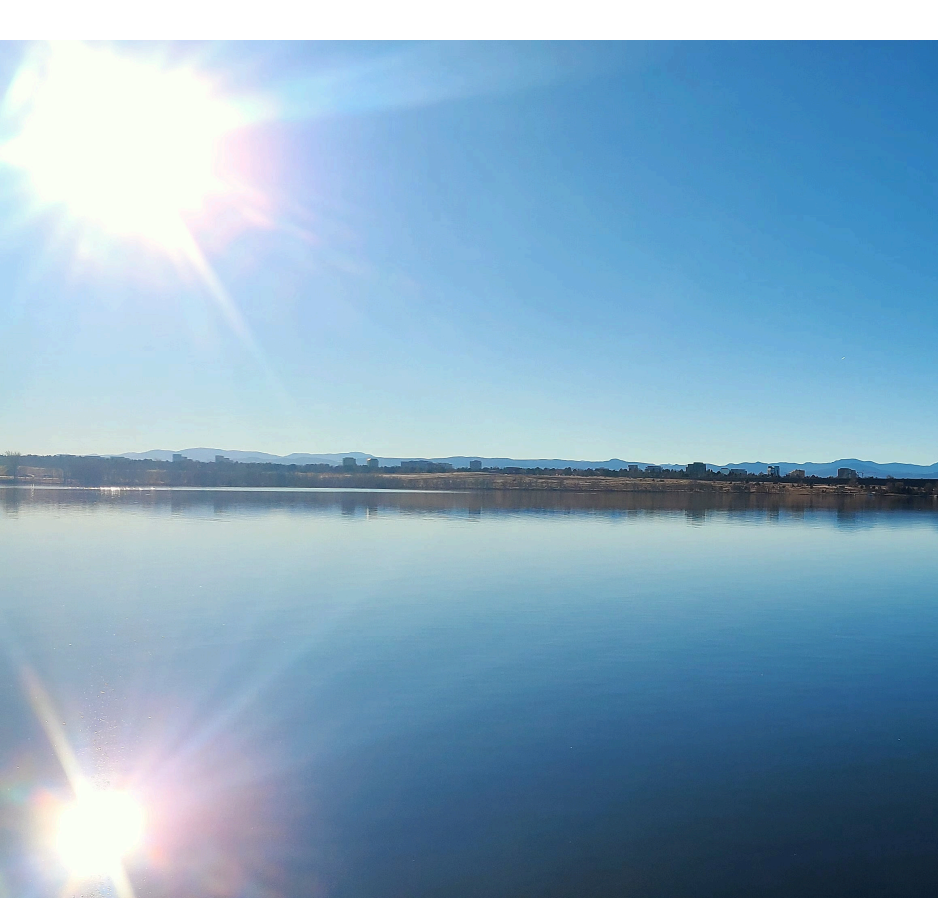
The provisional version of the WY 2023 Monitoring Report can be found at this link:

[WY 2023 Monitoring Report - PROVISIONAL](#)

Please note that this is NOT a review draft but a provisional copy for the TAC to provide general context and familiarity with the content and analysis that will be presented at the January 4, 2024 meeting. Questions related to this provisional draft can be discussed at the January presentation to the TAC, with additional opportunity for input prior to acceptance of the report at the February TAC Meeting.

Please note that some data and measurements normally collected under the monitoring program are not available for WY 2023 due to factors outside of the CCBWQA's control including damage to monitoring equipment and lost data due to excessive precipitation and associated flooding. Alternative calculations using the relative inflows of Cherry Creek and Cottonwood Creek and storage information from the USACE will be used and provided with the amended report later in 2024 to update pollutant load-related information.

Once the additional analysis is completed, the draft will be amended and provided as a final draft for comprehensive review.



CHERRY CREEK BASIN WATER QUALITY AUTHORITY MONITORING REPORT WATER YEAR 2023



SUBMITTED TO:
**Cherry Creek Basin Water
Quality Authority**
PO Box 3166
Centennial, CO 80161

PREPARED BY:
LRE Water
1221 Auraria Parkway
Denver, CO 80224



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PROVISIONAL

ACRONYMS/ABBREVIATIONS

Acronym	Definition
AF	Acre-feet
AOAC	Association of Official Analytical Chemists, now AOAC INTERNATIONAL
ASTM	American Society for Testing and Materials
Authority	Cherry Creek Basin Water Quality Authority
BMPs	Best Management Practices
CCBWQA	Cherry Creek Basin Water Quality Authority
CCR	Code of Colorado Regulations
CCSP	Cherry Creek State Park
CDPHE	Colorado Department of Public Health and Environment
Cells/mL	Cells per milliliter (phytoplankton)
CPW	Colorado Parks and Wildlife
CFR	Code of Federal Regulations
cfs	Cubic feet per second
chl α	Chlorophyll α
CM	Control Measures
CR72	Cherry Creek Reservoir Control Regulation 72
DM	Daily Maximum (for Temperature)
DO	Dissolved Oxygen
DOC	Dissolved Organic Carbon
EPA	U. S. Environmental Protection Agency
IEH	IEH Laboratories
M	Meters
mg/L	Milligrams per liter
mV	Millivolts
$\mu\text{g/L}$	Micrograms per liter
Mi	Mile
μm	Micrometers
$\mu\text{m}^3/\text{mL}$	Cubic micrometers per milliliter
$\mu\text{S/cm}$	Microsiemens per centimeter
MS4	Municipal Separate Storm Sewer System
MWAT	Maximum Weekly Average Temperature
N	Nitrogen
N:P	Nitrogen to Phosphorus Ratio
NOAA	National Ocean and Atmospheric Administration

Acronym	Definition
ND	Non-detect
NH ₃ -N	Ammonia Nitrogen
NO ₃ +NO ₂ -N	Nitrate plus Nitrite Nitrogen
#/L	Number of animals per liter (zooplankton)
ORP	Oxidation Reduction Potential
%	Percent
POR	Period of record
PRF	Pollutant Reduction Facility
PRISM	Parameter-elevation Regression on Independent Slopes Model
QA/QC	Quality Assurance/Quality Control
QAPP	Quality Assurance Project Plan
Reg 31	WQCC Regulation No. 31
Reg 38	WQCC Regulation No. 38
SAP	Sampling and Analysis Plan
Reservoir	Cherry Creek Reservoir
SM	Standard Methods
SRP	Soluble Reactive Phosphorus
TDN	Total Dissolved Nitrogen
TOC	Total Organic Carbon
TN	Total Nitrogen
TDP	Total Dissolved Phosphorus
TP	Total Phosphorus
TSI	Trophic State Index
TSS	Total Suspended Solids
TVSS	Total Volatile Suspended Solids
USACE	U.S. Army Corps of Engineers
USGS	U.S. Geological Survey
VSS	Volatile Suspended Solids
WY	Water Year
WQCC	Water Quality Control Commission
WWTP	Wastewater Treatment Plant

EXECUTIVE SUMMARY

The Cherry Creek Basin Water Quality Monitoring Report – Water Year 2023 provides a comprehensive description of monitoring completed for the Cherry Creek Basin Water Quality Authority (CCBWQA) of Cherry Creek Reservoir (Reservoir) and watershed for the 2023 Water Year (WY 2023) between October 1, 2022 and September 30, 2023. The Reservoir and watershed monitoring programs are completed in accordance with the Cherry Creek Sampling and Analysis Plan (SAP), Quality Assurance Project Plan (QAPP), and regulatory requirements. The data were collected to evaluate how successful the requirements specified in Cherry Creek Reservoir Control Regulation 72 (CR 72) are at achieving the chlorophyll- α (chl α) water quality standard and the water quality standards for associated parameters as outlined in Water Quality Control Commission (WQCC) Regulation No. 31 (Reg 31) and Regulation No. 38 (Reg 38), as directed by the CCBWQA's Statute. The program includes regular monitoring of biological, physical, and chemical conditions of the Reservoir, the streams and tributaries that feed the Reservoir, and precipitation and groundwater in the basin. Highlights of the findings from the monitoring completed during WY 2023 in relation to water quality standards, beneficial uses, and other notable details are outlined in the Executive Summary below. All CCBWQA data can be accessed at the CCBWQA's data portal at <https://www.ccbwqportal.org/>.

Please note that some data and measurements normally collected under the monitoring program are not available for WY 2023 due to factors outside of the CCBWQA's control including damage to monitoring equipment and lost data due to excessive precipitation and associated flooding. Alternative calculations using the relative inflows of Cherry Creek and Cottonwood Creek and storage information from the USACE will be used and provided with the amended report in 2024 to update pollutant load-related information.

STANDARDS

Regulation 38 (Reg 38) assigns water quality standards for Cherry Creek Reservoir to protect aquatic life and other beneficial uses. Cherry Creek Reservoir did not meet the chl α standard of 18 $\mu\text{g/L}$ established in Reg 38 in WY 2023 (Figure ES-1), although concentrations were lower than the three prior years, despite much higher phosphorus loading associated with major flood events. Cherry Creek Reservoir met the standards for temperature, pH, and dissolved oxygen (DO), which are protective of the Class 1 Warm Water Aquatic Life use.



Figure ES-1. Seasonal Chl α concentrations in Cherry Creek Reservoir

RESERVOIR HIGHLIGHTS

The water quality in Cherry Creek Reservoir during WY 2023 was atypical due to the well-above-average precipitation in the spring and flooding that occurred on both Cherry Creek and Cottonwood Creek. Following the multi-day storm during May 11th and 12th, the Reservoir elevation increased by almost 10 feet and remained above pool, or normal operating elevation for the remainder of the season. The benefit of the increased inflow and precipitation were that the cooler water and increased water exchange through the Reservoir kept the algal blooms at bay in the early summer. However, in July, as soon as the precipitation tapered to more average

levels and the temperatures warmed, the high nutrient concentrations present from the storm nutrient loads contributed to increased algal growth, chl-a concentrations, and cyanobacteria blooms.

Although the Reservoir met the DO standard, low DO concentrations were present at and near the bottom of the Reservoir during the warm summer months, that increase the potential for internal loading of phosphorus from the sediments due to anoxic conditions.

The seasonal phosphorus concentrations exceeded the interim nutrient criteria adopted by the WQCC in 2012 as well as the phosphorus standard that will be adopted statewide in lakes and reservoirs unless are adopted by 2027 (Figure ES-2). Although the seasonal nitrogen in the Reservoir was below the 2012 nutrient criteria, it exceeded the nitrogen standard that could be adopted in 2027¹.

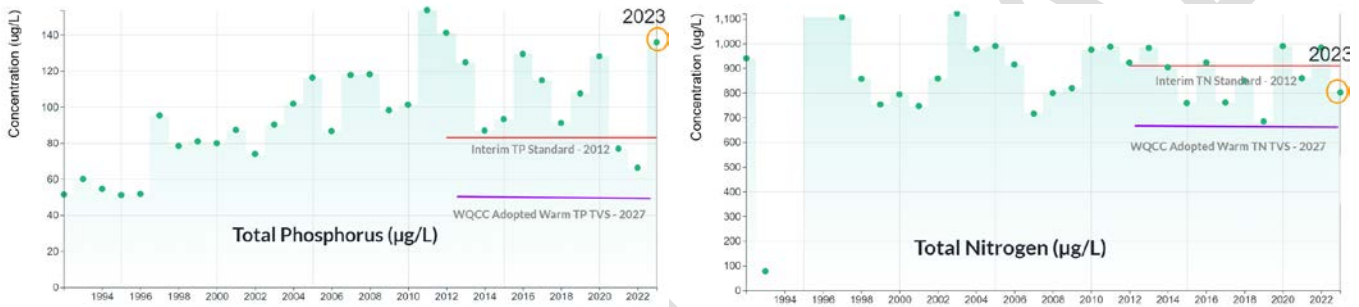


Figure ES-2. Seasonal Phosphorus and Nitrogen Concentrations in Cherry Creek Reservoir

The Trophic State Index (TSI) is a relative expression of the biological productivity of a lake using total phosphorus (TP), chl α , and transparency. The WY 2023 TSI for Cherry Creek Reservoir indicates that Cherry Creek Reservoir continues to be classified as eutrophic based on water transparency and chl α concentrations and hypereutrophic based on TP concentrations (Figure ES-3). Eutrophic and hypereutrophic conditions indicate elevated nutrient concentrations and often excessive productivity with increased probabilities of encountering nuisance algal blooms. Although there has been some fluctuation of the historical trophic state, Cherry Creek Reservoir has remained in the eutrophic to hypereutrophic range for the last 20+ years.

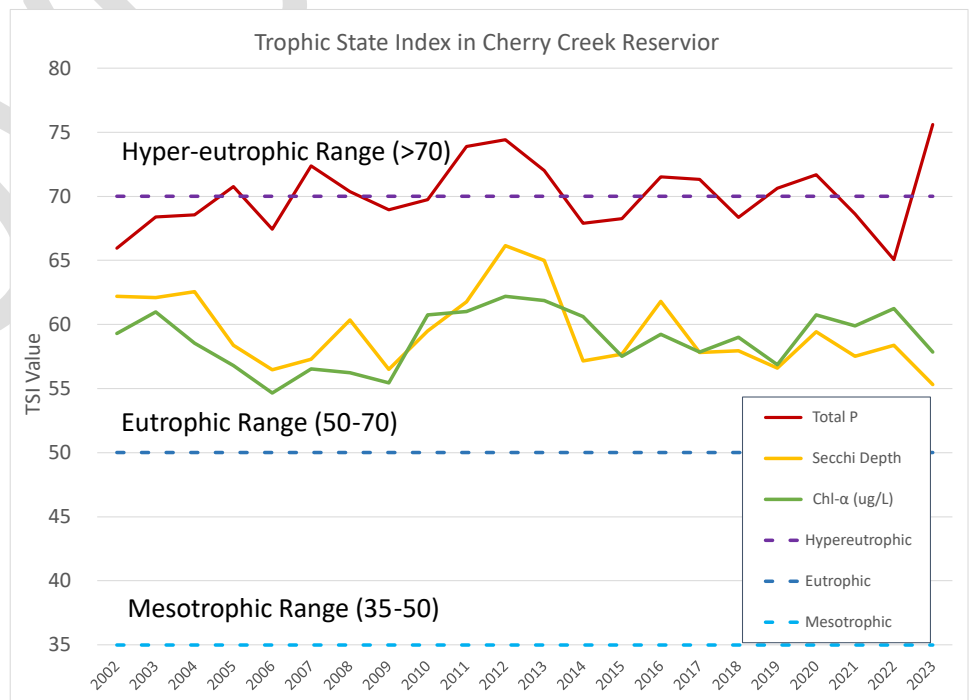


Figure ES-3. Trophic State in Cherry Creek Reservoir

In mid-July, a cyanobacteria bloom prompted Colorado Parks and Wildlife to post “Caution” signs to inform the public of the potential recreational risk. Ongoing monitoring detected toxin concentrations above the recreational threshold and a closure was implemented on July 28th in

¹ The Reservoir may adopt site-specific standards that differ from the statewide standards.

the area of the bloom and “Danger” signs were posted. Less than a week later, toxin levels had decreased to below the recreational threshold and the closure was lifted on Aug 4th and by mid-August the bloom had dissipated.

WATERSHED HIGHLIGHTS

The spring of WY 2023 received much higher-than-average precipitation, that caused major flooding and damage along Cherry Creek and Cottonwood Creek. Multiple monitoring stations, equipment, and other

The watershed received approximately 172% of the historical average precipitation in WY 2023, with 15-16” in May and June accounting for over 60% of the entire year.



infrastructure were damaged and required repair. The extended elevated water levels and equipment damage also impacted stream flow calculations.

In WY 2023, the Cherry Creek State Park (CCSP) meteorological station measured a total of 22.3 inches of precipitation. NOAA’s Centennial Airport weather station KAPA site measured 25.6 inches, which is 172% of the historical average.

The WY 2023 median TP concentrations were higher in storm flows (💧) than baseflows, as usual. Median TP

concentration were lower than the baseline (long-term) median at all sites under both baseflow and storm flow conditions except for the two sites on Cherry Creek (CC-7 and CC-10) during storm conditions and at the outlet to the Reservoir (CC-0,) which were higher in WY 2023 (Figure ES-4; See Figure 2 for monitoring locations). The higher TP concentrations in WY 2023 can be attributed to the major storm events that caused above-average concentrations of suspended solids and phosphorus.

In contrast to TP, relatively higher total nitrogen (TN) concentrations were not consistently observed during storm events. The WY 2023 median TN concentrations were higher than the baseline median at three sites on Cottonwood Creek (CT-P1, CT-1, and CT-2) during baseflows and during storm events at CT-2

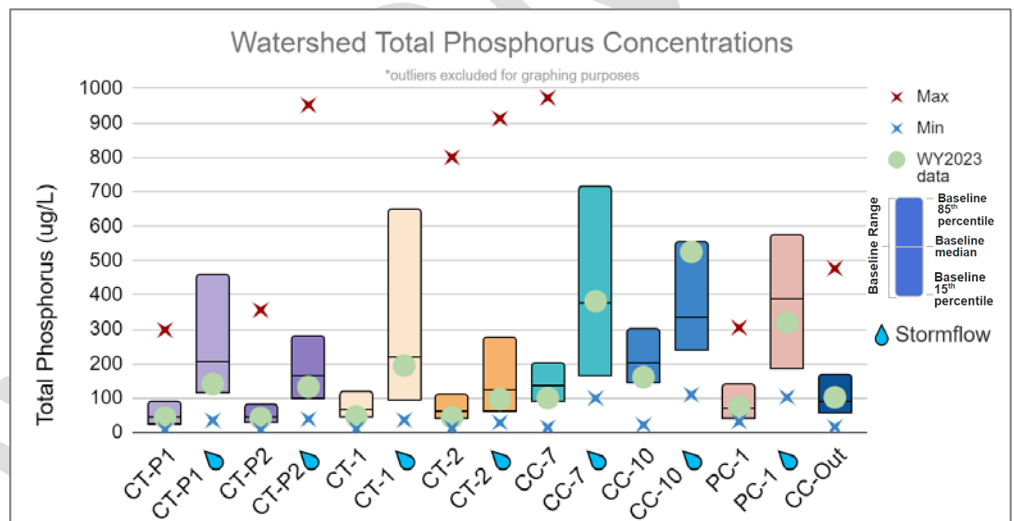


Figure ES-4. Cherry Creek Watershed Phosphorus Concentrations (CT = Cottonwood Creek; CC = Cherry Creek ad PC = Piney Creek)

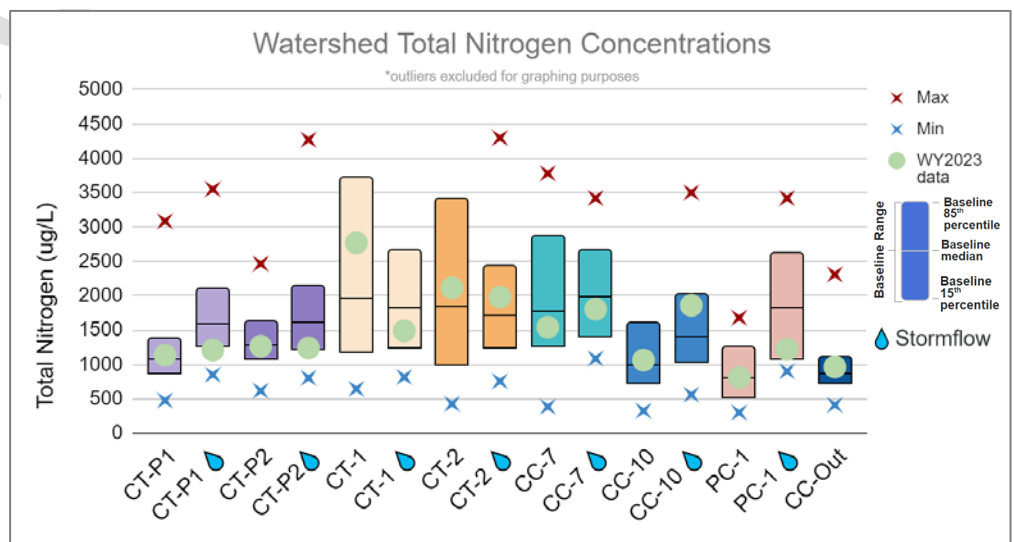


Figure ES-5. Cherry Creek Watershed Nitrogen Concentrations (CT = Cottonwood Creek; CC = Cherry Creek ad PC = Piney Creek)

(Figure ES-5). The WY 2023 median TN on Cherry Creek at CC-10 and the outlet to the Reservoir (CC-0) were also higher than the baseline median during baseflow conditions.

POLLUTION REDUCTION FACILITIES (PRF) HIGHLIGHTS

The Pollution Reduction Facilities (PRFs) in the watershed are monitored on an ongoing basis to determine the effectiveness of water quality benefits upstream to downstream annually and over time.

Based on the water quality concentrations in baseflow and stormflow events during WY 2023 and the last 10-years, the Cottonwood Creek PRF ponds and treatment train as a whole reduced phosphorus and suspended sediment concentrations in downstream stormflows. During WY 2023, the Cottonwood Treatment Train showed statistically significant removal of TP, TSS and VSS upstream to downstream during storm flows which is also true when evaluating the trend over the last 10 years. Both forms of suspended solids were also significantly lower in baseflow during WY 2023 through the whole treatment train. Peoria Pond also showed significant removal of TP and TSS upstream to downstream during stormflow conditions over the same period. The Perimeter Pond PRF demonstrated significant reductions in TP and TSS concentrations in base flow and stormflow during WY 2023 and the last 10 years. The McMurdo Gulch upstream to downstream concentration analysis also demonstrated a statistically significant reduction of all nutrients in WY 2023 TP and TN. This is similar to the 2014 and 2023 significance of downstream reductions with the exception of TN.

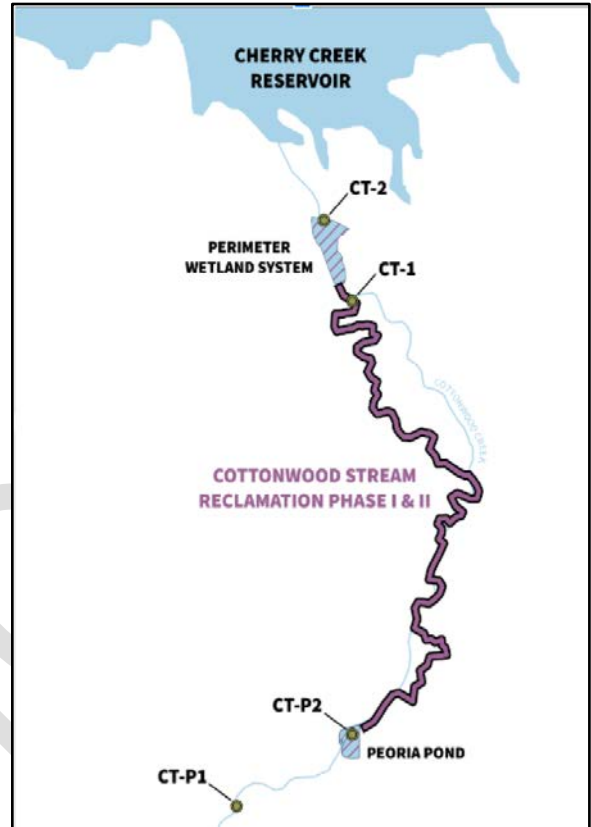


Figure ES-6. Cottonwood Creek Pollution Reduction Facilities (PRFs)

Table ES-1. Summary of Reductions in Nutrient and Suspended Solids in CCBWQA PRFs, WY 2023. *

PRF	Cottonwood Treatment Train		Peoria Pond		Perimeter Pond		Cottonwood Creek btw Ponds		McMurdo Gulch
	Base	Storm	Base	Storm	Base	Storm	Base	Storm	Base
Nitrate+ Nitrite					●				●
Ammonia									●
Nitrogen, Total					●				○
Phosphorus, Soluble Reactive									●
Phosphorus, Dissolved									●
Phosphorus, Total		●		●	●	●			●
Total Suspended Solids	○	●		●	●	●			
Volatile Suspended Solids	●	●				●			

*Legend: ○ reduction of net median downstream in WY 2023, □ significant reductions of net median downstream (2014-2023), ● significant net reduction in WY 2023 and 10-year median downstream, blank cells indicate no significant reduction or an increase upstream to downstream

GROUNDWATER HIGHLIGHTS

The groundwater and alluvium of Cherry Creek also play a role in nutrient dynamics as water moves down the watershed and flows into the Reservoir.

Total dissolved phosphorus (TDP) and soluble reactive phosphorus (SRP) are used for long-term evaluation of groundwater phosphorus since they are the representative forms found in groundwater and a longer period of record is available for these forms. TP concentrations in groundwater samples can be elevated by sediment disturbance during groundwater sampling. A Mann-Kendall

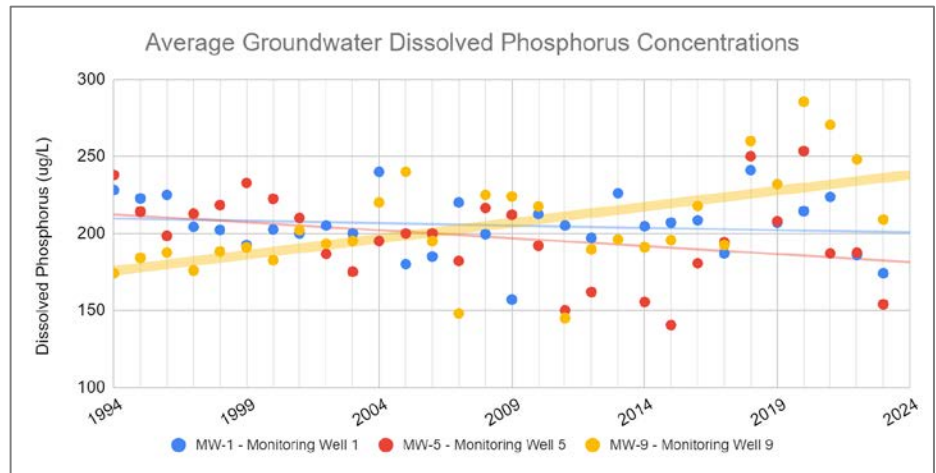


Figure ES-7. Groundwater Dissolved Phosphorus

statistical trend analysis demonstrated that TDP and SRP in the groundwater at MW-9 upstream of the Reservoir are significantly increasing over time, although there is not a similar statistically significant trend at the upstream monitoring well sites (MW-1 and MW-5) (Figure ES-7). Conductivity in the groundwater has a statistically significant increasing trend from upstream (MW-1) to downstream towards the Reservoir (MW-9).

WATER BALANCE HIGHLIGHTS

Due to circumstances beyond the control of the CCBWQA, including equipment damage due to significant flooding, some of the inflow data required for the calculations for the annual water balance is not available. As an alternative, the relative inflow discharge ratio of Cherry Creek to Cottonwood Creek from 2016-2022, along with the inflow, outflow and reservoir storage provided by the USACE will be used. However, as of the end of 2023, the storage information provided by the USACE is not available due to a discrepancy in the elevation datum shift. USACE plans to address this discrepancy by the end of January 2024, at which time the storage information will be provided, the required calculations can be completed, and an amended Monitoring Report will be issued.

NUTRIENT BALANCE HIGHLIGHTS

The nutrient concentrations of the inflows and the outflow of Cherry Creek Reservoir are used to calculate the mass storage on an annual basis. The flow-weighted influent phosphorus goal, derived as part of the 2009 Reg 38 rulemaking process to achieve the 18 $\mu\text{g/L}$ chl α standard, is 200 $\mu\text{g/L}$. Flow-weighted nutrient concentrations and mass storage in the Reservoir for WY 2023 will be provided after the water balance has been completed.

WY 2023 CONCLUSIONS

The CCBWQA's comprehensive monitoring program and WY 2023 data provide insight into current conditions and long-term trends in the watershed and Cherry Creek Reservoir. Although Cherry Creek Reservoir did not meet the chl α seasonal standard for WY 2023, it did meet the Reg 38 standards for temperature, pH, and DO to support the Class 1 Warm Water Aquatic Life classification. Cherry Creek Reservoir's trophic state continues to remain eutrophic to hypereutrophic with elevated phosphorus concentrations, reduced water transparency, and

algal growth. A cyanobacteria bloom in WY 2023 resulted in a brief closure to recreational users of the Reservoir due to the presence of toxins.

The WY 2023 weather conditions resulted in above-average stream inflows, higher water levels, and shorter residence time in Cherry Creek Reservoir which would normally be beneficial to water quality. However, the high phosphorus concentrations that entered the Reservoir during the flood events increased the potential for algal growth, cyanobacteria blooms, and high chl α concentrations during the summer months.

There are notable differences in water quality between Cherry Creek, Cottonwood Creek, and Piney Creek. Cherry Creek has much higher concentrations of phosphorus, whereas Cottonwood Creek has higher relative concentrations of nitrogen. Although median watershed TP concentrations in WY 2023 were lower than baseline medians at most sites, the WY 2023 TP concentrations on Cherry Creek just upstream of the Reservoir were well above baseline medians. Since Cherry Creek contributes approximately $\frac{3}{4}$ of the annual stream inflows, water quality in the Reservoir is usually most impacted by changes at this site (CC-10).

Conductivity in the streams and groundwater is significantly increasing over time, which impacts the Reservoir water quality and dynamic. Although the sources of the increased conductivity have not been identified, potential sources could be deicing chemicals and other discharges.

In WY 2023, the constructed wetland PRF ponds on Cottonwood Creek functioned effectively to remove phosphorus and suspended solids during stormflow conditions. In addition, the PRF Ponds on Cottonwood Creek have been functioning effectively when evaluating upstream to downstream concentrations on a long-term basis.

The above average spring precipitation in the Cherry Creek watershed caused flooding along Cherry Creek and Cottonwood Creek that damaged multiple monitoring stations and impacted data collection and stream flow calculations. Due to these factors and the associated data gaps, alternative calculations will be used in the water balance, nutrient balance and mass storage once data is available. These results will be provided in an amended report in 2024.

1.0 INTRODUCTION

The mission and vision of the Cherry Creek Basin Water Quality Authority (CCBWQA) are to benefit the public by improving, protecting, and preserving water quality in Cherry Creek and Cherry Creek Reservoir (Reservoir) for recreation, fisheries, and other warm water aquatic life, water supplies, and agriculture to achieve and maintain current water quality standards. The CCBWQA also supports effective efforts by partner counties, municipalities, special districts, and landowners within the basin providing for the protection of water quality, ensuring that new developments and construction activities pay their equitable share of costs for water quality preservation and facilities, and promoting public health, safety, and welfare.

The CCBWQA was formally created by statute in 1988 by the Colorado State Legislature. The CCBWQA Board consists of representatives from two counties and eight cities, along with one representative from each of the seven special districts that provide water and wastewater treatment in the basin, and seven public representatives appointed by the Governor.

The Cherry Creek Basin watershed includes over 386 square miles and 600 miles of creeks and streams (Figure 1). The U.S. Army Corps of Engineers (USACE) states that Cherry Creek Reservoir has a maximum surface area of 850 surface acres at an operating pool of 5550 ft elevation. The Reservoir is located near the base of the watershed, south of I-225 and west of Parker Rd., in Cherry Creek State Park (CCSP or the Park).. The Park covers approximately 4,000 acres and is one of the most productive fisheries and widely enjoyed recreational areas in Colorado. The Park has miles of trails to view birds and wildlife with scenic views of the Rocky Mountains in the background.

USACE constructed the Reservoir between 1948 and 1950 for flood control. In 1951, the State Park Board leased Cherry Creek recreation area from the USACE and created Colorado's first state park, which was opened in 1959. In addition to providing flood control, the Reservoir is a recreational and aquatic life amenity, and water released from the Reservoir supports downstream agriculture and water supply uses.

The Water Quality Control Commission (WQCC) adopted use classifications and water quality standards for the Reservoir and watershed, most recently effective August 30, 2023. These numeric standards, as specified in Regulation No. 38 (5 CCR 1002-38) (Reg 38), include the mainstem of Cherry Creek to the inlet of the Reservoir and from the outlet to the confluence with the South Platte River, Cherry Creek Reservoir, Cottonwood Creek, and other tributaries, lakes, and reservoirs within the watershed. These standards are set to protect recreation, aquatic life, agriculture, and water supply uses. The CCBWQA focuses on improving, protecting, and preserving the water quality of Cherry Creek and Cherry Creek Reservoir, and on achieving and maintaining the existing water quality standards.

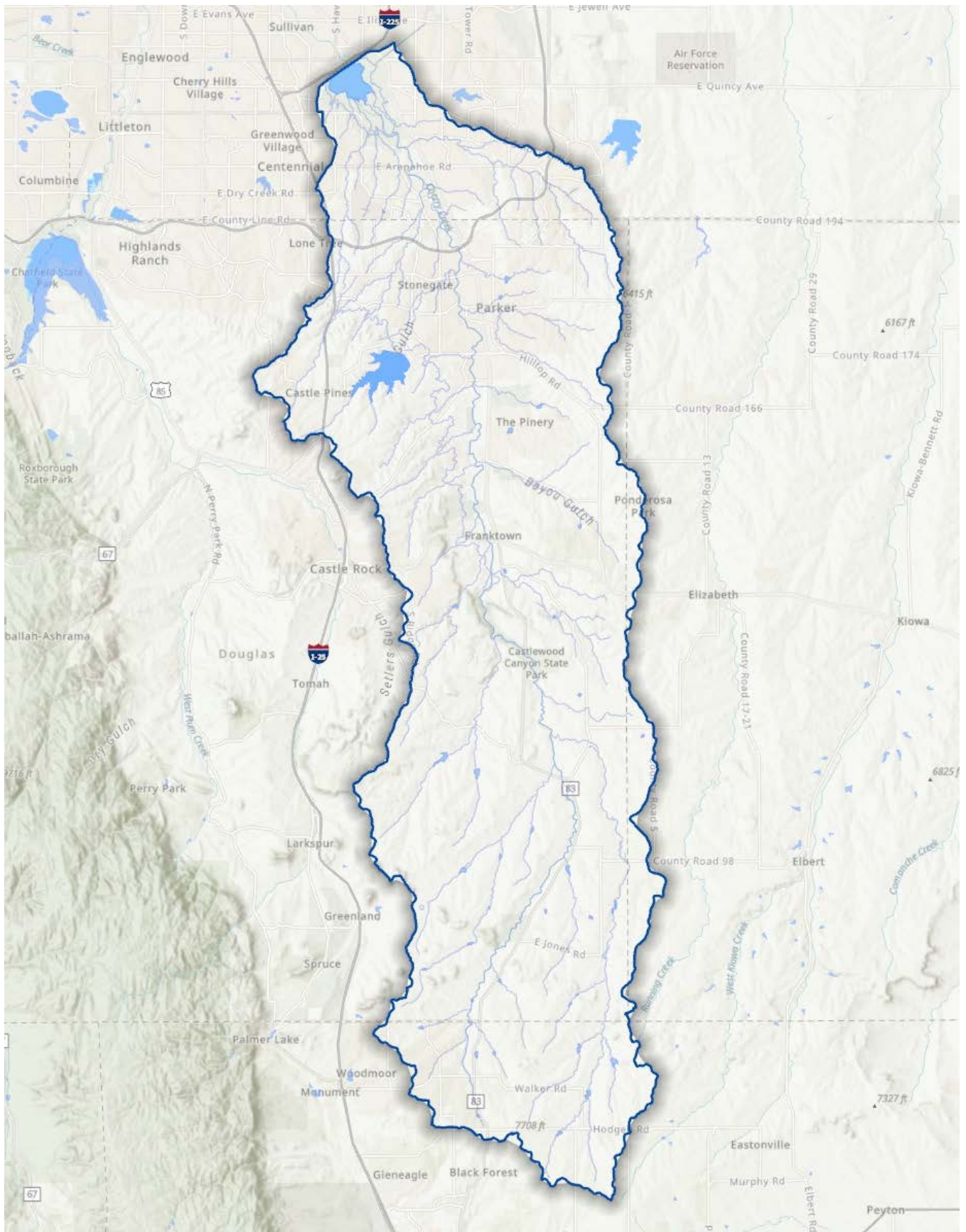


Figure 1. Cherry Creek Basin.

2.0 MONITORING PROGRAM

The WQCC's Cherry Creek Reservoir Control Regulation No. 72 (5 CCR 1002-72), (CR 72), requires that the CCBWQA execute a water quality monitoring program of the Cherry Creek watershed and Reservoir for water quality, inflow volumes, alluvial water quality, and non-point source flows. The program is implemented to determine total annual flow-weighted concentrations of nutrients to the Reservoir and to monitor the Pollutant Reduction Facilities (PRFs) to determine inflow and outflow nutrient concentrations. The sample collection and analysis provide data required to evaluate the nutrient sources and transport, characterize reductions in nutrient concentrations, and calculate and document compliance with associated water quality standards. In addition, these data are used to update the Reservoir and Watershed models.

The CCBWQA Sampling and Analysis Plan/Quality Assurance Project Plan (SAP/QAPP) provides the foundation for the sampling and analysis program, including sampling methods, QA/QC (quality assurance/quality control) and protocols. The monitoring program was designed to understand and quantify the relationships between nutrient loading and Reservoir productivity. The routine monitoring of surface water and groundwater was implemented to promote the concentration-based management strategy for phosphorus control in the basin, to determine the total annual flow-weighted concentration of nutrients to the Reservoir, to evaluate watershed nutrient sources and transport mechanisms, and to evaluate the effectiveness of PRFs including the cumulative effect of stormwater control measures (SCM, also known as BMPs) implemented in the basin.

All monitoring activities and analytical work are performed in accordance with the SAP/QAPP, which includes details of the current monitoring program (monitoring locations, frequency, parameters analyzed, etc.) and can be found on the CCBWQA website, <https://www.cherrycreekbasin.org/plans>. The monitoring sites and details regarding type, frequency, events, and telemetry is displayed in Figure 2.

This WY 2023 Monitoring Report summarizes data collected during the 2023 water year and includes an assessment and evaluation of data and results from the Reservoir and watershed sampling and analysis, including water quality and quantity of surface water, groundwater, stormwater, and the effectiveness of Pollutant Reduction Facilities (PRFs). The water quality data and results described herein available on the CCBWQA's Data Portal, <http://www.ccbwqportal.org>.

2.1 MONITORING METHODS AND ANALYTE DESCRIPTIONS

The parameters analyzed in the monitoring program are useful in determining the suitability of the water for aquatic life, recreational use, and attaining water quality standards, collectively referred to as “beneficial uses.” These parameters are also used to define lake trophic state and interactions between the chemical and biological components of lake ecosystems.

All analyses were conducted using approved methods described by the U.S. EPA and/or Standard Methods and are detailed in the SAP/QAPP. A summary of the key parameters and metrics described in this report are described below.

pH

The hydrogen ion activity, indicating the balance of acids and bases in water, determines pH. A pH of 7 is considered neutral, a pH less than 7 is considered acidic, while a pH greater than 7 is considered basic. Reg 38 has a standard range for pH between 6.5 and 9.0 for aquatic life. Since pH is expressed on a logarithmic scale, each 1-unit change in pH represents a ten-fold increase or decrease in hydrogen ion concentration. Therefore, a pH of 6 would be 10 times more acidic than a pH of 7 and 100 times more acidic than a pH of 8. The pH of normal rainwater (containing no pollutants) is about 5.6. As the rainwater travels over and through rocks and soil, chemical reactions with minerals affect the pH and increase the buffering capacity of the water.

Oxidation Reduction Potential

Oxidation reduction potential (ORP) measurements are used to quantify the exchange of electrons during chemical reactions in which the oxidation states of atoms are changed, also known as redox or oxidation-reduction reactions. Electrical activity is reported in millivolts (mV). At the water/sediment boundary layer, microbial organisms facilitate the chemical reactions but do not actually oxidize or reduce the compounds. Redox reactions provide energy for microbial cells to carry out their metabolic processes (Wetzel 2001). The combination of microbial organisms and redox reactions are responsible for the breakdown of organic matter and development of anoxic conditions near the sediment boundary in reservoirs during the summer. Higher ORP values indicate an oxidizing environment and high potential to break down organic matter in the water. Low and negative values indicate a reducing environment and usually correlate to lower dissolved oxygen concentrations and higher microbial decomposition activity normally present at deeper sites and in the sediments of lakes.

Conductivity

Conductivity (Specific Conductance) is the ability of water to conduct an electrical current and is based on the dissolved inorganic solids (positive and negative ions) present. Conductivity is a useful general measure of water quality since values increase with salinity and can be an indicator of dissolved solids that can be considered “pollutants” in the water. The geology of the area, water source, and watershed affect conductivity. Conductivity values of 50-1500 $\mu\text{S}/\text{cm}$ are typical for surface water. Conductivity also varies in direct proportion with temperature with higher temperature increasing the conductivity. Thus, to allow direct comparison of samples collected at different temperatures, conductivity is typically corrected to 25°C and reported as specific

conductance ($\mu\text{mhos/cm}$ @ 25 °C). For the sake of simplicity, specific conductance is referred to as “conductivity” in this report.

Dissolved Oxygen

Dissolved oxygen (DO) is the amount of oxygen gas dissolved in the water column. Small amounts of oxygen enter the water column by direct diffusion at the air/water interface and oxygen is also produced during photosynthesis. DO gradients provide an indication of mixing patterns and the effectiveness of mixing processes in a lake. DO concentrations also have an important bearing on the physical-chemical properties of lakes and the composition of a lake's biota. Lakes impacted by heavy sediment loads may experience low DO levels since the increased turbidity caused by suspended particles can reduce light penetration and limit photosynthesis. The breakdown of organic matter or decomposition can consume large amounts of oxygen from the water column and reduce DO. Fish require oxygen for respiration and may become stressed at levels less than 5.0 mg/L. DO can be expressed as concentration (mg/L) or as percent saturation. DO saturation is directly related to temperature and the capacity of water to absorb oxygen decreases as temperature increases.

Temperature

Water temperature affects the DO concentration of the water, the rate of photosynthesis, rates of chemical reactions, metabolic rates of aquatic organisms, and the sensitivity of organisms to toxins, parasites, and disease. All aquatic organisms are dependent on certain temperature ranges for optimal health. If temperatures are outside of this optimal range for a prolonged period of time, the organisms become stressed and can die. Water temperature generally increases with turbidity; as the particles absorb heat, the DO levels are reduced. Temperature is primarily controlled by climatic conditions but can also be impacted by human activities.

Secchi Depth

The Secchi depth of a waterbody is a way to quantify turbidity or water clarity. It is measured with an 8” black and white disk. The disk is slowly lowered into the water column and the depth at which it is no longer visible becomes the Secchi depth. The measurement is based on both light absorption and the amount of light scattered by particles in the water column. The Secchi depth is higher when there is greater clarity or fewer particles in the water and is usually a representation of productivity of the water. Secchi depths of less than 6.6 feet (2.0 meters) have traditionally been considered undesirable for recreational uses in natural lakes; however, lower clarity is usually tolerated in reservoirs.

Light Transmission

Light transmission is a measurement of light absorption in the water column. The depth at which 1% of the surface light penetrates is considered the lower limit of algal growth and is referred to as the photic zone (see below). The measurement of 1% light transmission is accomplished by using both an ambient and an underwater quantum sensor attached to a data logger. The ambient quantum sensor remains on the surface, while the underwater sensor is lowered into the water on the shady side of the boat. The underwater sensor is lowered until the value displayed on the data logger is 1% of the value of the ambient sensor, and the depth is recorded.

Photic Zone

The Photic Zone of an aquatic resource is calculated as the depth at which light can penetrate or the depth of the water column where phytoplankton could complete photosynthesis based on light availability. Samples in Cherry Creek Reservoir are collected as a composite from what represents the common photic zone based on conditions, typically from 0-3m. See Light Transmission above.

Chlorophyll *a*

Chlorophyll is the green pigment that allows plants to photosynthesize. The measurement of chl α in water provides an indirect indication of the quantity of photosynthesizing phytoplankton found in the water column. It is found in all algal groups, as well as in cyanobacteria. More specifically, chl α is a measurement of the portion of the pigment that was still actively photosynthesizing at the time of sampling and does not include dead biomass. In surface water, lower chl α concentrations correspond to oligotrophic or mesotrophic conditions (0-6 $\mu\text{g/L}$), where higher concentrations indicate a eutrophic (6-40 $\mu\text{g/L}$) or hypereutrophic state (>40 $\mu\text{g/L}$).

Phosphorus

Phosphorus can be found in several forms in freshwater, but the biologically available form of soluble inorganic orthophosphate, operationally referred to as soluble reactive phosphorus results in nuisance plant and/or algal growth. Inorganic phosphates quickly bind to soil particles and plant roots and, consequently, much of the phosphorus in aquatic systems is bound and moves through the system as sediment particles. Organic phosphates are phosphorus forms found in the cells of plants and other organisms and are biologically unavailable. Under anoxic (low oxygen) conditions, bound phosphorus can be released from bottom sediments, and the concentration of biologically available orthophosphate can increase dramatically. The erosion of soil particles from steep slopes, disturbed ground, and stream channels is often an important source of phosphorus in aquatic systems. Surface runoff containing phosphorus from fertilizers, wastewater effluent, and decaying organic matter also contribute to biologically available phosphorus enrichment.

Total Phosphorus (TP) is the measure of all phosphorus in a sample as measured by persulfate digestion and includes inorganic, oxidizable organic, and polyphosphates. This includes what is readily available, has the potential to become available, and stable forms. In lakes and reservoirs, concentrations <12 $\mu\text{g/L}$ are considered oligotrophic; 12-24 $\mu\text{g/L}$ mesotrophic; 25-96 $\mu\text{g/L}$ eutrophic; and >96 $\mu\text{g/L}$ hypereutrophic.

Soluble Reactive Phosphorus (SRP) is the measure of dissolved inorganic phosphorus (PO_4^{3-} , HPO_4^{2-} , H_2PO_4^- , and H_3PO_4). This form is readily available in the water column for phytoplankton growth.

Total Dissolved Phosphorus (TDP) is a measure of all phosphorus forms (inorganic, organic, and polyphosphate) that are dissolved in water.

Nitrogen

Nitrogen has a complex cycle and can exist in organic, inorganic, particulate, gaseous, and soluble forms. The soluble, inorganic oxidized forms are nitrate (NO_3^{-1}), and nitrite (NO_2^{-1}), which are normally found in surface water. The reduced inorganic form is ammonia (NH_3), which is normally found in low-oxygen environments. The inorganic forms, NO_3^{-1} , NO_2^{-1} , and NH_3 are the most available for primary productivity or algal growth. However, atmospheric nitrogen (N_2) can also be used as a nutrient source by some species of algae or cyanobacteria, and various other reduced forms of nitrogen can be produced by decomposition processes. Particulate and dissolved organic forms of nitrogen are not immediately available to drive algal growth but can be converted to ammonia by bacteria and fungi and can be oxidized to form nitrites and then nitrates. Surface runoff can contain inorganic nitrogen from fertilizers and organic nitrogen from animal waste, wastewater, etc.

Total Nitrogen (TN) is the quantity of all nitrogen in the water and is calculated by adding the measured forms of organic nitrogen, oxidized nitrogen, and ammonia.

Nitrates and Nitrites ($\text{NO}_3^- + \text{NO}_2^-$) are the sum of total oxidized nitrogen, often readily available for algal uptake.

Ammonia (NH₃-N) is a reduced form of dissolved nitrogen that is readily available for phytoplankton uptake. NH₃ is found where dissolved oxygen is lacking, such as in a eutrophic hypolimnion, and is produced as a by-product by bacteria during decomposition.

Nitrogen/Phosphorus Levels and Ratios

Phytoplankton require both macronutrients, such as phosphorus, nitrogen, and carbon, and trace nutrients, including iron, manganese, and other minerals, for growth. Biological growth is limited by the substance that is present in the minimum quantity with respect to the needs of the organism. The ratio of total nitrogen (TN) to total phosphorus (TP) in a water body provides insight into nutrient limitation in the water body. Since many species of harmful cyanobacteria (blue-green algae) have the ability to fix nitrogen from the atmosphere, they have a competitive advantage over other algae in phosphorus-rich environments when nitrogen is limited and can become dominant over the more beneficial green algae species. Maintaining a molar TN:TP ratio greater than 16:1, or 7:1 ratio by weight, will favor a balanced phytoplankton diversity and reduce the potential for a cyanobacteria-dominated environment. The ratio of total inorganic nitrogen (nitrate, nitrite, and ammonia) to soluble reactive phosphorus (TIN:SRP) can sometimes be more indicative of phytoplankton growth potential since these are the nutrient forms most available in the water column.

Trophic State

The Trophic state as described by Vollenweider (1970) is used as a guideline for describing water quality as it relates to the trophic state or biological productivity potential. Many indices assign numerical values to trophic state based on multiple water quality parameters. The following are typical characteristics of various trophic states:

Oligotrophic - lack of plant nutrients, low productivity, sufficient oxygen at all depths, clear water, deeper lakes can support trout,

Mesotrophic - moderate plant productivity, hypolimnion may lack oxygen in summer, moderately clear water, warm water fisheries only,

Eutrophic - contains excess nutrients, blue-green algae dominate during summer, algae scums are probable at times, hypolimnion lacks oxygen in summer, poor transparency, rooted macrophyte problems may be evident,

Hypereutrophic - algal scums dominate in summer, few macrophytes, no oxygen in hypolimnion, fish kills possible in summer and under winter ice.

Alkalinity

Alkalinity, expressed as mg CaCO₃/L, represents the presence of bicarbonates and carbonates in water and indicates the buffering capacity. A higher buffering capacity can reduce the potential for pH swings during photosynthesis (removing CO₂) primary producers (algae) and plant growth. A minimum alkalinity of 20mg/L is the aquatic life criteria recommended by the EPA.

Anions: Chloride and Sulfate

Chloride and sulfate are the major anions (negative ions) that play a role in conductivity and can be indicators of pollutants entering a watershed due to de-icing activities, treated wastewater discharge, stormwater runoff, naturally elevated conditions in groundwater, etc. Conductivity is a measure of the ability of water to conduct electricity, which is a function of all the dissolved ions in solution. Since chloride and sulfate are ions in solution, any increase in their concentrations increases conductivity.

Cations: Calcium, Magnesium, Sodium, and Potassium

The major cations (positive ions) that contribute to dissolved solids concentration in water typically are calcium, magnesium, sodium, and potassium. These ions can also indicate pollutants entering a watershed such as de-icing products, treated wastewater discharge, stormwater runoff, etc. Starting in 2022, these parameters were included in the data analysis for one reservoir site and three surface water sites twice a year so the major contributions to conductivity can be evaluated when enough data has been collected.

Suspended Solids

Total Suspended Solids (TSS) is a quantification of concentrations of suspended sediment and other particulates in water. Suspended solids in lakes include both organic material, such as algal cells and other microorganisms, and inorganic particulate matter, such as silt and clay particles. Algae and other organisms appear to be the main source of TSS in the open waters, while suspended silts and clays appear to be the primary suspended solids in stream or groundwater samples. Volatile Suspended Solids (VSS) is a measure of the amount of particulate organic material that is present in water. Suspended solids in the water can indirectly impact chl α concentrations by reducing the opportunity for algae to photosynthesize.

Organic Carbon

Organic carbon provides a measure of all organic compounds in a water body and can provide an assessment of the carbon-based components or pollution of water. Plant material is often a major component of organic carbon and refractory organic compounds from plants can impart a dark color to lake water. Both total and dissolved organic carbon are measured in analytical samples.

2.2 WATER QUALITY ANALYSIS

The water quality data collected during the CCBWQA monitoring program is analyzed to evaluate short- and long-term changes or trends, seasonal and spatial variability, as well as compliance with applicable water quality standards. The Cherry Creek Watershed experiences seasonal fluctuations that influence water quality and trends over time.

To conduct the analysis, summary statistics are calculated for each parameter and monitoring location based on the entire or specified period of record (POR) which represents the baseline. The summary statistics and associated graphs in this report illustrate the median, 15th, and 85th percentiles of the POR data, as well as WY 2023 values. The central value of the dataset, known as the median, signifies the point where half of the sample set measurements are below, and half are above that value. The 85th percentile indicates that 85% of the measured values fall below this statistic. Conversely, the 15th percentile represents the statistic that 15% of the measured values fall below. The use of 85th/15th percentile serves as upper/lower indicators while mitigating the influence of potential minimum and maximum measurement outliers.

In addition to characterizing times series data using statistical summary values, it also is important to determine if there are significant trends in long-term data sets. Since water quality data are typically non-parametric (do not conform to a normal distribution) a Mann Kendall (MK) trend Analysis can quantify if time series data for a given location and parameter have a statistically significant trend. A p-value obtained from the MK trend test of less than 0.05 provides evidence of a significant monotonic trend in the time series. Conversely, if the p-value is greater than 0.05, it suggests that there is not enough evidence to conclude the presence of a significant monotonic trend.

3.0 WATERSHED MONITORING RESULTS

The watershed monitoring program includes an analysis of the quantity and quality of potential nutrient source inputs to Cherry Creek Reservoir. During WY 2023, surface water and groundwater sites in the watershed were monitored either monthly, every other month, on a bi-annual frequency, and/or during storm events to characterize spatial and temporal variability and differences in base and stormflow conditions.

The spring of 2023 received much higher-than-average precipitation, which caused major flooding and damage along Cherry Creek and Cottonwood Creek. Multiple monitoring stations, equipment, and other infrastructure were damaged and required repairs. These equipment issues and the extended elevated water levels impacted stream flow calculations. Please note that some data and measurements normally collected under the monitoring program are not available due to these factors outside of the CCBWQA’s control and alternative calculations will be used and provided with an amended report in 2024.

3.1 PRECIPITATION

Precipitation in the watershed and on the surface of the Reservoir plays a major role in water quality in the streams and overall Reservoir dynamics. Historically, precipitation in the Cherry Creek watershed has been measured at NOAA’s Centennial Airport weather station (KAPA) located at Latitude (Lat) 39.56°N, Longitude (Long) -104.85°W, and an elevation of 5,869 ft.

The meteorological station at Cherry Creek State Park (CCSP, located at Latitude (Lat) 39.63°N, Longitude (Long) -104.83°W, and an elevation of 5,631 ft was installed in 2021 (Figure 2). In WY 2023, the CCSP station measured a total of 22.3 inches of precipitation and the KAPA site measured 25.6 inches.

The watershed received approximately 172% of the historical average precipitation in WY 2023, with 15-16” in May and June accounting for over 60% of the entire year.

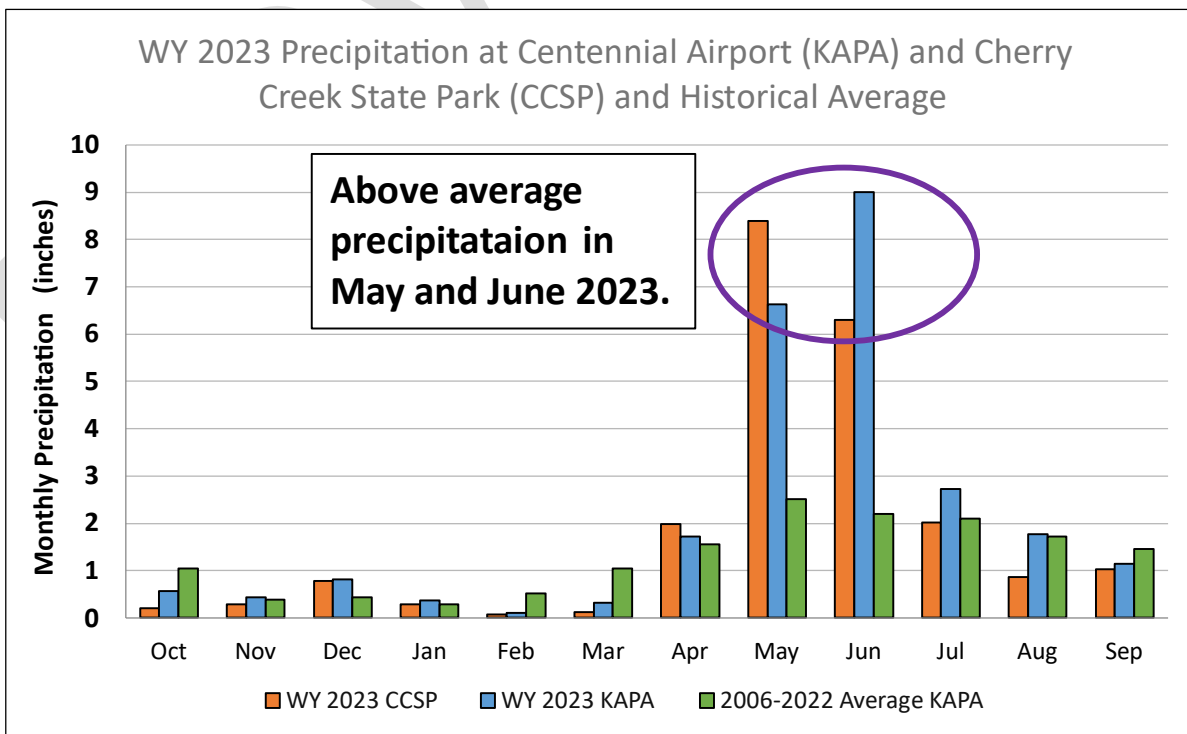



Figure 3. Monthly Watershed Precipitation in WY 2023 compared to (2006-2022) average.

Due to the closer proximity, the CCSP station should better represent the precipitation on the surface of the Reservoir and is used in water balance calculations. However, the KAPA site will continue to be used as a comparison and as a historical reference until a representative period of record can be developed for the CCSP site.

October 2022, and February, March and September 2023 received below average precipitation at both locations. However, May and June 2023 received much higher-than-average precipitation, accounting for over 60% of the precipitation for the entire year. The KAPA station measured a total of 25.6 inches of precipitation in WY 2023, approximately 172% of the historical average from 2006 to 2022 for this weather station (Figure 3).

STORM FREQUENCY ANALYSIS

The 24-hour total precipitation recorded was 3.35" on May 11th 2023, and 2.33" on June 22nd 2023, which were the top 2 ranked daily averages since 2006 at the KAPA site. When evaluating the frequency of dates that have received greater than 1" of precipitation, the probability of these two events was ~5 and 7% respectively. Probability and recurrence intervals are likely to be different if evaluated based on a more specific time-period or at different locations in the watershed. This is especially true for the June 22nd storm which was greater than 2" in 2 hours at the KAPA site near where the most notable stream flooding was observed on Cottonwood Creek but was much less at the CCSP site (Figure 3).

Evaluation of these storms using the Point Precipitation Frequency (PF) Estimates (NOAA, 2017) at the Cherry Creek Dam, Site ID 05-1547 indicates that the May 11th/ 12th storm, which recorded 6.48" inches in a 48-hour period at the CCSP Met station, has a recurrence interval of 500 years (500-year flood) (Figure 4) and the >2" of precipitation observed in two hours at the KAPA site is estimated to have a 25-year recurrence interval (Figure 5).

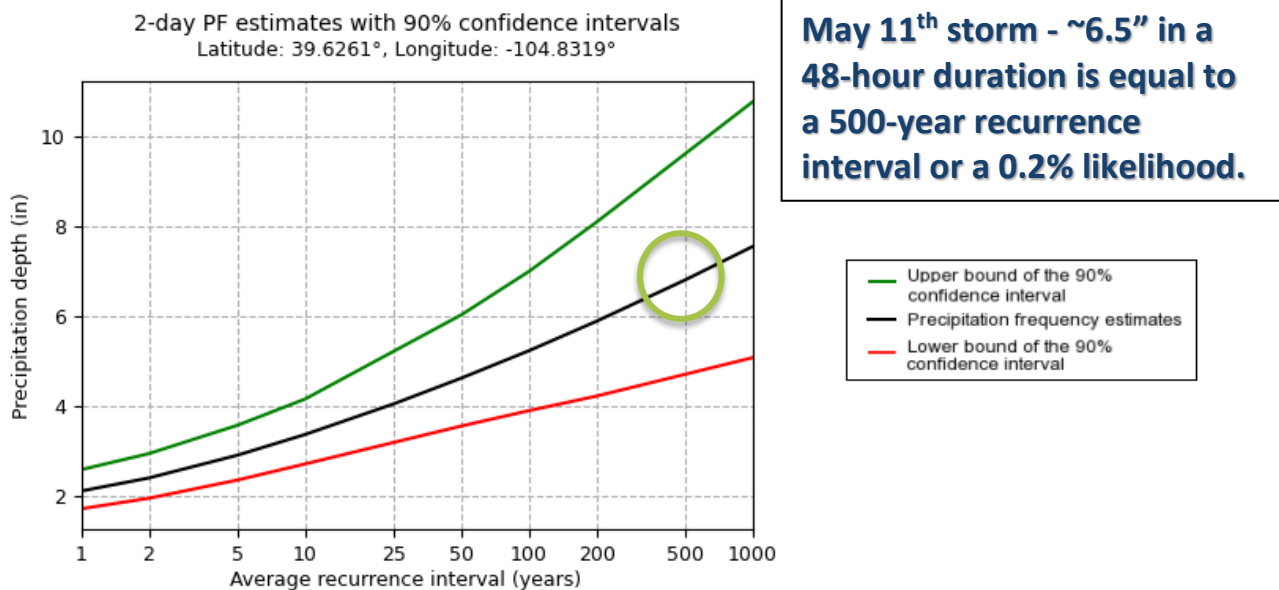
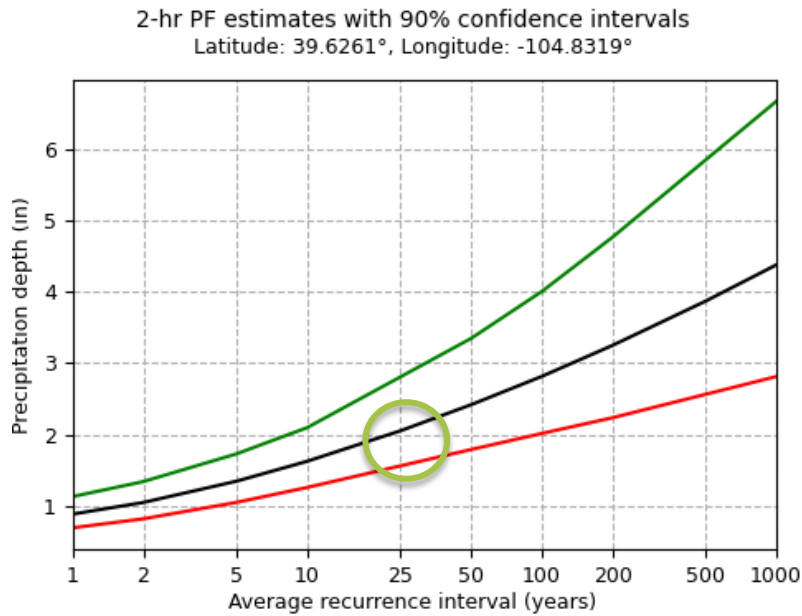


Figure 4. Average Recurrence Interval based on 48-hour Precipitation Frequency at Cherry Creek Dam site (NOAA)



June 22nd storm - >2" in a 2-hour duration is equal to a 25-year recurrence interval or a 4% likelihood.

— Upper bound of the 90% confidence interval
— Precipitation frequency estimates
— Lower bound of the 90% confidence interval

Figure 5. Average Recurrence Interval based on 2-hour duration at Cherry Creek Dam site, NOAA)

Additionally, when looking at NOAA's annual precipitation information, nearly all areas of the watershed received precipitation ranging between approximately 122 to 225 percent of normal when compared to the 30-year Parameter-elevation Regression on Independent Slopes Model (PRISM) normal from 1991-2020 (Figure 6). The watershed received approximately 200% of the 30-year average, while areas just above Cherry Creek Reservoir generally received less than average precipitation. This data is based on observed National Weather Service (NWS) precipitation from the CONUS River Forecast Centers and is displayed as a gridded resolution of roughly 4x4 km using bilinear interpolation in GIS.

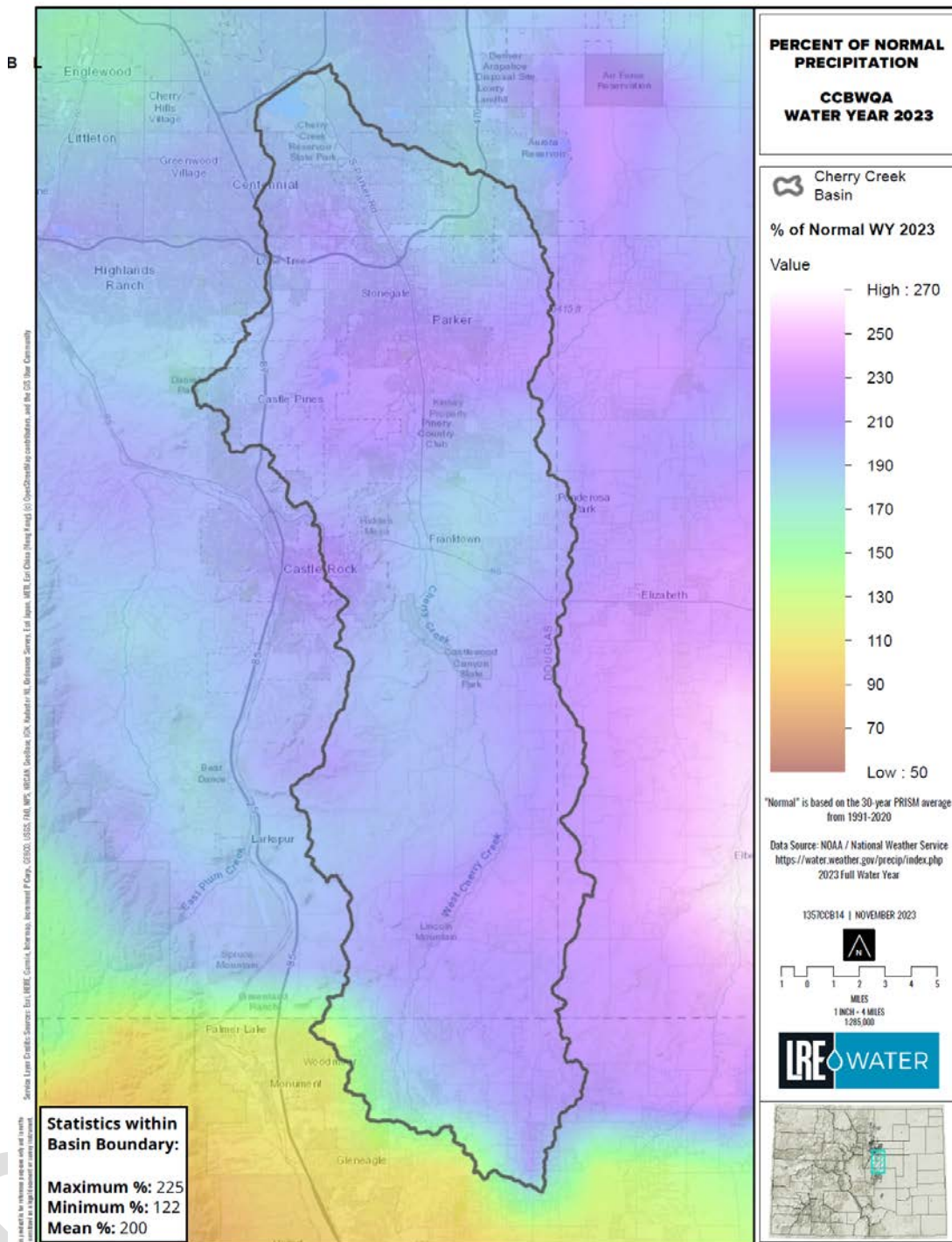


Figure 6. Percent of Normal Precipitation (30-year PRISM Average)

3.2 STREAM FLOWS

The U.S. Geological Survey (USGS) operates two gauging stations on Cherry Creek upstream of the Reservoir which are used as surface water monitoring locations for the SAP. The “Cherry Creek Near Franktown, CO” station (0671200) has an 80-year period of record (POR) and the “Cherry Creek near Parker, CO” station (393109104464500) has a 29-year POR.

3.2.1 CHERRY CREEK NEAR USGS FRANKTOWN SITE

The USGS Cherry Creek Near Franktown station is in Castlewood Canyon State Park at Lat 39°21'21", Long 104°45'46", Douglas County, CO, Hydrologic Unit 10190003 (Figure 2). The station, which represents the upper portion of the watershed is 1.3 mi downstream from Castlewood Dam site, and 2.5 mi south of Franktown. The USGS WY 2023 summary statistics are listed in the table to the right; Figure 7 shows the estimated daily discharge along with the historical daily mean from the last 82 years.

<p>USGS Gage - Cherry Creek near Franktown 2023 Statistics Drainage Area: 169 sq mi. Total Annual Flow: 4519 cfs/ 8960 AF/ Year Annual Mean Flow Rate: 12.4 cfs 24.5 AF/day Percent of Long-term Average (1940-2023): 139% Percent of 31-year average (1992-2023): 163%</p>

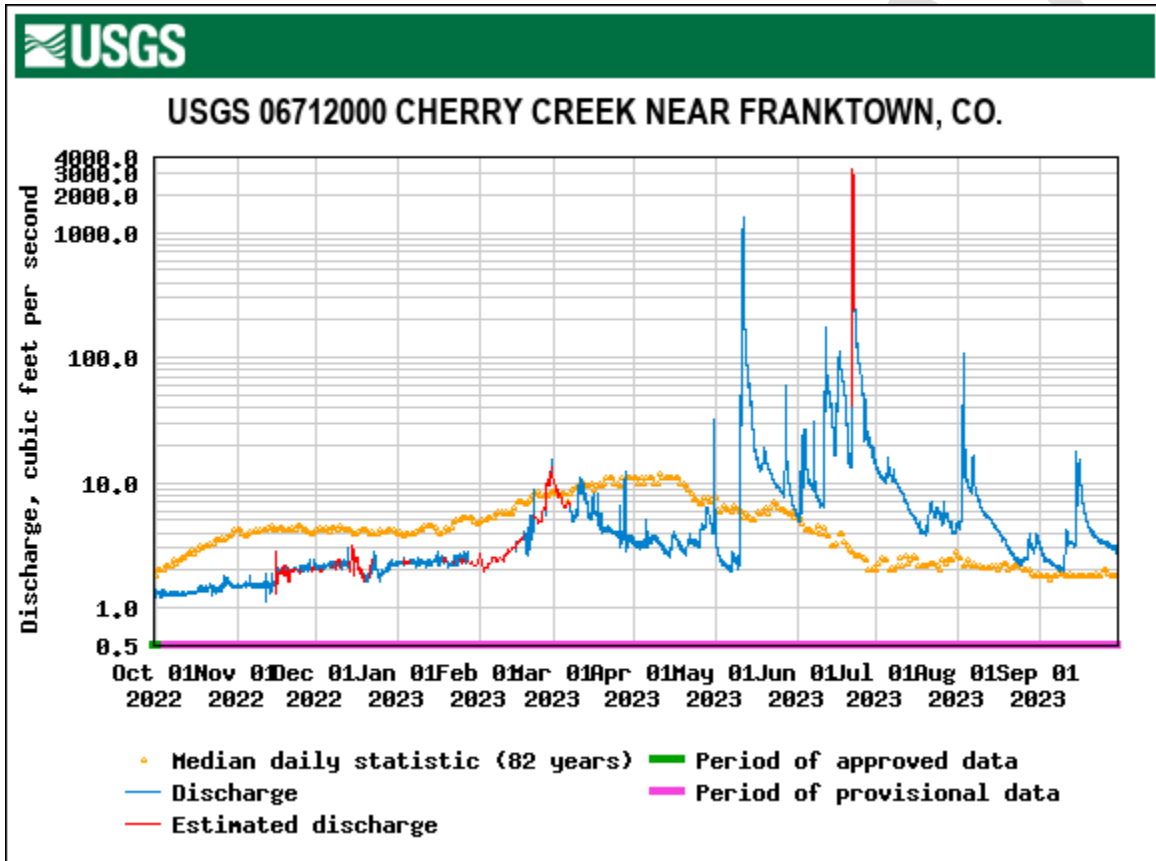


Figure 7. WY 2023 Daily Mean Discharge and Historical Median Flows for USGS Gauge near Franktown.

3.2.2 CHERRY CREEK NEAR USGS PARKER SITE

The USGS Cherry Creek near Parker station is located at Lat 39°31'09", Long 104°46'45", Douglas County, CO, Hydrologic Unit 10190003, 200 ft upstream from Main Street, 1,100 ft downstream from mouth of Sulphur Gulch, and 0.8 mi west of Parker Rd. This site is representative of the conditions in the middle of the watershed. The USGS WY 2023 summary statistics are listed in the table to the right; Figure 8 shows the estimated daily discharge along with the historical daily mean from the last 31 years.

<p>USGS Gage - Cherry Creek near Parker 2023 Statistics Drainage Area: 287 sq mi Total Annual Flow: 8042 cfs/ 15947 AF/year Annual Mean Flow Rate: 22.0 cfs/ 44 AF/day Percent of 31-year average (1992-2023): 190.5%</p>
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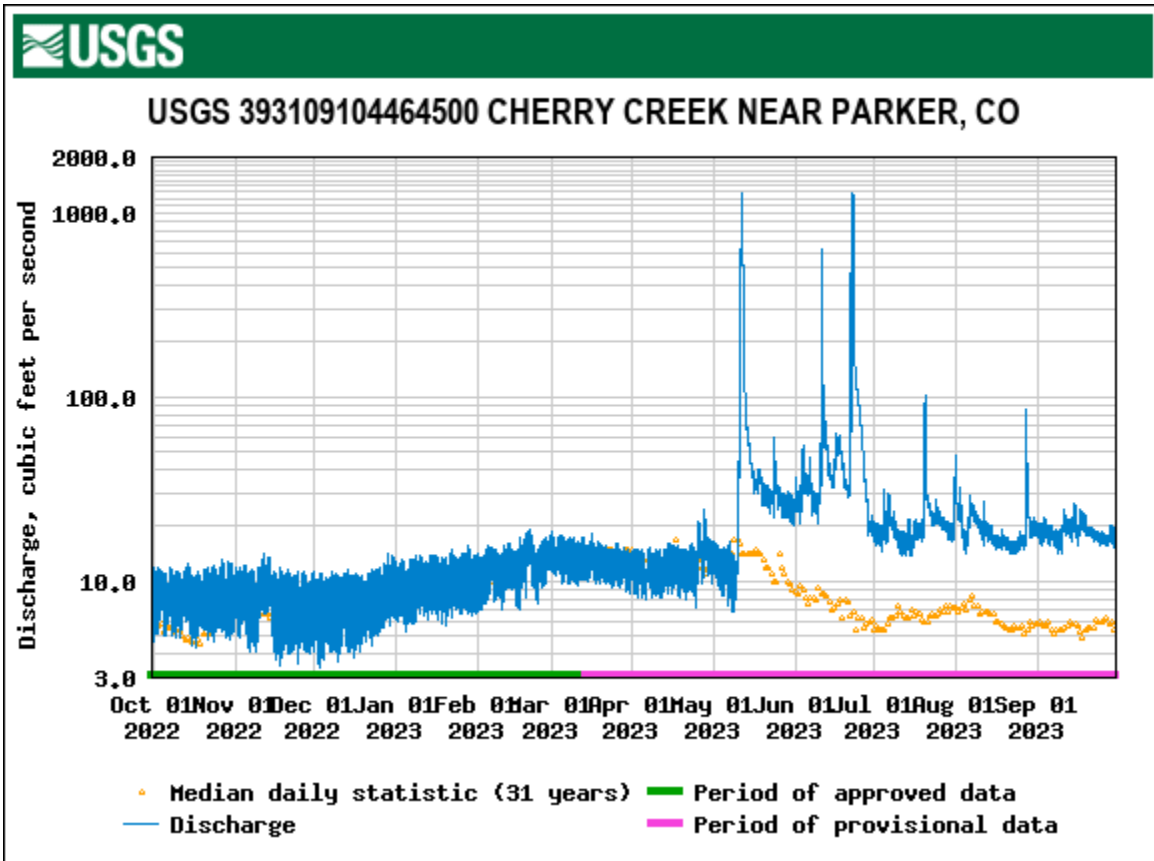


Figure 8. WY 2023 Daily Mean Discharge and Historical Median Flows for USGS Gauge near Parker.

3.2.3 CHERRY CREEK BELOW CHERRY CREEK LAKE

Water is released from the Reservoir through the dam’s outlet works. The USGS measures outflow at Station 06713000, Cherry Creek below Cherry Creek Lake, CO. The gauge is located approximately 2,300 ft downstream of the Reservoir. Other than releases from the Reservoir, there are no major surface water contributions to flow measured at this gauge. The USGS WY 2023 summary statistics are listed in the table to the right; Figure 9 shows the estimated daily discharge along with the historical daily mean from the last 31 years.

<p>USGS Gage - Cherry Creek below Cherry Creek Lake</p> <p>2023 Statistics:</p> <p>Total Annual Flow: 20,015.3 cfs/ 39,690 AF</p> <p>Annual Mean Flow Rate: 54.8 cfs/ 108 AF/day</p> <p>Percent of 31-year average (1992-2023): 277%</p>
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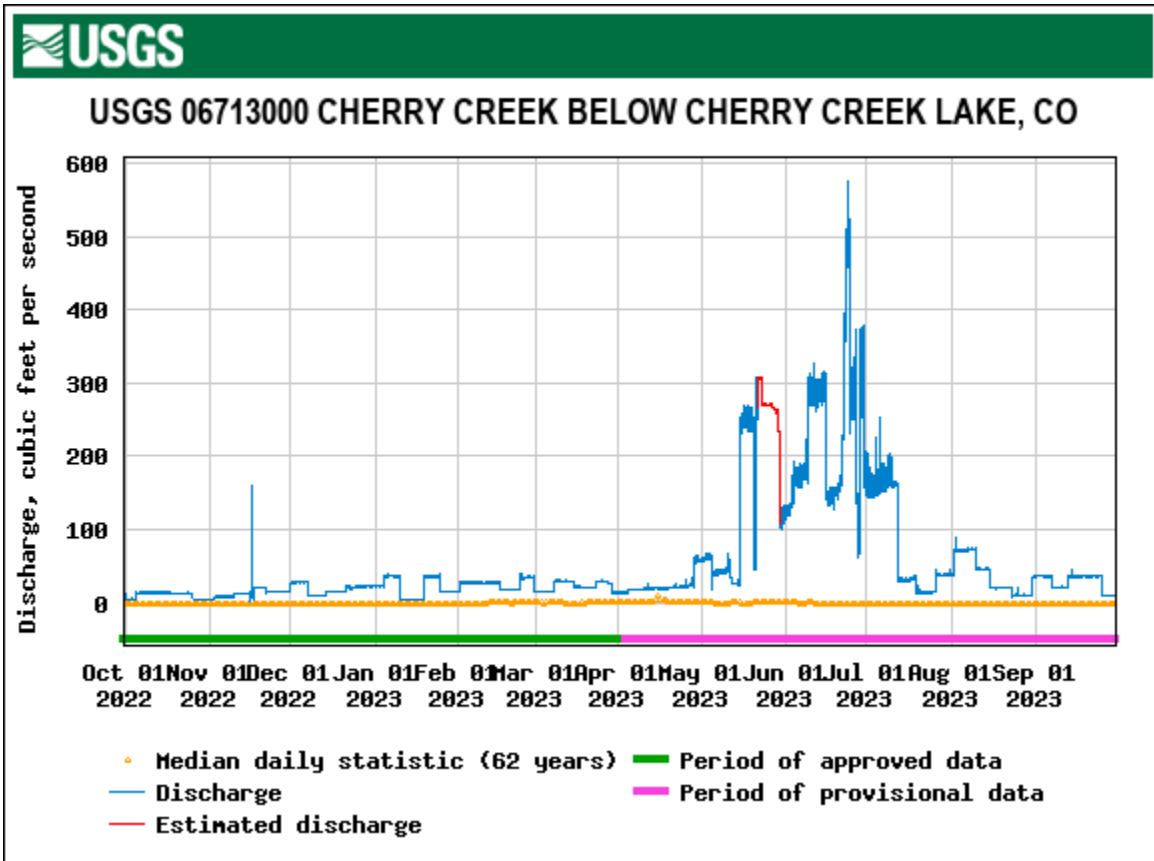


Figure 9. WY 2023 Daily Mean Discharge and Historical Median Flows for USGS Gauge Below Cherry Creek Lake.

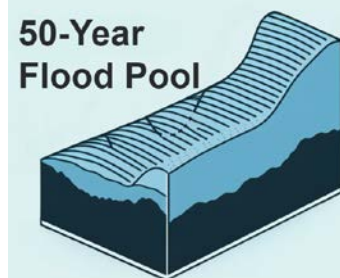
3.3 RESERVOIR INFLOWS

Chery Creek, the main inflow to Cherry Creek Reservoir, flows from south to north to the Reservoir through a 234,000-acre drainage basin. The basin includes various types of land use, including agriculture in the upper basin and higher-density development closer to the Reservoir, as well as permitted discharges in and around Cherry Creek. Cottonwood Creek has the second largest surface water input to Cherry Creek Reservoir with a sub-basin of 9,050 acres, which includes developed land use, and multiple wastewater dischargers.

The multiple large storm events in the Cherry Creek Watershed during 2023 affected stage measurements and associated flow calculations due to damaged equipment and inaccurate readings at the two stations on Cherry Creek and Cottonwood Creek upstream of the Reservoir that are used to calculate inflows (CC-10 and CT-2, respectively). (See Section 3.1)

Multiple calculations of the recurrence frequency of the precipitation events and stream flows in the Cherry Creek watershed are presented in Section 3.1. In addition, based on a Flood Hazard Delineation completed by the USACE (Figure 10), the mid-

The elevation of Cherry Creek Reservoir reached 5556.7 ft, near the **50-year Flood Pool** during the storm event in May 2023.



May storm reached near the 50-year flood pool when Reservoir elevation reached 5556.74 ft and near the 10-year flood pool in late June reaching 5555.90 ft (Figure 11). The Reservoir rose ~10 feet during the May storm-- the largest single event increase since the flood of 1965. The storm made local and national news due to the damage on Lakeview Drive, the main roadway through the State Park.

Based on the annual peak streamflow data at USGS station on Cherry Creek near Parker, CO from 1992-2023, the peak streamflow of 1,290 cfs on May 12th was the second highest daily flow since 2012 with a probability of only ~6% likelihood of occurrence.

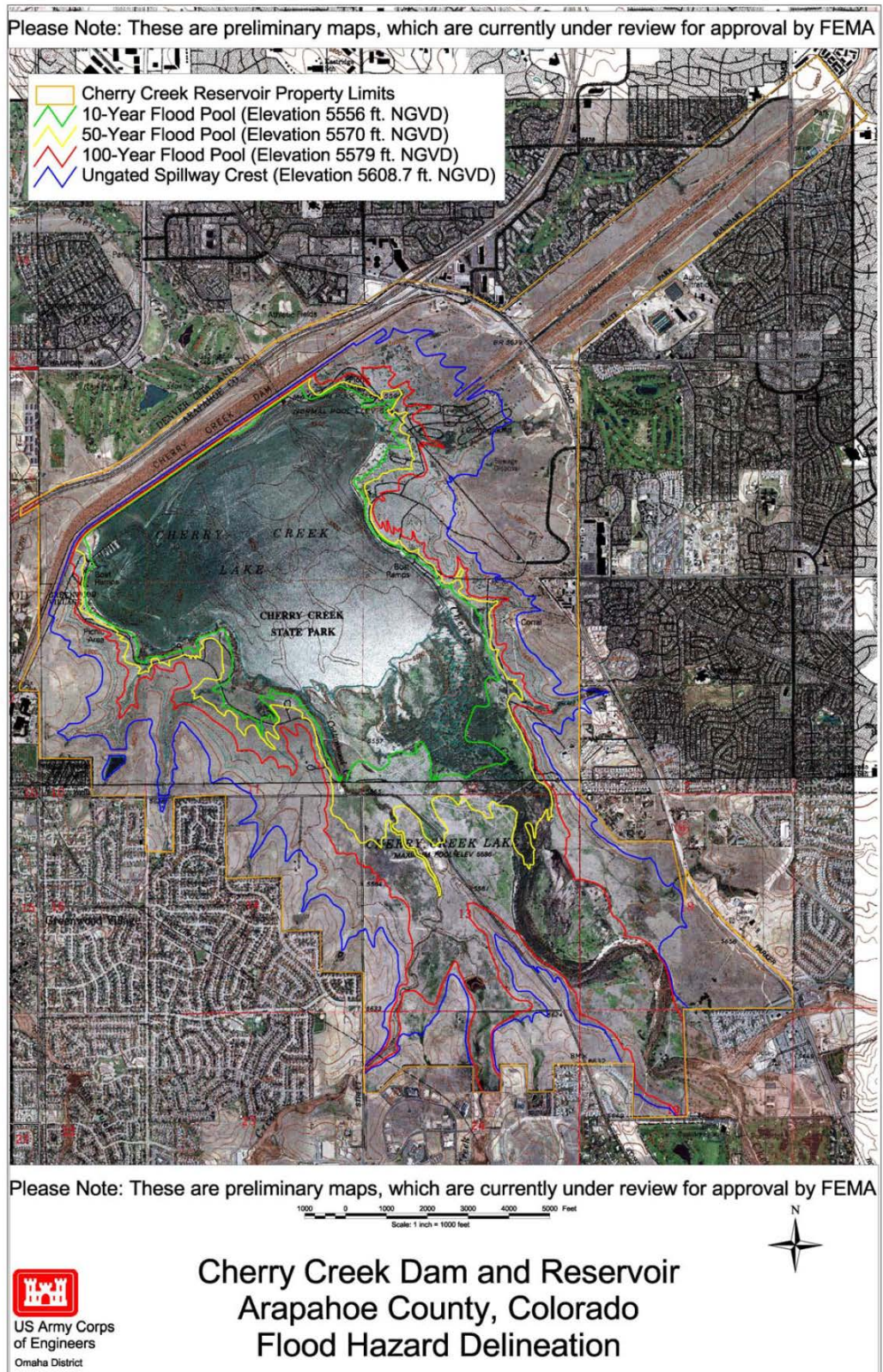


Figure 10. Cherry Creek Dam and Reservoir Flood Hazard Delineation

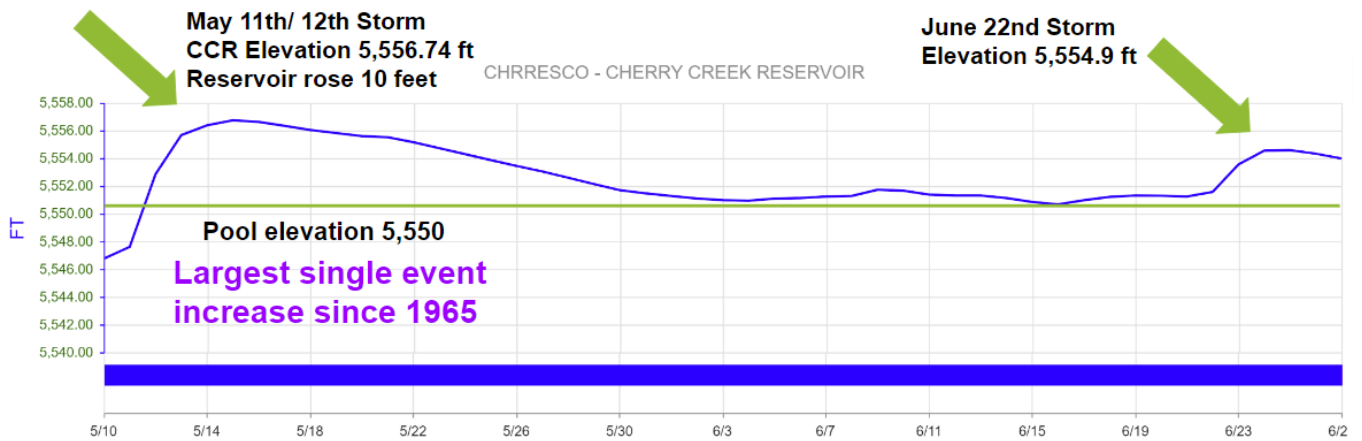


Figure 11. Reservoir Elevation May- June 2023 (CHRRESCO) (Colorado DWR).

3.3.1 CHERRY CREEK

CC-10 is the site upstream on Cherry Creek just before it enters the Reservoir, and it is representative of inflow water quality. The other sites on Cherry Creek and monitoring results are discussed in Section 3.4 below.

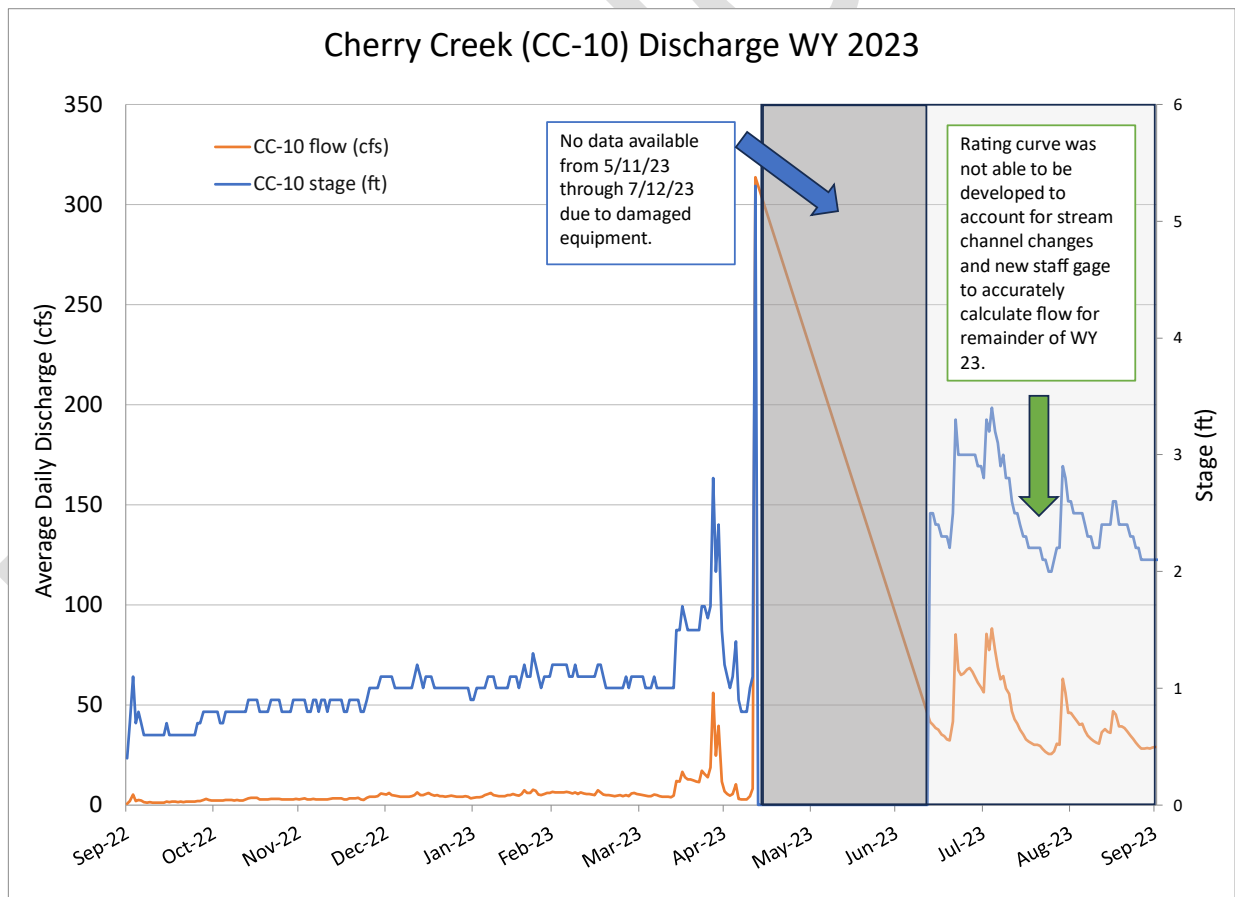
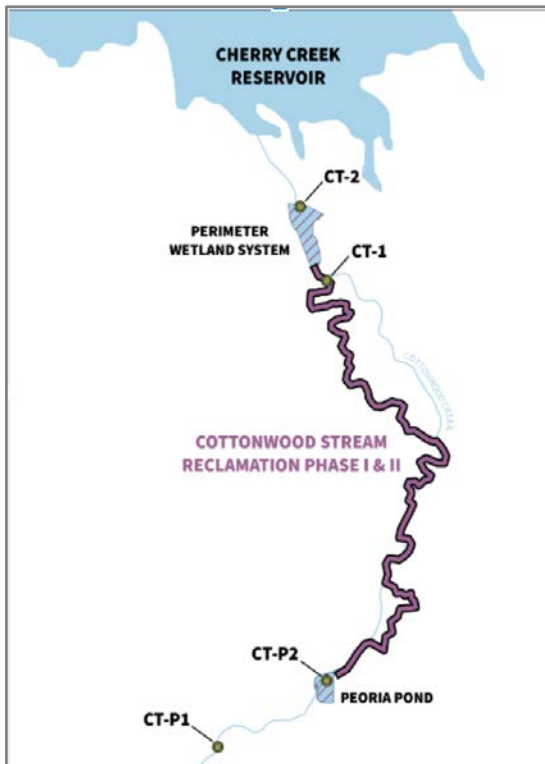


Figure 12. Cherry Creek Discharge at CC-10 upstream of Cherry Creek Reservoir.

Due to the major storm events in the Cherry Creek Watershed during 2023, stage measurements upstream of Cherry Creek Reservoir on Cherry Creek (CC-10) were not able to be collected following the May 11th storm

event due to damaged equipment. Although the pressure transducer-based level sensor and staff gage at the CC-10 were re-installed in mid-July, flow calculations require an updated rating curve due to the changes in stream channel and relocation of the staff gauge. A rating curve at CC-10 could not be generated since the water levels in the reservoir created a backwater effect at the site and the damage to the road upstream impacted the collection of accurate manual flow measurements to develop a new stage-discharge relationship in WY 2023. Stage and flow from CC-10 with the conditions affecting the values are displayed in Figure 12.

3.3.2 COTTONWOOD CREEK



Cottonwood Creek is the second largest surface water input to Cherry Creek Reservoir. Cottonwood Creek has a sub-basin of 9,050 acres. Compared to Cherry Creek, Cottonwood Creek sub basin has more developed land use, and multiple wastewater discharges. There are four monitoring sites on Cottonwood Creek. There are two sites upstream on Cottonwood Creek off Peoria St. and two sites in Cherry Creek State Park. These sites are monitored regularly and CT-1, CT-2, CT-P1, and CT-P2 have equipment to monitor stream levels and collect storm samples upstream and downstream of the PRF ponds and wetland systems (Figure 13).

CT-2 is the site upstream on Cottonwood Creek just before it enters the Reservoir, and it is representative of inflow water quality. The other Cottonwood Creek sites are discussed regarding the evaluation of the effects of the PRFs in Section 3.5 below.

The stage measurements at Cottonwood Creek at CT-2 were not accurately captured during one or more high flow events in WY2023 due to the location of the level sensor in the pond outlet structure. The max stage recorded at CT-2 on Cottonwood Creek

Figure 13. Cottonwood Creek Monitoring Locations and PRFs

was 2.6 ft on May 15, 2023, however on May 11, 2023, Cottonwood Pond overflowed the side of the pond to the East. Since the level sensor is located in the weir outlet structure, it does not accurately record the level in the pond during higher flow events (>2.2ft). In 2024, the survey of the pond elevations will be reviewed to see if level sensor at another location may more accurately represent high flow events. WY 2023 stage and flow from CT-2 with the conditions affecting the values are displayed in Figure 14.

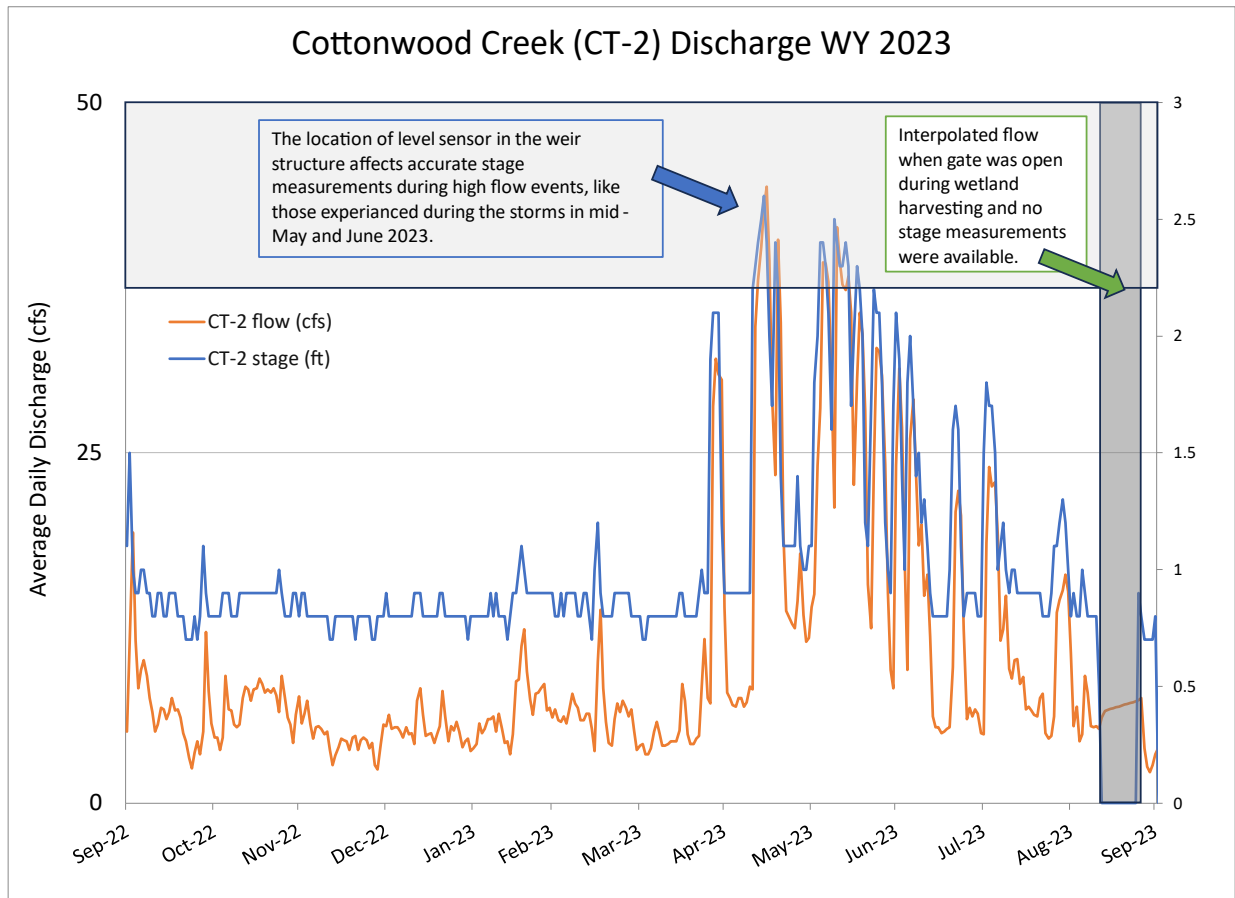


Figure 14. Cottonwood Creek Discharge at CT-2 upstream of Cherry Creek Reservoir.








3.4 WATERSHED WATER QUALITY

The SAP includes monitoring of all the stream sites either monthly or every other month, as well as during storm events (Figure 2). In addition, all sites along Cherry Creek from upstream to downstream are monitored two times per year in the spring and fall. Table 1 outlines the locations, monitoring frequency, and POR that each of the stream sites has been monitored. The sections below outline the major parameters monitored, summary statistics, notable seasonal variation, and major trends observed using an MK analysis (see section 2.2) during the POR for each site.

Table 1. Watershed Monitoring Locations, Frequency, and Period of Record.

B-Bi-annual, EO – Every other Month, M-Monthly, 💧 - Storm

Location Name	#/Yr	LOCID	Earliest Sampling Event	Most Recent Sampling Event	POR (Years)
CC-USGSFRANKTOWN	B	USGS-Franktown	8/11/1994	5/3/2023	29
CC-1 - Cherry Creek Station 1	B	CC-1	8/10/1994	5/3/2023	29
CC-2 - Cherry Creek Station 2	B	CC-2	11/8/1994	5/3/2023	29
CC-USGSPARKER	B	USGS-Parker	5/9/2017	5/3/2023	6
CC-4 - Cherry Creek Station 4	B	CC-4	8/10/1994	5/3/2023	29
CC-5 - Cherry Creek Station 5	B	CC-5	8/9/1994	5/3/2023	29
CC-6 - Cherry Creek Station 6	B	CC-6	8/9/1994	5/3/2023	29
CC-7 - Cherry Creek Station 7	M / 💧	CC-7	5/15/2012	9/13/2023	11

CC-8 - Cherry Creek Station 8	B	CC-8	3/15/1995	5/3/2023	28
CC-9 - Cherry Creek Station 9	B	CC-9	8/8/1994	5/3/2023	29
CC-10 - Cherry Creek Station 10	M / 	CC-10	4/3/1992	9/13/2023	31
CC-Out - Cherry Creek Reservoir Outflow	M	CC-Out	4/3/1992	9/13/2023	31
CT-1 - Cottonwood Creek PRF Site 1	M / 	CT-1	4/9/1992	9/13/2023	31
CT-2 - Cottonwood Creek PRF Site 2	M / 	CT-2	4/2/1996	9/13/2023	27
CT-P1 - Cottonwood Creek PRF Site P1	M / 	CT-P1	5/24/2002	9/13/2023	21
CT-P2 - Cottonwood Creek PRF Site P2	M / 	CT-P2	2/20/2002	9/13/2023	21
MCM-1 - McMurdo Gulch Station 1	EO	MCM-1	1/18/2012	8/9/2023	11
MCM-2 - McMurdo Gulch Station 2	EO	MCM-2	1/18/2012	8/9/2023	11
PC-1 - Piney Creek	M / 	PC-1	4/25/2018	9/13/2023	5
Rain Sampler		PRECIP	4/4/2014	7/5/2023	9
MW-1 Monitoring Well 1	B	MW-1	8/10/1994	5/3/2023	29
MW-5 Monitoring Well 5	B	MW-5	8/16/1994	5/3/2023	29
MW-9 Monitoring Well	B	MW-9	8/12/1994	5/3/2023	29
MW-Kennedy Monitoring Well	B	MW- Kennedy	6/1/1999	5/4/2023	24

3.5.1 PHYSICAL PARAMETERS

The stream sites in the Cherry Creek Watershed are monitored monthly for physical conditions such as temperature, pH, dissolved oxygen, and conductivity which indicate major changes in water chemistry upstream to down and between streams and tributaries.

TEMPERATURE

The water temperatures in the streams monitored monthly in the Cherry Creek watershed vary seasonally and between locations. Overall, the sites on Cherry Creek (CC) and Piney Creek (PC) demonstrate less temperature variability than the sites on Cottonwood Creek (CT). The median water temperature in 2023 was at or below the baseline medians at all sites, except for the most upstream site, CT-P1, where it was slightly higher.

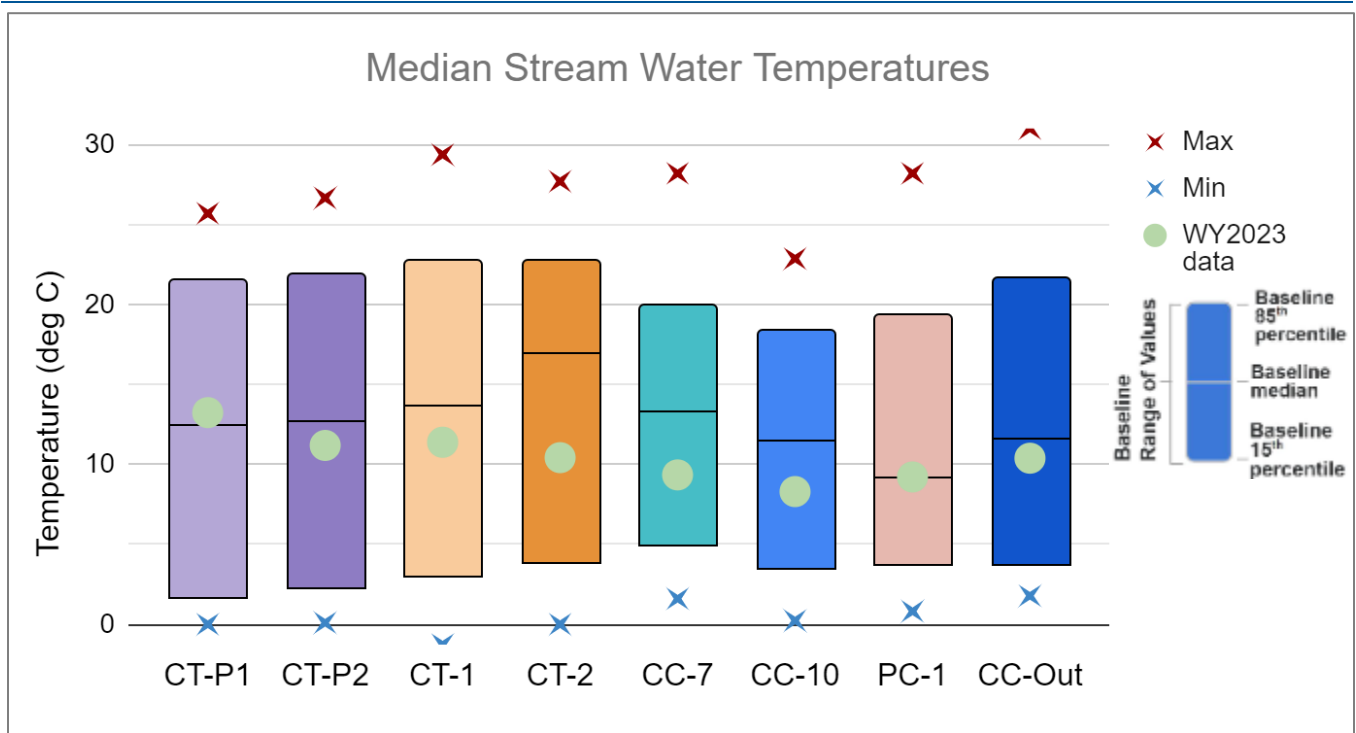


Figure 15. Stream Temperature Summary Statistics and WY 2023 medians.

PH

The pH in streams can affect aquatic life as well as alter the behavior of other compounds in the water. Often, major changes in pH can be traced back to human activities in the watershed, but plants and algae can also increase the pH as they remove carbon dioxide from the water during photosynthesis.

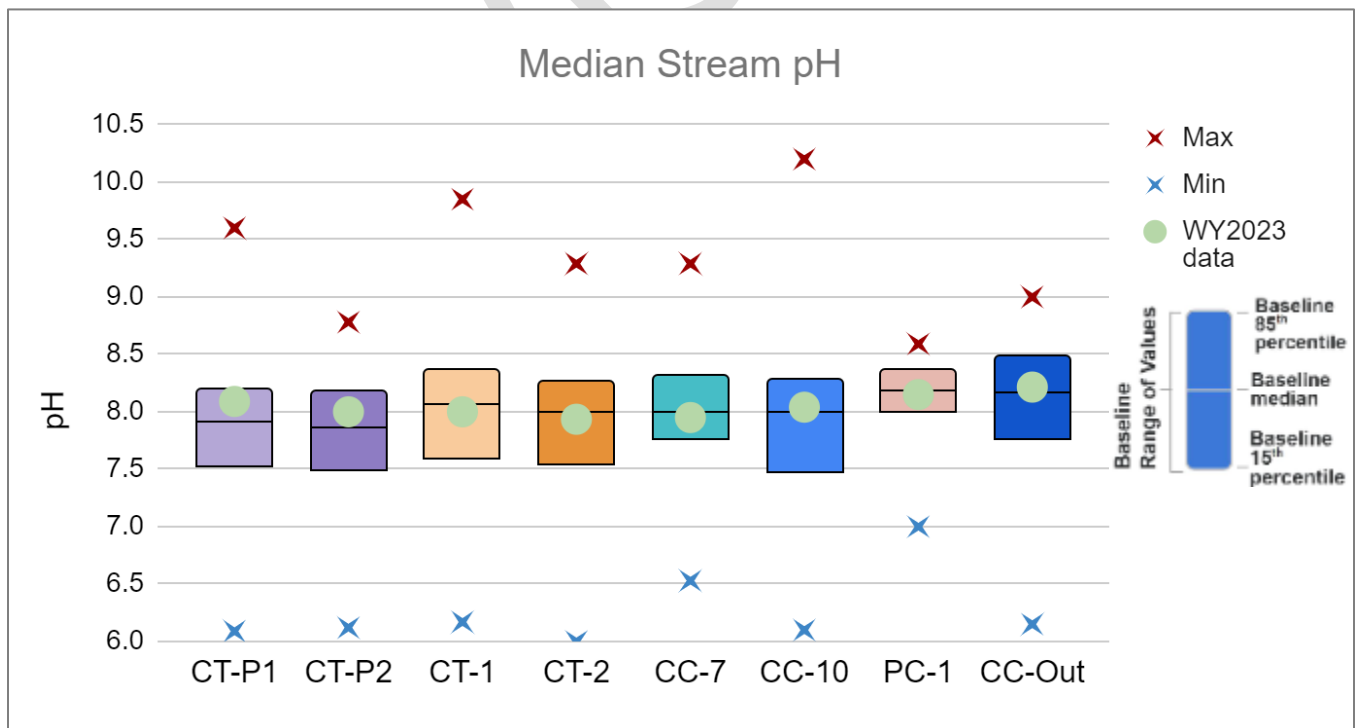


Figure 16. Stream pH Summary Statistics – POR Median and 15th/85th percentiles and WY 2023 Median

As illustrated in Figure 16, the pH in the streams monitored monthly during WY 2023 did not demonstrate any major differences spatially or temporally.

UP TO DOWNSTREAM CHERRY CREEK

Figure 17 shows the pH upstream to downstream on Cherry Creek from the bi-annual monitoring events from WY 2023 along with POR summary statistics. pH was similar or higher than November 2022 during May 2023 except for the outlet of the Reservoir, which was higher in November 2023. These pH values correlate with the fact that pH tends to be higher during the warmer months (e.g., May) as biological productivity in the water increases.

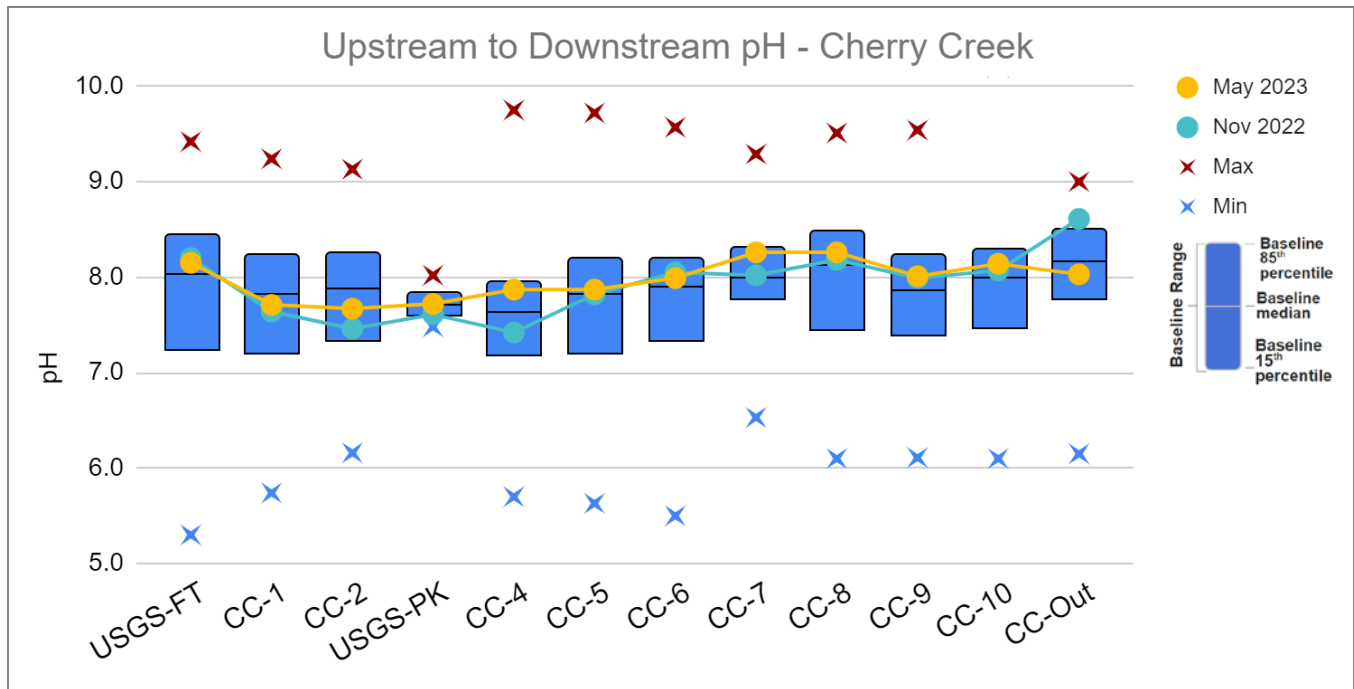


Figure 17. pH Upstream to Downstream on Cherry Creek, Baseline data 1994-2023 and WY 2023 – Nov 2022 and May 2023.

DISSOLVED OXYGEN

Dissolved oxygen in the water is required for aquatic life and generally decreases as water temperatures increase in the warmer months. The DO concentrations in the watershed demonstrate some variability seasonally and between sites (Figure 18.)

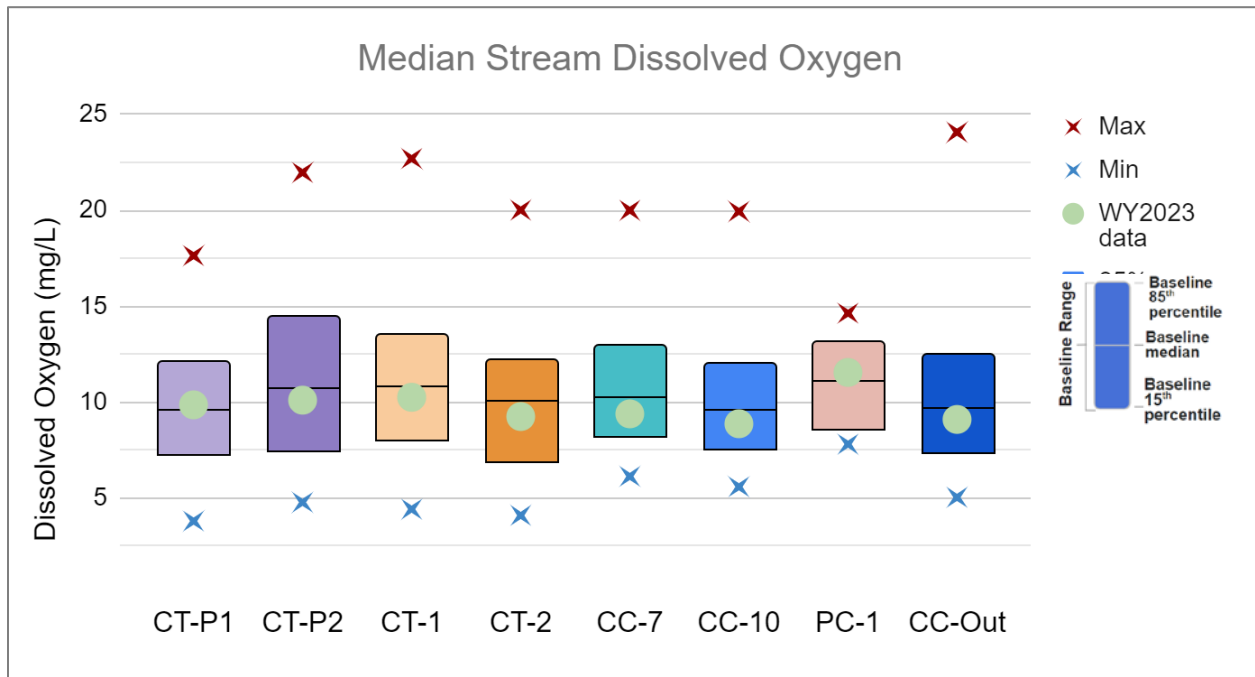


Figure 18. Dissolved Oxygen Concentration Summary Statistics - POR Median, 15th/85th percentile and WY 2023 Median.

MONTHLY STREAM SITES THROUGH THE WATERSHED

The baseline median DO concentration of the monthly stream sites is around 10 mg/L, and the WY 2023 median is slightly lower at 9.6mg/L.

UP TO DOWNSTREAM CHERRY CREEK

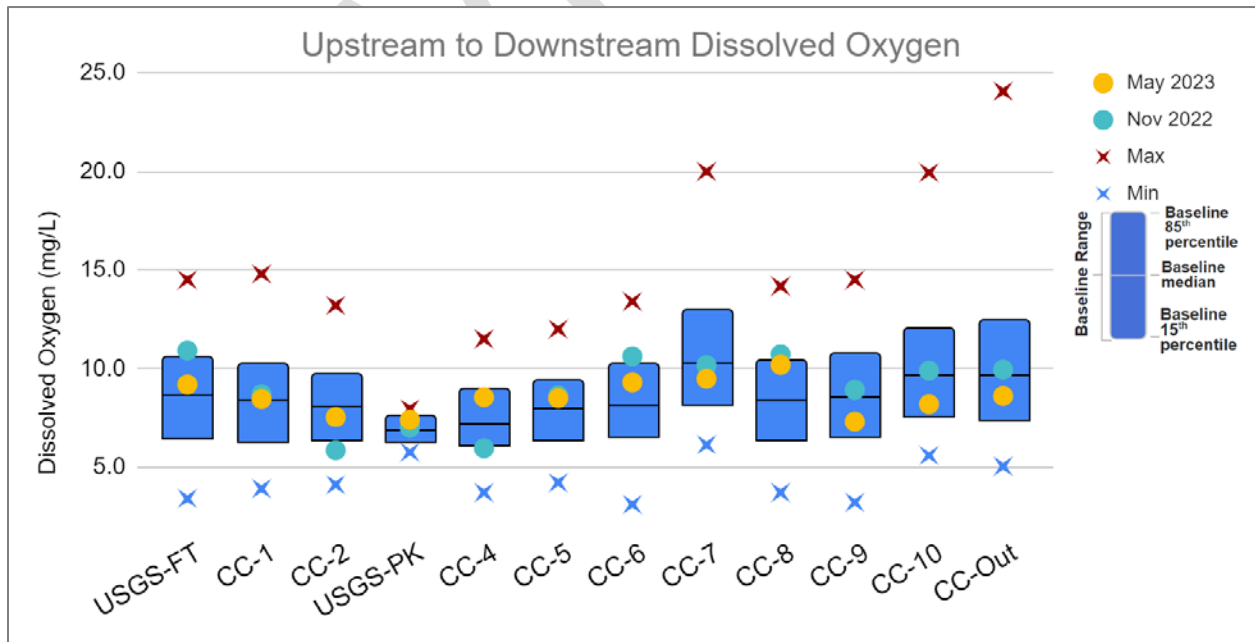


Figure 19. Dissolved Oxygen Concentrations Upstream to Downstream on Cherry Creek, Summary Statistics for POR and WY 2023 – Nov 2022 and May 2023.

Because higher water temperature decreases the solubility of oxygen in water, higher concentrations are usually observed in the colder months. In WY 2023, higher dissolved oxygen concentrations were observed in

November during the bi-annual upstream to downstream monitoring event on Cherry Creek, except for CC-2, USGS-Parker, and CC-4, which were higher in May (Figure 19).

CONDUCTIVITY

The conductivity, measured as specific conductance, which indicates dissolved solids (i.e., salts minerals, etc.) demonstrates spatial variability within the Cherry Creek watershed. Although there are no conductivity standards for streams in the basin, the US EPA considers levels above 1,500 $\mu\text{S}/\text{cm}$ above average for most streams in the US.

Figure 20 depicts the specific conductance at the sites monitored monthly over the entire period of record as well as the median values observed in WY 2023, with the EPA value displayed on the graph as a benchmark reference. Over the POR, the highest conductivity values are observed at the furthest upstream sites (CT-P1 and CT-P2) on Cottonwood Creek and decrease downstream towards the Reservoir. High conductivity has also been recorded on Piney Creek although the POR is shorter (2019-present). The lowest conductivity values are observed upstream on Cherry Creek at CC-7 and increase downstream at CC-10, just upstream of the Reservoir. The median conductivity at the outlet is slightly higher than Cherry Creek but lower than Cottonwood due to the relative inflow concentrations and mixing that occurs in the Reservoir. The WY 2023 median conductivity is similar to the baseline median at CC-7 but the WY 2023 medians are higher than the baseline medians for all other sites.

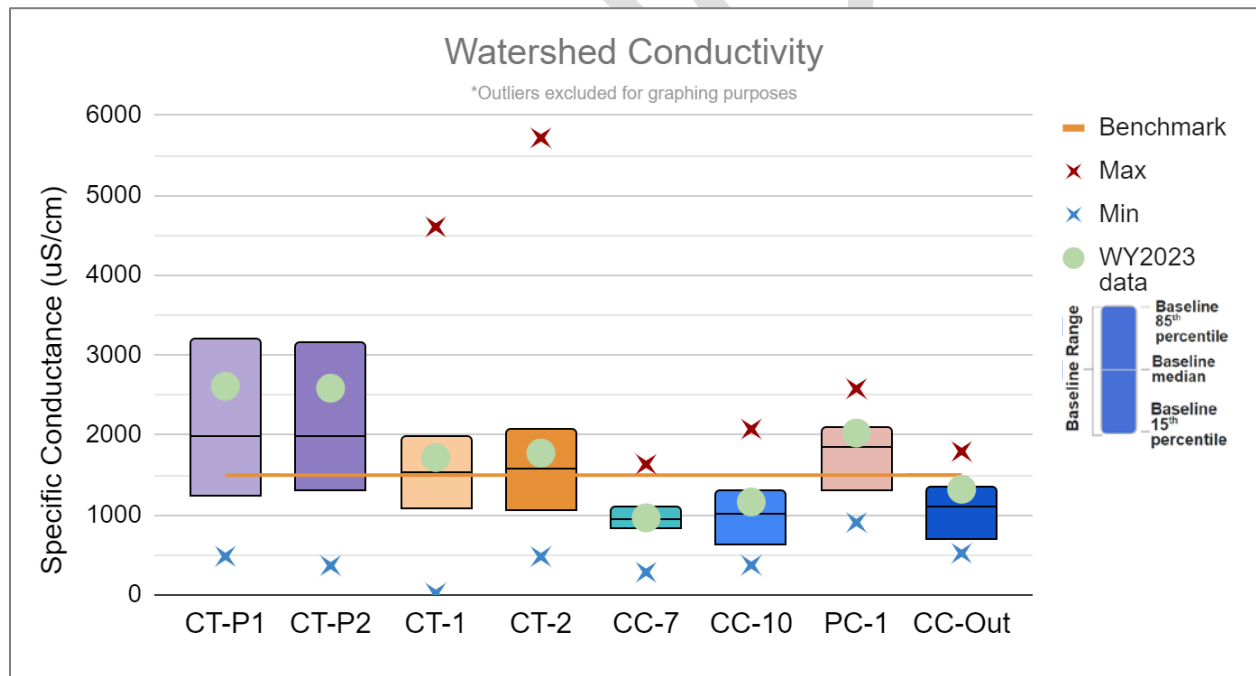


Figure 20. Watershed Stream Conductivity, Summary Statistics for POR and WY 2023 median.

MONTHLY STREAM SITES THROUGH THE WATERSHED

Within the watershed, conductivity varies seasonally; February has the highest historical maximum conductivity and January having the maximum in WY 2023 in Cherry Creek, Cottonwood Creek, and Piney Creek (Figure 21, Figure 22, and Figure 23). Lower conductivity values are observed during the summer months. The median conductivity on Cherry Creek was below the 1500 $\mu\text{S}/\text{cm}$ EPA criteria during WY 2023 (Figure 21) but the median values exceeded this threshold on Cottonwood Creek all months except June, July, and September (Figure 22). The conductivity on Piney Creek demonstrated a similar pattern with only June and August values below the

benchmark. Notably, the fall winter and spring months of October through April, with the exception of February, all had conductivity at or near the maximum observed since monitoring started on Piney Creek in 2018 (Figure 23).

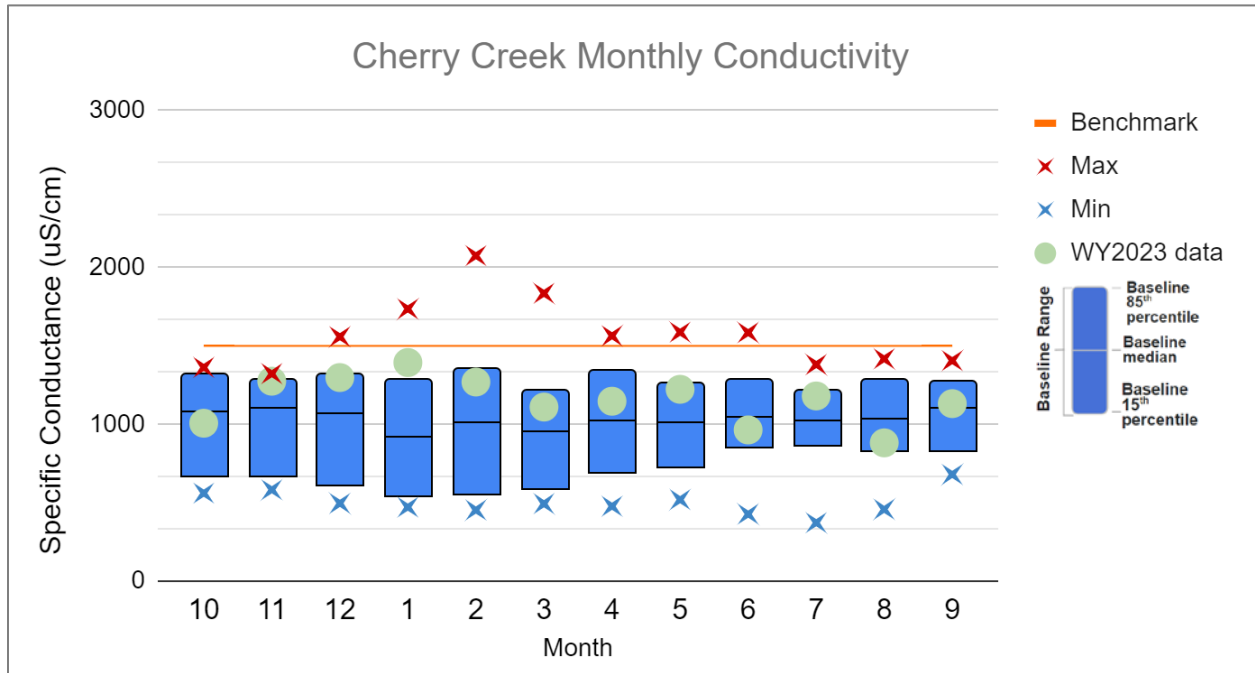


Figure 21. Monthly Conductivity on Cherry Creek at CC-10, POR Summary Statistics, and WY 2023.

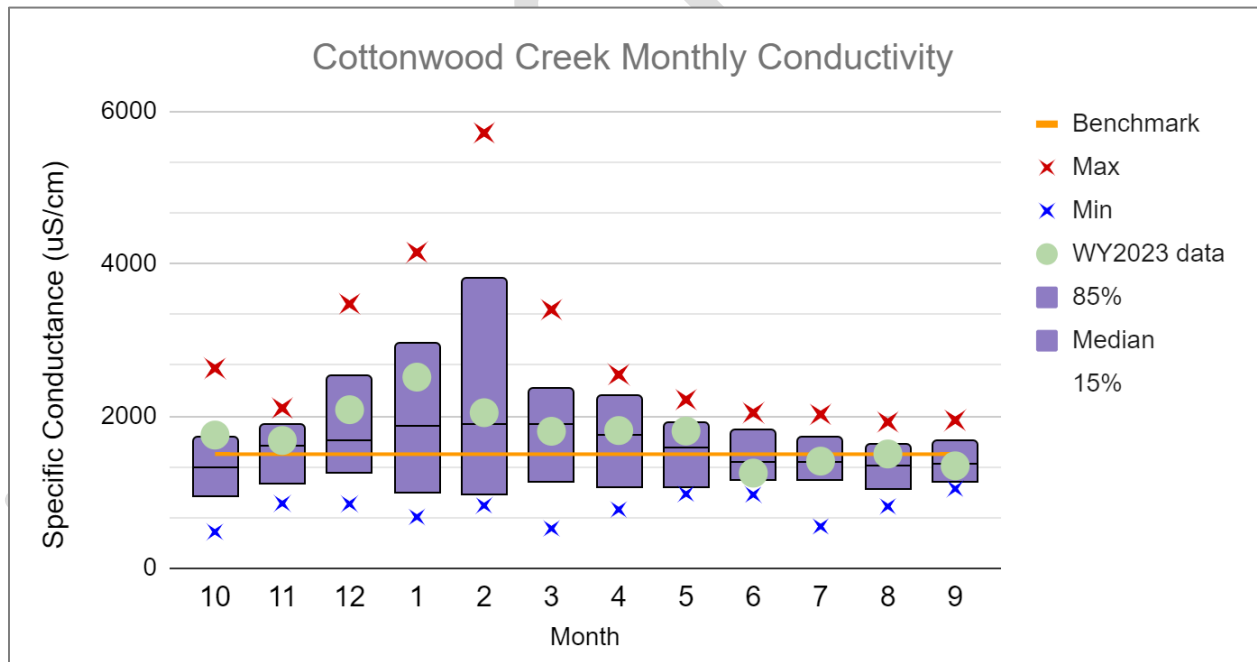


Figure 22. Monthly Conductivity on Cottonwood Creek at CT-2, POR Summary Statistics, and WY 2023.

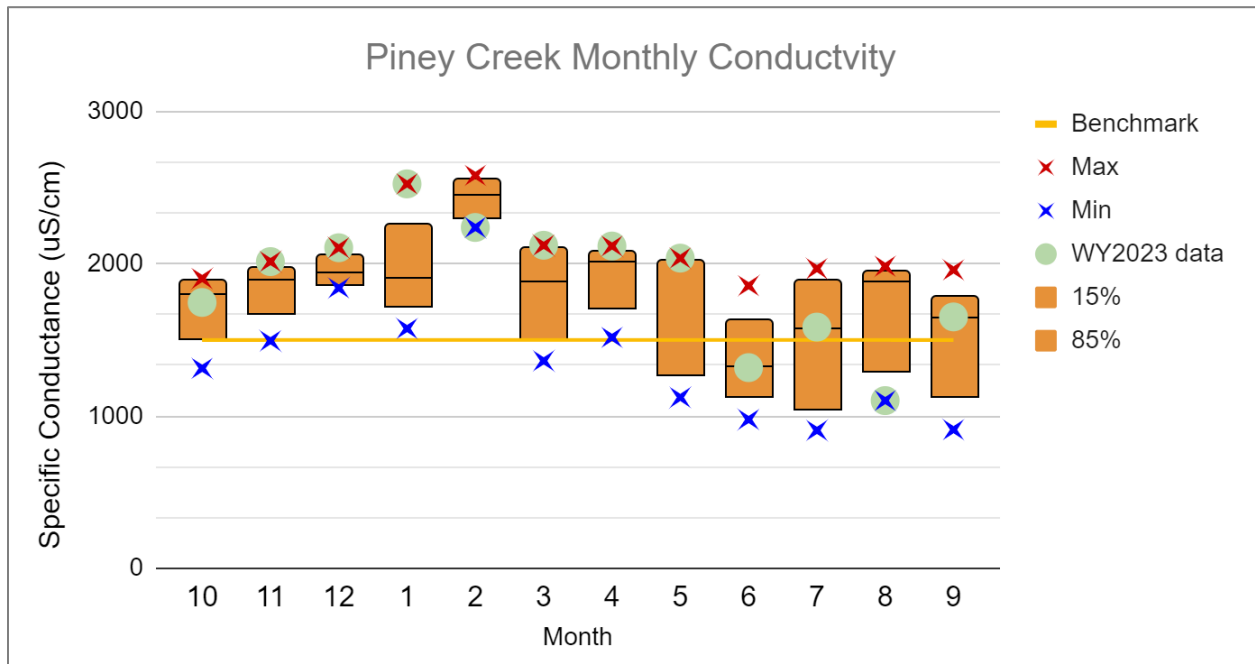


Figure 23. Monthly Conductivity on Piney Creek at PC-1, POR Summary Statistics, and WY 2023.

UP TO DOWNSTREAM CHERRY CREEK

Figure 24 illustrates the median conductivity upstream to downstream measurements in Nov 2022 and May 2023 on Cherry Creek along with the 1994 to 2023 POR summary statistics. A MK trend analysis determined that the baseline and WY 2023 median conductivity significantly increases upstream to downstream (Figure 24). In addition, a MK trend analysis demonstrates that the increasing trend of the annual mean conductivity of inflows to the Reservoir (Cherry Creek at CC-10 and on Cottonwood Creek at CT-2) is significant (Figure 25).

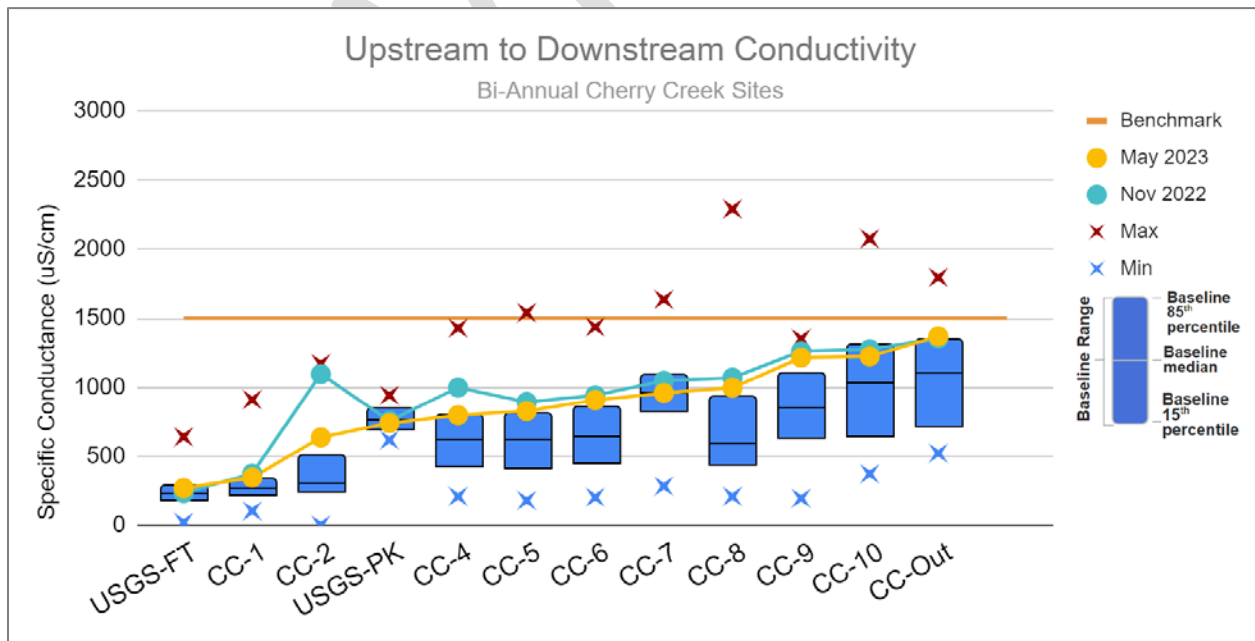


Figure 24. Conductivity Upstream to Downstream on Cherry Creek, Summary Statistics for POR and WY 2023 – Nov 2022 and May 2023.

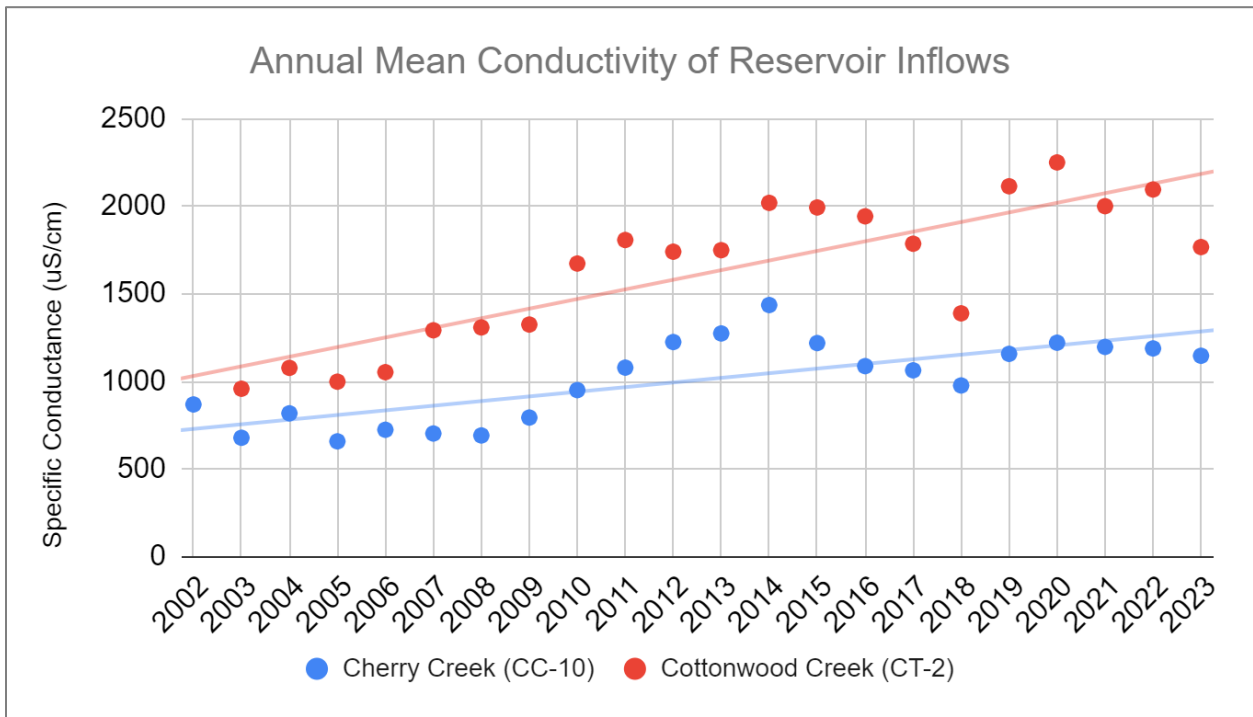


Figure 25. Historical Mean Conductivity on Cherry Creek and Cottonwood Creek.

NUTRIENTS AND SUSPENDED SOLIDS

Nutrients and suspended solids in the streams in the Cherry Creek Watershed have a direct impact on the water quality in the Reservoir. Nutrients demonstrate variable patterns and trends between sites and flow patterns. High stream flow increases suspended particles in the water which is directly correlated to increased phosphorus concentrations.

PHOSPHORUS

Figure 26 and Table 2 show the total phosphorus (TP) POR summary statistics and WY 2023 base and stormflow medians for each of the monthly stream sites.. The maximum TP concentrations are observed during storm events with some values excluded for graphing purposes. The WY 2023 median TP concentrations were higher in stormflows than baseflows and were lower than the baseline median at all sites and flow conditions except for the two sites on Cherry Creek (CC-7 and CC-10) in storm conditions and at the outlet to the Reservoir (CC-0). The higher TP concentrations in WY 2023 at these sites can be attributed to the major storm events that caused above average concentrations of suspended solids and phosphorus.

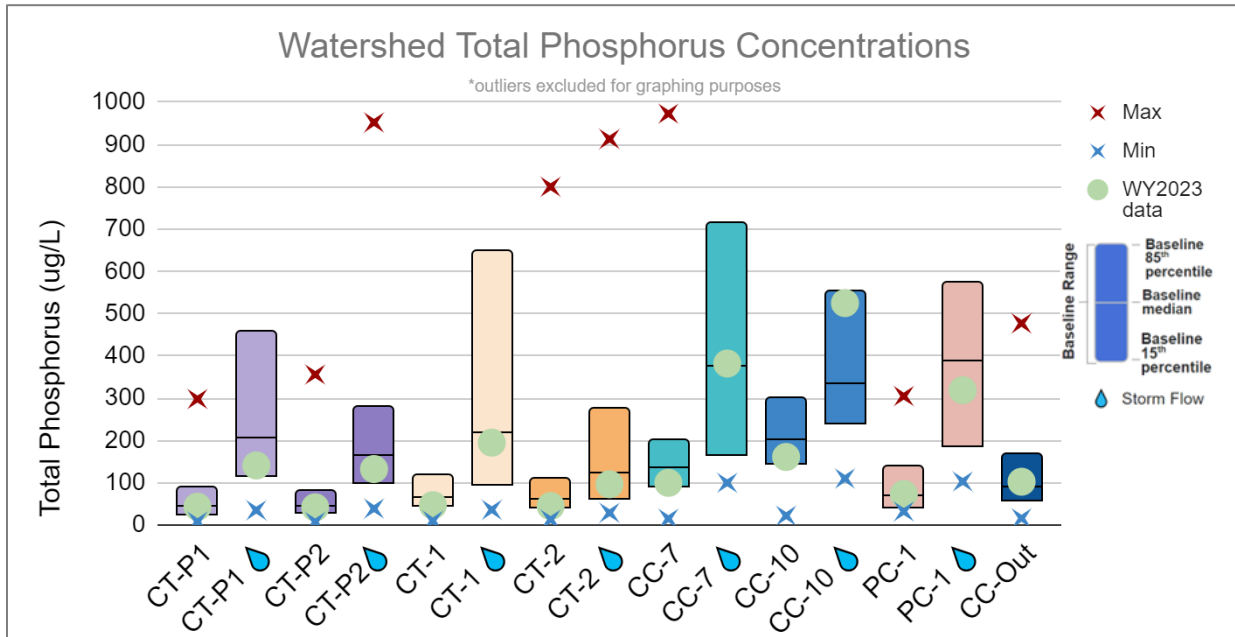


Figure 26. Watershed Phosphorus Concentrations (Base and Stormflow Conditions) POR Summary Statistics, and WY 2023.

Table 2. Total Phosphorus Concentration (µg/L) Baseline Summary Statistics and WY 2023 values, Base and Stormflow Conditions.

Site	Site/ Flow	POR Min	POR Median	POR Max	Count	WY2023 median	Count
CT-P1 - Cottonwood Creek PRF Site P1	CT-P1	8	47	298	240	44	12
CT-P1 - Cottonwood Creek PRF Site P1	CT-P1	35	210	2235	134	141	8
CT-P2 - Cottonwood Creek PRF Site P2	CT-P2	7	50	356	238	42	12
CT-P2 - Cottonwood Creek PRF Site P2	CT-P2	39	168	952	124	133	8
CT-1 - Cottonwood Creek PRF Site 1	CT-1	10	69	1461	370	47	12
CT-1 - Cottonwood Creek PRF Site 1	CT-1	36	222	3570	162	195	8
CT-2 - Cottonwood Creek PRF Site 2	CT-2	13	64	800	349	45	12
CT-2 - Cottonwood Creek PRF Site 2	CT-2	29	127	913	163	97	8
CC-7 - Cherry Creek Station 7	CC-7	15	137	973	124	100	12
CC-7 - Cherry Creek Station 7	CC-7	100	378	2684	43	382	7
CC-10 - Cherry Creek Station 10	CC-10	22	207	2532	378	161	12
CC-10 - Cherry Creek Station 10	CC-10	110	336	3110	145	525	7
PC-1 - Piney Creek	PC-1	32	74	305	60	74	12
PC-1 - Piney Creek	PC-1	103	390	2250	13	319	6
CC-Out - Cherry Creek Reservoir Outflow	CC-Out	16	95	477	340	103	12

Stormflow indicated with after site name.

*Values in *italics* were excluded from Figure 26 for graphing purposes.

UP TO DOWNSTREAM CHERRY CREEK

During the upstream to downstream monitoring events in WY 2023, the TP concentrations were higher in May 2023 than November 2022 (Figure 27). Both events had TP concentrations that were lower than the respective baseline medians except for CC-7 and the outlet (CC-0) to the Reservoir in May.

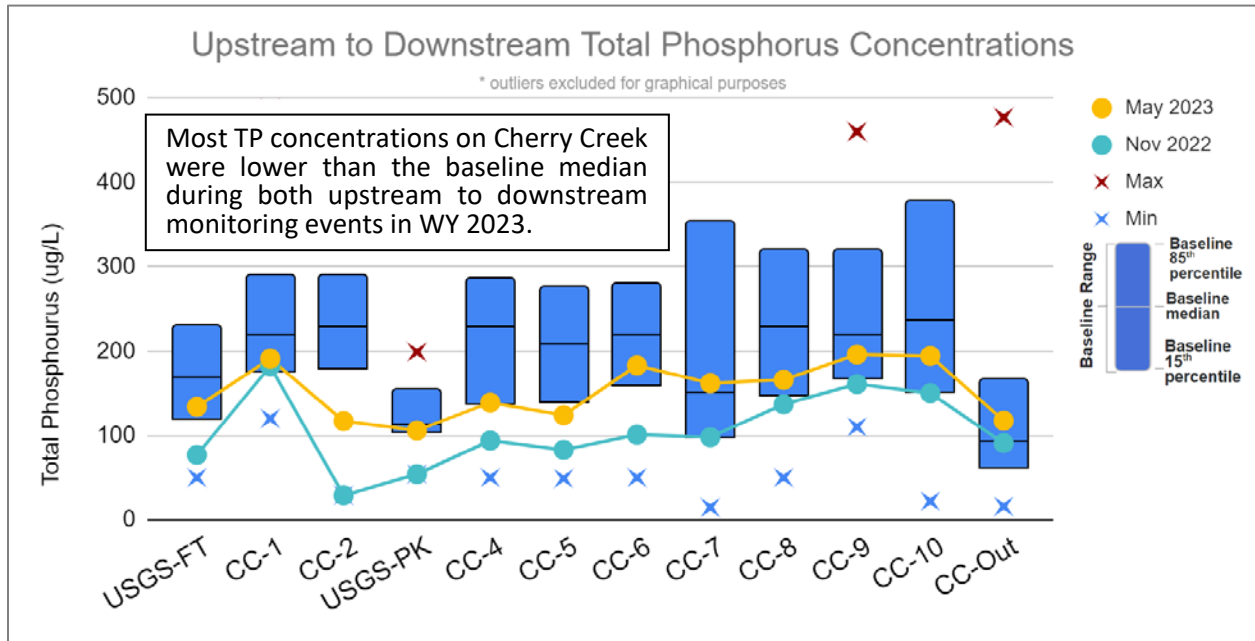


Figure 27. Upstream to Downstream Total Phosphorus Concentrations on Cherry Creek, Summary Statistics for POR and WY 2023 – Nov 2022 and May 2023.

NITROGEN

Nitrogen concentrations in the streams vary spatially throughout the watershed, seasonally and with different flow patterns. Figure 28 and Table 3 show the total nitrogen (TN) POR summary statistics and WY 2023 base and stormflow medians for each of the monthly stream sites. In contrast to TP, the maximum TN concentrations were not always observed during storm events (Table 2). The WY 2023 median TN concentrations were higher than the baseline median at three sites on Cottonwood Creek (CT-P1, CT-1, and CT-2) during baseflows and during storm events at CT-2. The WY 2023 median TN on Cherry Creek at CC-10 and the outlet to the Reservoir (CC-0) were also higher than the baseline medians during baseflow conditions.

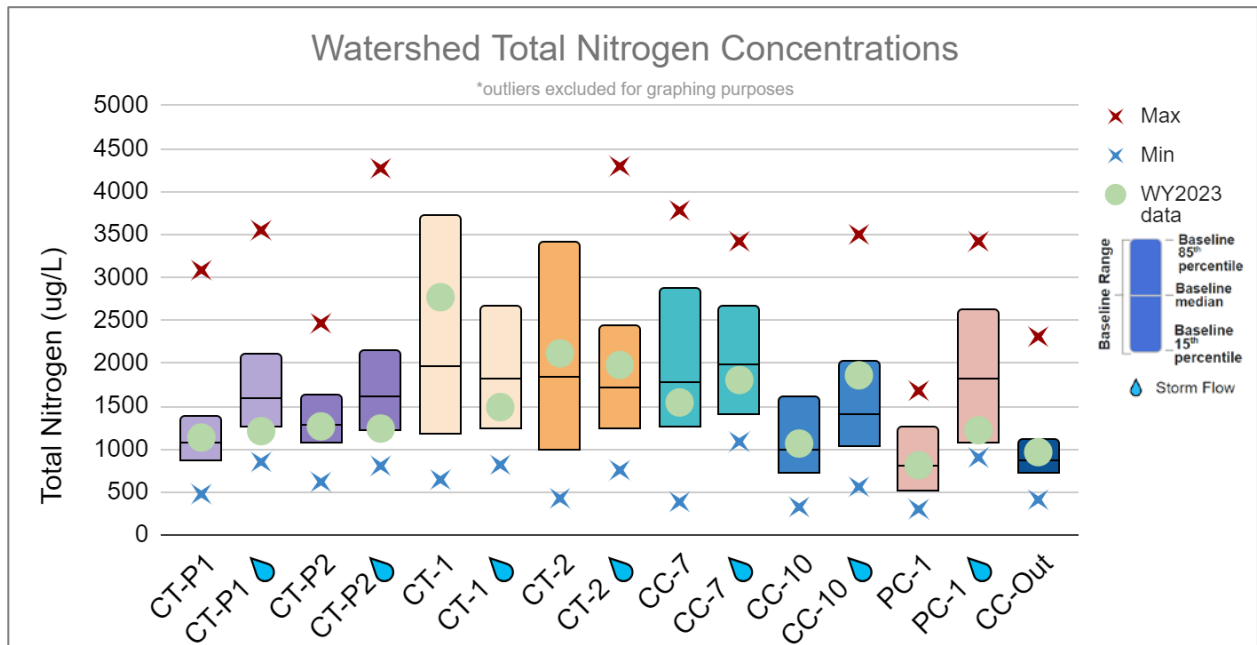


Figure 28. Watershed Nitrogen Concentrations (Base and Stormflow Conditions) Baseline Summary Statistics, and WY 2023.

Table 3. Total Nitrogen Concentration (µg/L) Baseline Summary Statistics and WY 2023 values, Base and Stormflow Conditions.

Site	Site/ Flow	Min	Median	Max	Count	WY2023 median	Count
CT-P1 - Cottonwood Creek PRF Site P1	CT-P1	477	1095	3084	239	1135	12
CT-P1 - Cottonwood Creek PRF Site P1	CT-P1 💧	851	1607	3550	133	1210	7
CT-P2 - Cottonwood Creek PRF Site P2	CT-P2	619	1294	2466	237	1265	12
CT-P2 - Cottonwood Creek PRF Site P2	CT-P2 💧	806	1615	4270	123	1240	7
CT-1 - Cottonwood Creek PRF Site 1	CT-1	645	1986	6300	301	2770	12
CT-1 - Cottonwood Creek PRF Site 1	CT-1 💧	818	1840	7670	129	1490	7
CT-2 - Cottonwood Creek PRF Site 2	CT-2	428	1858	5761	297	2115	12
CT-2 - Cottonwood Creek PRF Site 2	CT-2 💧	756	1733	4295	147	1980	7
CC-7 - Cherry Creek Station 7	CC-7	386	1800	3780	119	1544	12
CC-7 - Cherry Creek Station 7	CC-7 💧	1086	1988	3420	42	1805	6
CC-10 - Cherry Creek Station 10	CC-10	327	1002	7980	312	1065	12
CC-10 - Cherry Creek Station 10	CC-10 💧	562	1422	3500	122	1860	6
PC-1 - Piney Creek	PC-1	301	822	1680	59	813	12
PC-1 - Piney Creek	PC-1 💧	902	1840	3420	12	1220	5
CC-Out - Cherry Creek Reservoir Outflow	CC-Out	412	884	2310	291	965	12

Stormflow indicated with 💧 after site name.

*Values in *italics* were excluded from Figure 28 for graphing purposes.

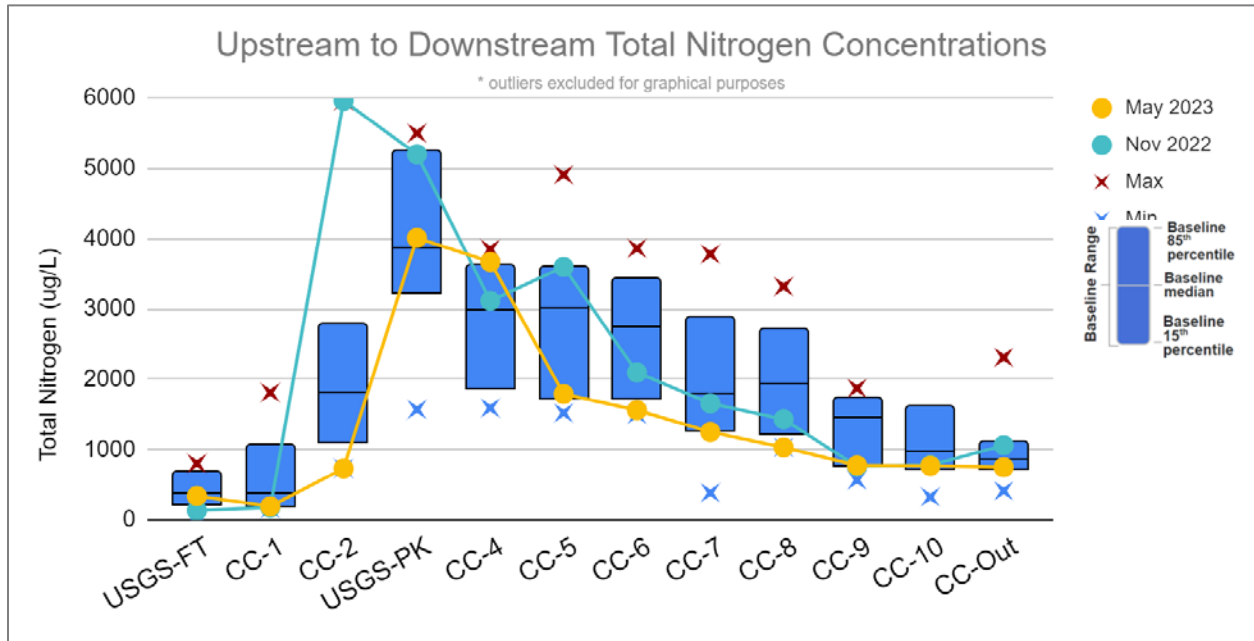


Figure 29. Upstream to Downstream Total Nitrogen Concentrations on Cherry Creek, Summary Statistics for POR and WY 2023 – Nov 2022 and May 2023.

During the upstream to downstream monitoring events in WY 2023, the TN concentrations were usually higher in November 2022 than in May 2023 (Figure 27). TN concentrations were only higher in May at the USGS Franktown site and CC-4. Both events had TN concentrations that followed a similar pattern to the baseline median with concentrations increasing between CC-2 and USGS Parker and then decreasing downstream towards the Reservoir.

SUSPENDED SOLIDS

Concentrations of suspended solids, or particles in the water, vary spatially throughout the watershed, seasonally and with different flow patterns. Figure 30 and Table 4 show the TSS POR summary statistics and WY 2023 base and stormflow medians for each of the monthly stream sites. As expected with high flow, TSS concentrations are higher during storm conditions when fast moving water is likely to pick up particles. The WY 2023 median TSS concentrations were only higher than the baseline medians on Cherry Creek (CC-7 and CC-10) during storm events (Table 4).

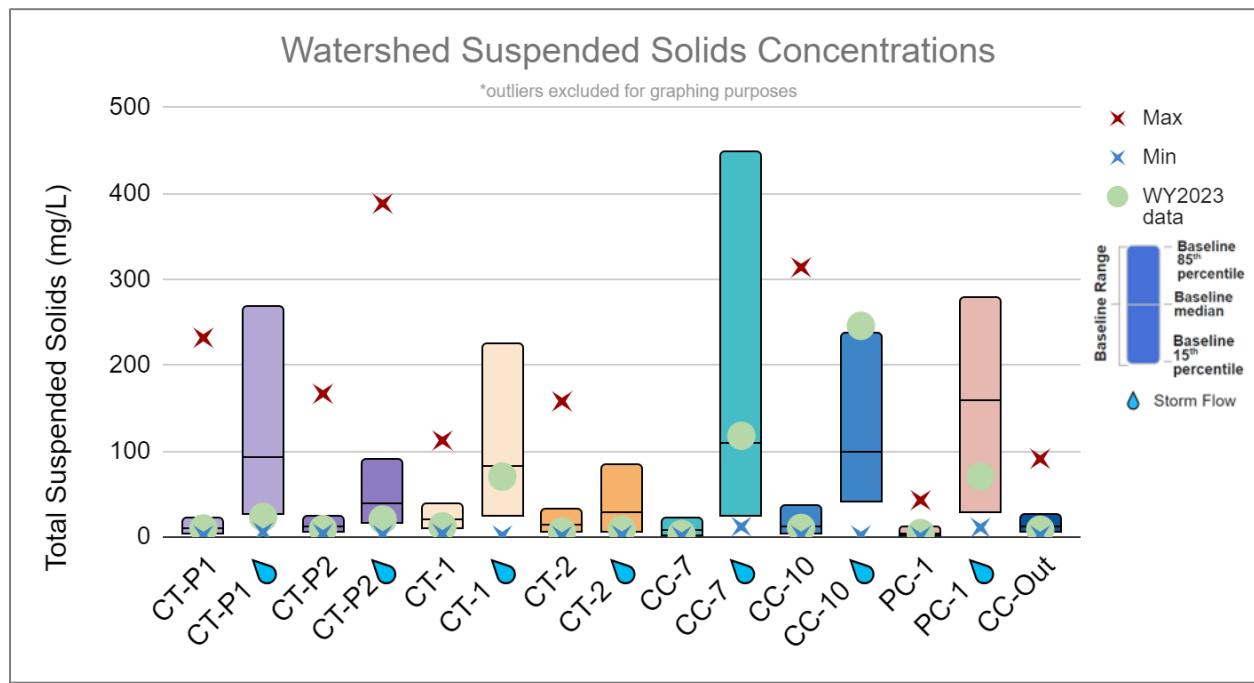









Figure 30. Median Suspended Solids Concentrations (Base and Stormflow Conditions) POR Summary Statistics, and WY 2023.

Table 4. Total Suspended Solids Concentration (mg/L) POR Summary Statistics and WY 2023 values.

Site	Site/ Flow	Min	Median	Max	Count	WY2023 median	Count
CT-P1 - Cottonwood Creek PRF Site P1	CT-P1	2	12	232	173	10	11
CT-P1 - Cottonwood Creek PRF Site P1	CT-P1 	6	94	<i>1053</i>	124	24	8
CT-P2 - Cottonwood Creek PRF Site P2	CT-P2	4	14	167	170	9	12
CT-P2 - Cottonwood Creek PRF Site P2	CT-P2 	3	40	388	114	21	8
CT-1 - Cottonwood Creek PRF Site 1	CT-1	4	22	113	192	13	12
CT-1 - Cottonwood Creek PRF Site 1	CT-1 	2	83	1337	110	71	8
CT-2 - Cottonwood Creek PRF Site 2	CT-2	1	15	158	197	7	12
CT-2 - Cottonwood Creek PRF Site 2	CT-2 	2	31	782	130	9	8
CC-7 - Cherry Creek Station 7	CC-7	1	8	1060	120	4	12
CC-7 - Cherry Creek Station 7	CC-7 	12	110	1360	43	118	7
CC-10 - Cherry Creek Station 10	CC-10	2	14	314	207	11	12
CC-10 - Cherry Creek Station 10	CC-10 	2	101	1660	110	246	7
PC-1 - Piney Creek	PC-1	1	5	43	60	5	12
PC-1 - Piney Creek	PC-1 	11	160	685	13	71	6
CC-Out - Cherry Creek Reservoir Outflow	CC-Out	2	14	91	196	9	12

Stormflow indicated with  after site name.

*Values in *italics* were excluded from Figure 30 for graphing purposes.

3.5 POLLUTANT REDUCTION FACILITIES (PRFS)

The CCBWQA has completed multiple pollutant abatement projects (PAPs), which include PRFs, in various locations through the watershed. WQCC CR 72 states:

"Pollutant Reduction Facility (PRF) means projects that reduce nonpoint source pollutants in stormwater runoff that may also contain regulated stormwater. PRFs are structural measures that include, but are not limited to, detention, wetlands, filtration, infiltration, and other technologies with the primary purpose of reducing pollutant concentrations entering the Reservoir or that protect the beneficial uses of the Reservoir."

The SAP includes an assessment of the effectiveness of selected PRF projects in relation to nutrients and sediment concentrations as water moves downstream. The current monitoring program includes assessment of the PRFs on Cottonwood Creek and McMurdo Gulch. Monitoring of PRFs is conducted in accordance with CR 72.8.1(b).

The Cottonwood Creek PRF is a series of wetland detention systems, along with an area where stream reclamation has been completed, collectively referred to as the Cottonwood Treatment Train (Figure 13). The monitoring program includes water quality samples during routine baseflow sampling and storm conditions above and below these sites.

Table 5. Significant Reductions in Nutrients and Suspended Solids in CCBWQA PRFs, WY 2023 and 2014-2023.*

PRF	Cottonwood Treatment Train		Peoria Pond		Perimeter Pond		Cottonwood Creek btw Ponds		McMurdo Gulch
	Base	Storm	Base	Storm	Base	Storm	Base	Storm	Base
Nitrate+ Nitrite					●				●
Ammonia									●
Nitrogen, Total					●				○
Phosphorus, Soluble Reactive									●
Phosphorus, Dissolved									●
Phosphorus, Total		●		●	●	●			●
Total Suspended Solids	○	●		●	●	●			
Volatile Suspended Solids	●	●				●			

*Legend: ○ reduction of net median downstream in WY 2023, □ significant reductions of net median downstream (2014-2023), ● significant net reduction in WY 2023 and 10-year median downstream, blank cells indicate no significant reduction or an increase upstream to downstream

While the limited results from each water year are often not sufficient to complete a robust statistical analysis, annual calculations are included for reference. This analysis leverages the “PRF Statistics Tool” from the data portal to evaluate the statistical significance of changes above and below PRFs during WY 2023. The tool applies a non-parametric Wilcoxon signed rank test to assess whether differences are present between two data sets, with statistically significant differences indicated by p values less than 0.05.


Table 5 summarizes if the net median upstream-to-downstream changes for each PRF are significant from WY 2023 and how that compares to the significance over the last 10 years (2014-2023) (Section 3.5.1).

Tables 11 through 15 summarize the median upstream and downstream concentrations, the median difference of the paired data, and if the data for the current water year indicates that the median downstream concentrations are significantly lower than the upstream.

Table 6, Table 7, Table 8, and Table 9 summarize the median upstream and downstream concentrations, the median difference of the paired data, and if the data for the current water year indicates that the median downstream concentrations are significantly lower than the upstream in base and stormflows for WY 2023.

During WY 2023, the median concentrations of TSS and VSS downstream of Cottonwood Treatment Train as a whole were lower during both base and storm events (Table 6). The median TN and TDN concentrations were lower downstream during baseflow and stormflow and the NH₃-N and TDP, and TP were lower during storms sampled. The difference of median TP and TSS concentrations downstream during storms was significant. Dissolved nutrient forms are typically harder to remove than particulate forms, which is supported by the water quality from the Cottonwood Creek PRFs.

Table 6. Pollutant Reduction Analysis, Cottonwood Creek Treatment Train PRF, WY 2023.

Cottonwood Treatment Train	Baseflow				Stormflow 			
	CT-P1	CT-2	Upstream to Downstream		CT-P1	CT-2	Upstream to Downstream	
Events (n)	12	12			8	8		
Analyte	Median Concentration		Median Difference	Significant	Median Concentration		Median Difference	Significant
NO ₂ +NO ₃ , µg/L	368	1,250	882		346	670	324	
NH ₃ -N, µg/L	21	42	33		25	19	-9	
TN, µg/L	5	5	1,105		30	44	600	
SRP, µg/L	1,135	2,115	-1		1,210	1,980	-5	
TDP, µg/L	5	5	2		30	44	-4	
TP, µg/L	10	13	-4		44	50	-41	Yes
TSS, mg/L	44	45	-4	Yes	141	97	-13	Yes
VSS, mg/L	10	6	-1		24	9	-3	

When evaluating the two sections individually (Peoria Pond and Perimeter Pond Wetland Systems shown in Table 7 and Table 8), although the Peoria Pond demonstrated reductions in median concentrations of all phosphorus and suspended solid forms in both base and stormflow conditions, the Perimeter Pond demonstrated higher reductions in TP, TSS, and VSS during storm events. The Perimeter Pond also demonstrated lower median concentrations downstream of all forms of NO₂+NO₃ and TN during baseflows. The median concentrations of TP and TSS downstream were significantly lower than upstream during WY 2023 storms sampled is similar to the long-term trends observed over time (Section 3.5.1).

Table 7. Pollutant Reduction Analysis, Peoria Pond PRF, WY 2023.


Peoria Pond	Baseflow				Stormflow 			
Site	CT-P1	CT-P2	Upstream to Downstream		CT-P1	CT-P2	Upstream to Downstream	
Events	12	12			8	8		
Analyte	Median Concentration		Median Difference	Significant	Mean Concentration		Median Difference	Significant
NO ₂ +NO ₃ , µg/L	368	536	104		346	388	47	
NH ₃ -N, µg/L	21	15	-9		25	58	66	
TN, µg/L,	1,135	1,265	180		1,210	1,240	20	
SRP, µg/L	5	4	-1		30	26	-6	
TDP, µg/L	10	9	-1		44	37	-7	
TP, µg/L	44	42	-2		141	133	-11	
TSS, mg/L	10	9	-1		24	21	-2	
VSS, mg/L	3	2	-1		6	6	-1	

Table 8. Pollutant Reduction Analysis, Perimeter Pond PRF, WY 2023.



Perimeter Pond	Baseflow				Stormflow 			
Site	CT-1	CT-2	Upstream to Downstream		CT-1	CT-2	Upstream to Downstream	
Events (n)	12	12			8	8		
Analyte	Median Concentration		Median Difference	Significant	Mean Concentration		Median Difference	Significant
NO ₂ +NO ₃ , µg/L	1,450	1,250	-280	Yes	453	670	342	
NH ₃ -N, µg/L	33	42	4		23	19	16	
TN, µg/L	2,770	2,115	-530	Yes	1,490	1,980	90	
SRP, µg/L	5	5	0		15	44	3	
TDP, µg/L	12	13	0		28	50	3	
TP, µg/L	47	45	-8	Yes	195	97	-112	Yes
TSS, mg/L	13	7	-4	Yes	71	9	-63	Yes
VSS, mg/L	3	2	-1		13	3	-10	

Table 9. Pollutant Reduction Analysis, Cottonwood Treatment Train between the PRF ponds, WY 2023

Cottonwood Ck btwn Pnds	Baseflow				Stormflow 			
Site	CT-P2	CT-1	Upstream to Downstream		CT-P2	CT-1	Upstream to Downstream	
Events (n)	12	12			8	8		
Analyte	Median Concentration		Median Difference	Significant	Mean Concentration		Median Difference	Significant
NO ₂ +NO ₃ , µg/L	536	1,450	972		388	453	-13	
NH ₃ -N, µg/L	15	33	22		58	23	-8	
TN, µg/L	1,265	2,770	1,440		1,240	1,490	150	
SRP, µg/L	4	5	0		26	15	-11	
TDP, µg/L	9	12	2		37	28	-10	
TP, µg/L	42	47	2		133	195	64	
TSS, mg/L	9	13	3		21	71	29	
VSS, mg/L	2	3	0		6	13	3	

There have been multiple stream restoration projects completed on Cottonwood Creek between the Peoria and Perimeter Pond. When evaluating the Cottonwood treatment train between the two ponds (Table 9), the concentrations downstream were the same or higher off all nutrients and suspended solids during baseflow. Although the median NO₂+NO₃, NH₃-N, SRP and TDP were lower in downstream stormflows during WY 2023, the difference was not significant.

One of the upper tributaries of Cherry Creek is McMurdo Gulch, which has multiple reclamation projects completed early in the area’s urbanization to install a proactive PRF designed to protect the gulch and reduce sediment and nutrient loading into Cherry Creek. In addition, over the last few years, other improvements have been completed in various reaches of the same area to further stabilize the channel. Routine water quality samples were collected every other month only under baseflow conditions from monitoring site MCM-1, upstream of the stream reclamation project area, and MCM-2, downstream.

In WY 2023, all median nutrients and suspended solids concentrations were similar or lower downstream of the McMurdo stream reclamation project (Table 10) when compared to the upstream site. The median concentrations of TN and TP were significantly lower downstream in WY 2023 which is similar to the long-term trend observed over the last 10 years (Section 3.5.1).

Table 10. Pollutant Reduction Analysis, McMurdo Gulch, WY 2023.

McMurdo Gulch	Baseflow			Upstream to Downstream
	Site	MCM-1	MCM-2	
Events	6	6		
Analyte	Median Concentration		Median Difference	Significant
NO ₂ +NO ₃ , µg/L	539	36	-403	Yes
NH ₃ -N, µg/L	3	3	0	
TN, µg/L	1,002	613	-432	Yes
SRP, µg/L	261	157	-64	Yes
TDP, µg/L	274	182	-70	Yes
TP, µg/L	335	243	-67	Yes
TSS, mg/L	3	2	0	
VSS, mg/L	1	1	0	

4.5.1 LONG-TERM PRF EVALUATION

The long-term PRF evaluation also evaluates the statistical significance of changes above and below PRFs and over time using non-parametric Wilcoxon signed rank test to assess whether the downstream concentrations are significantly lower than upstream during the period evaluated. Activities such as implementation of CMs and maintenance (e.g., dredging and wetland harvesting) may affect results during various time periods. If more detailed analysis is required to evaluate projects, maintenance activities, or other changes in the watershed, specific evaluations can be completed using the PRF Statistics Tool available on the CCBWQA data portal (<https://www.ccbwqportal.org/prf-statistics-tool>).

Using this tool, an analysis of upstream to downstream concentrations over the last 10 water years (WY 2014-2023) was completed to assess changes (Δ) in median concentrations during baseflow and stormflow conditions. Cottonwood Treatment Train as a whole (Table 11), Peoria Pond (Table 12) and Perimeter Pond (Table 13) all showed statistically significant reductions of TP and TSS during stormflow conditions. Additionally, the Perimeter Pond PRF demonstrated statistically significant reductions in median TP, TN, and TSS concentrations in baseflow conditions as well. There was no significant difference in base or stormflow concentrations upstream to downstream between the two ponds from WY 2014-2023 (Table 10).

For the McMurdo Gulch PRF during WY 2014-2023 (Table 15), the upstream to downstream concentrations of TP and TN during baseflow conditions demonstrated a statistically significant reduction. Statistically significant changes during baseflow conditions were not present for TSS; however, TSS concentrations were extremely low.

Table 11. Pollutant Reduction Analysis of Cottonwood Treatment Train (2014-2023).


PRF	Cottonwood Treatment Train			
Flow Condition	Base		Storm 	
Analyte	Median Δ	Significant	Median Δ	Significant
TP, $\mu\text{g/L}$	1	No	-124	Yes
TN, $\mu\text{g/L}$	720	No	105	No
TSS, mg/L	-2	No	-89	Yes

Table 12. Pollutant Reduction Analysis of Peoria Pond (2014-2023).


PRF	Peoria Pond			
Flow Condition	Base		Storm 	
Analyte	Median Δ	Significant	Median Δ	Significant
TP, $\mu\text{g/L}$	0	No	-26	Yes
TN, $\mu\text{g/L}$	210	No	10	No
TSS, mg/L	1	No	-43	Yes

Table 13. Pollutant Reduction Analysis of Perimeter Pond (2014-2023).


PRF	Perimeter Pond			
Flow Condition	Base		Storm 	
Analyte	Median Δ	Significant	Median Δ	Significant
TP, $\mu\text{g/L}$	-8	Yes	-95	Yes
TN, $\mu\text{g/L}$	-300	Yes	-20	No
TSS, mg/L	-7	Yes	-64	Yes

Table 14. Pollutant Reduction Analysis of Cottonwood Creek Between Ponds (2014-2023).


PRF	Cottonwood Creek Between Ponds			
Flow Condition	Base		Storm 	
Analyte	Median Δ	Significant	Median Δ	Significant
TP, $\mu\text{g/L}$	10	No	58	No
TN, $\mu\text{g/L}$	1020	No	150	No
TSS, mg/L	7	No	45	No

Table 15. Pollutant Reduction Analysis of McMurdo Gulch – 2014-2023 Significance.

PRF	McMurdo Gulch	
Flow Condition	Base	
Analyte	Median Δ	Significant
TP, $\mu\text{g/L}$	-93	Yes
TN, $\mu\text{g/L}$	-176	Yes
TSS, mg/L	2	No

3.2 GROUNDWATER

Groundwater in the Cherry Creek watershed is monitored to gain insight into interactions with surface water and the impacts of groundwater on the Reservoir. Although additional wells have been monitored historically, there are currently four active wells sampled twice a year in the spring and fall. The wells are located throughout the basin, including the top of the basin (MW-1), the middle of the basin (MW-5), and just upstream (MW-9) and downstream of the Reservoir (MW-Kennedy) (Figure 2) that are monitored bi-annually. (Table 1).

5.5.1 GROUNDWATER WATER QUALITY

Groundwater is monitored for physical parameters such as temperature, pH, and dissolved oxygen and chemical composition including nutrients and dissolved solids.

PH

pH in the Cherry Creek Watershed tends to be relatively stable in groundwater, ranging between 6 and 8.5. Although there has been more variability in the pH of the monitoring wells historically, the pH during both upstream to downstream monitoring events were within or near the 15th and 85th percentile baseline ranges (Figure 31).

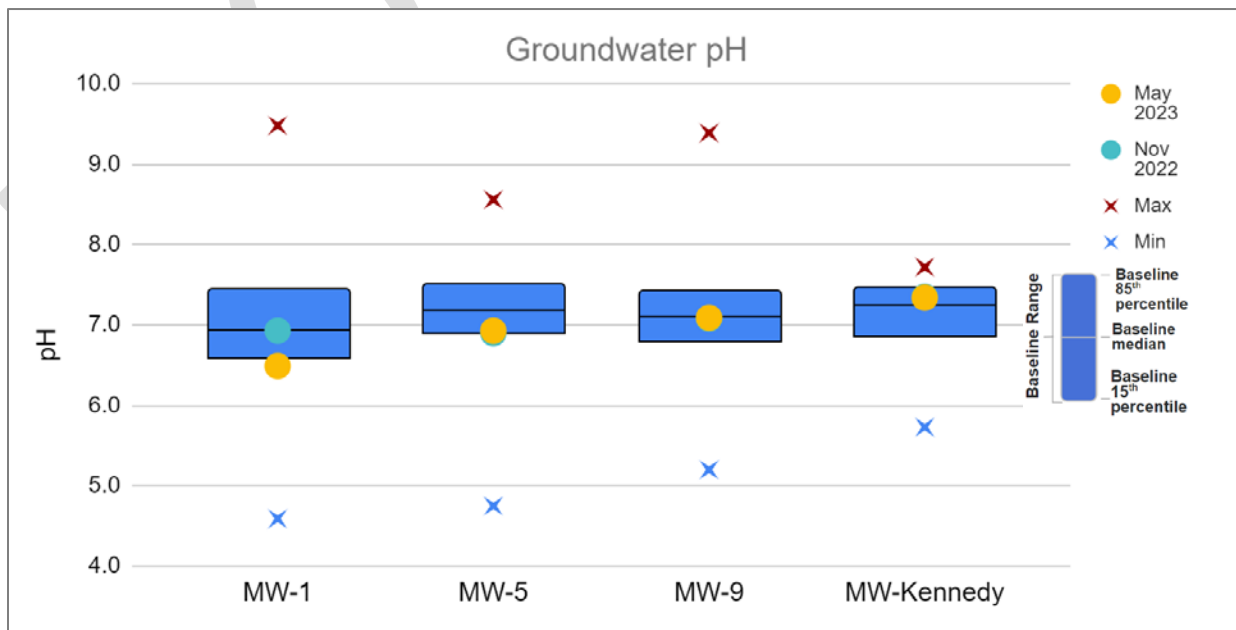


Figure 31. Median pH Groundwater Monitoring Wells

CONDUCTIVITY AND DISSOLVED SOLIDS

In addition to natural sources, conductivity in groundwater can be impacted due to interactions with surface water. Figure 32 shows the conductivity from the bi-annual monitoring events from WY 2023 along with POR summary statistics. All monitoring well results, with the exception of November MW-1, were higher than the 85th percentile POR value. A MK trend analysis demonstrates that the increasing trend of the annual mean conductivity of all monitoring wells upstream of the Reservoir as well as MW-Kennedy below the Reservoir is significant (Figure 33).

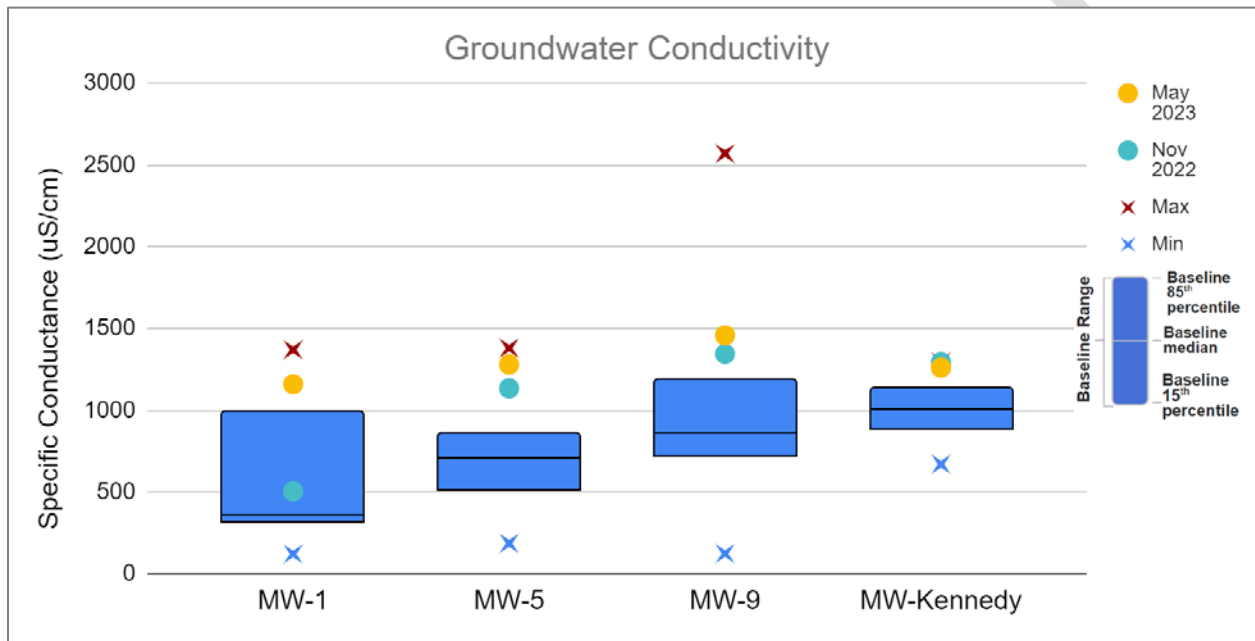


Figure 32. Groundwater Conductivity Summary Statistics and WY 2023 values (Nov 2022 and May 2023).

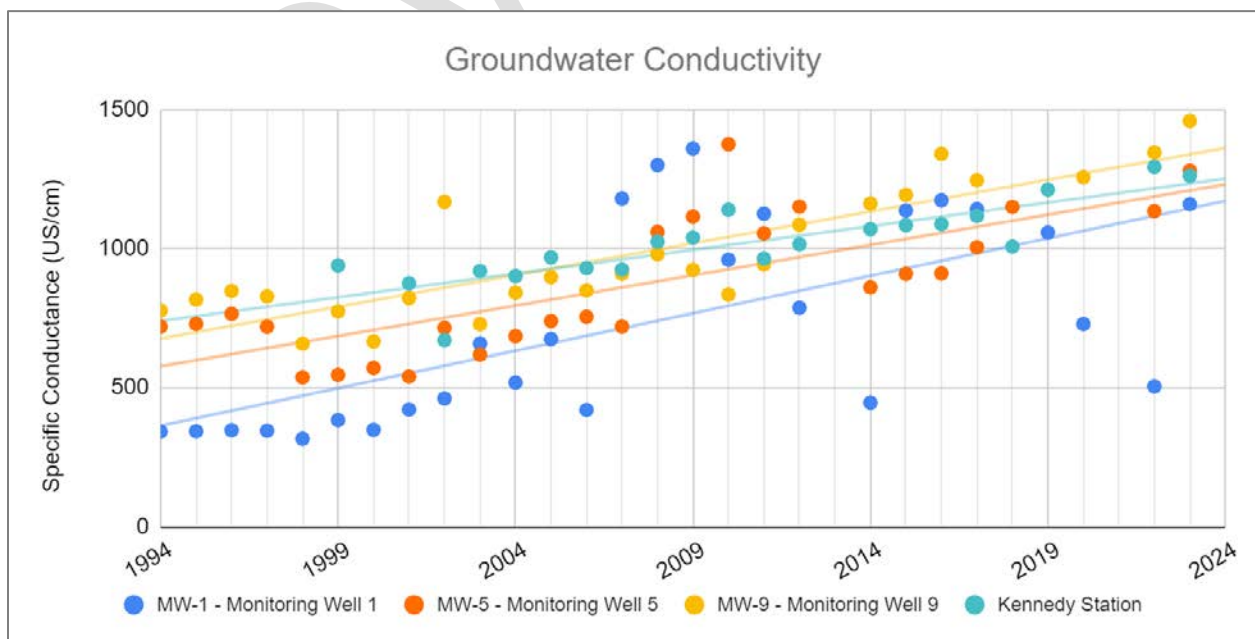


Figure 33. Historical Mean Conductivity in Groundwater Monitoring Wells

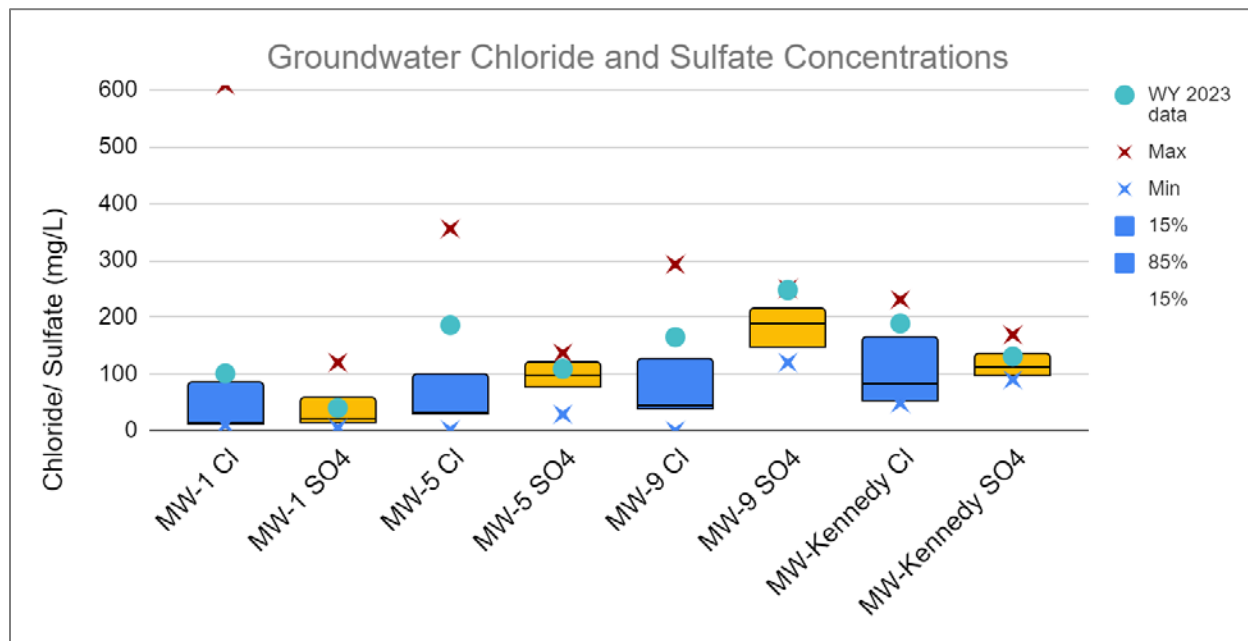


Figure 34. Groundwater Chloride and Sulfate Concentrations

Two of the major dissolved solids components contributing to conductivity are chloride and sulfate. Chloride and sulfate concentrations from the monitoring wells are depicted in Figure 34 with the median from the two monitoring events in WY 2023. The WY 2023 median chloride concentrations were higher than the baseline median and above the 85th percentile for the POR. The WY 2023 median sulfate concentrations were above the baseline median at all sites and MW-9 was above the 85th percentile for the POR. Although these are not drinking water wells, the state water supply standard for both chloride and sulfate is 250 mg/L (5 CCR 1002-41.8). MW-9 approached but did not exceed this value in May 2023 with a concentration of 248mg/L.

PHOSPHORUS

Although total phosphorus is the form evaluated most frequently in surface water, total dissolved phosphorus (TDP) and soluble reactive phosphorus (SRP) concentrations are more useful to compare in groundwater. These forms also have a longer POR and provide more representative concentrations because manual bailing used to sample the wells can increase suspended solids containing particulate phosphorus that skew the results.

Figure 35 shows the median groundwater TDP concentrations and Figure 36 shows the summary statistics for soluble reactive phosphorus in all the monitoring wells that have been monitored historically in addition to the median concentrations from WY 2023 (November 2022 and May 2023). The concentrations of both TDP and SRP were higher in November at all 3 sites upstream but lower at the well (Kennedy) below the Reservoir.

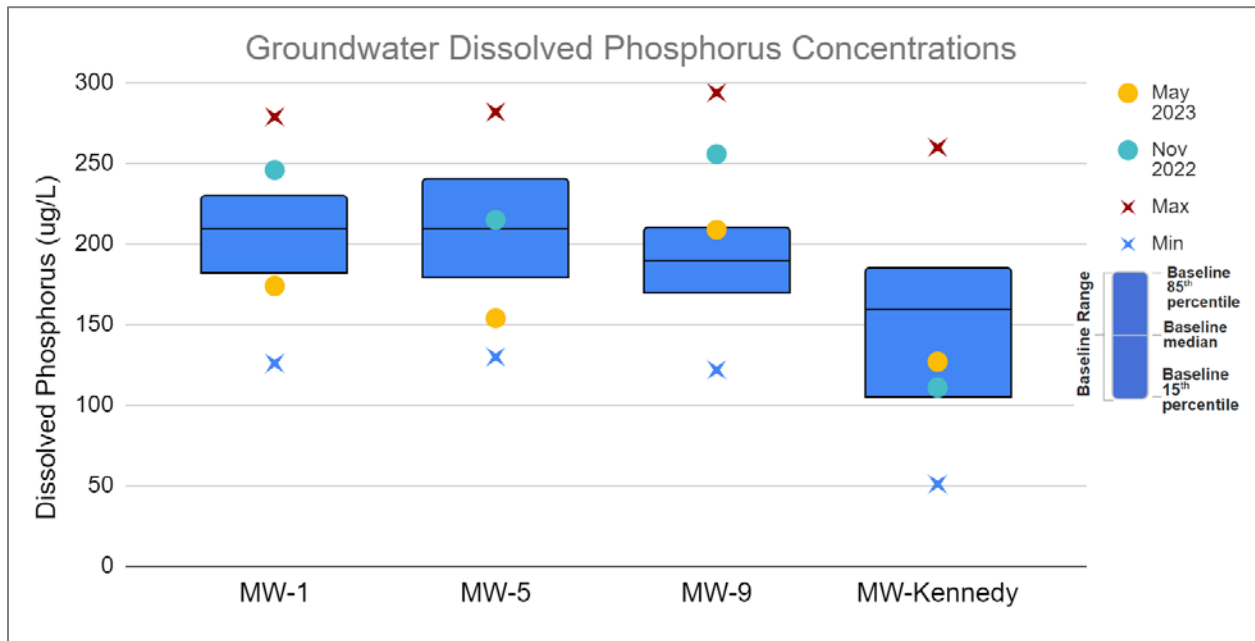


Figure 35. Groundwater Total Dissolved Phosphorus Concentrations, POR Summary Statistics, and WY 2023 (November 2022 and May 2023).

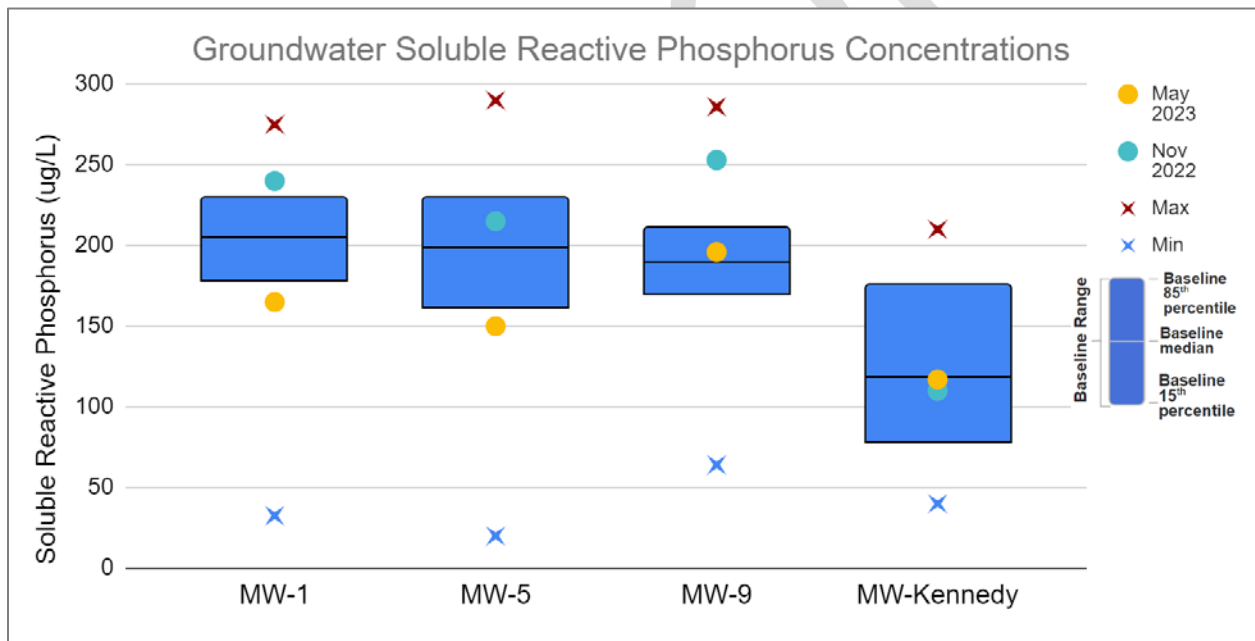


Figure 36. Groundwater Soluble Reactive Phosphorus Concentrations, POR Summary Statistics, and WY 2023 (November 2022 and May 2023).

On average, SRP makes up 86-88% of the TDP concentrations in MW-1 and MW-9 and 95% of the TDP concentration observed in MW-9 just upstream of the Reservoir. Table 16 includes the summary statistics for TDP concentrations for the POR and the median of the WY 2023 values.

Figure 37 depicts the annual mean TDP at the three monitoring wells upstream of the Reservoir. A MK analysis demonstrates that the annual TDP at the monitoring well just above the Reservoir, MW-9, is significantly increasing although there is no significant trend for the wells upstream in the watershed at MW-1 or MW-5.

A MK trend analysis demonstrates that the increasing trend of the annual groundwater TDP concentration above the Reservoir (MW-9) is significant (Figure 37) but the other two wells are not.

Table 16. Dissolved Phosphorus Concentrations ($\mu\text{g/L}$) Summary Statistics (1994-2023) and WY 2023 Median.

Site	Site Abv.	Min	Median	Max	Count	WY2023 median
MW-1 - Monitoring Well 1	MW-1	126	210	279	121	210.0
MW-5 - Monitoring Well 5	MW-5	130	210	282	120	184.5
MW-9 - Monitoring Well 9	MW-9	122	190	294	142	232.5
Kennedy Station	MW-Kennedy	51	160	260	41	119.0

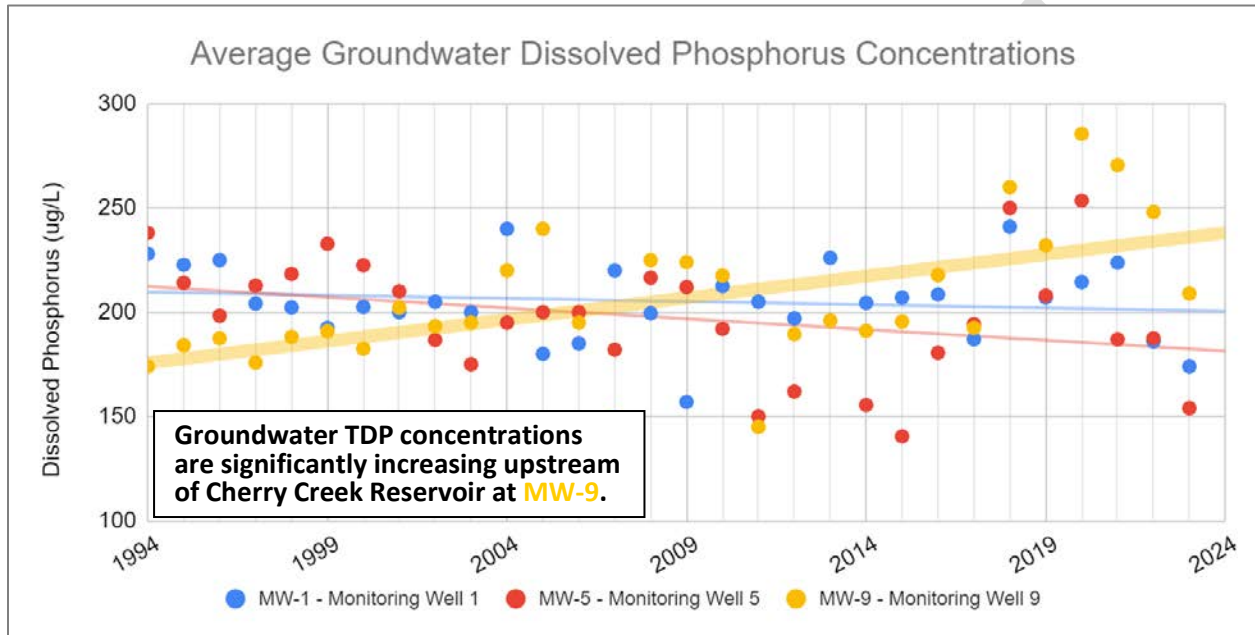


Figure 37. Annual Mean Dissolved Phosphorus in Groundwater Monitoring Wells Upstream of Cherry Creek Reservoir.

NITROGEN

Total Nitrogen (TN) in groundwater has been monitored since 2016 and Nitrate + Nitrite ($\text{NO}_3 + \text{NO}_2 - \text{N}$) since 2013. TN concentration summary statistics for TN in all the monitoring wells that have been monitored historically in addition to the median concentrations from WY 2023 (November 2022 and May 2023) are depicted in Figure 38 and $\text{NO}_3 + \text{NO}_2 - \text{N}$ in Figure 39.

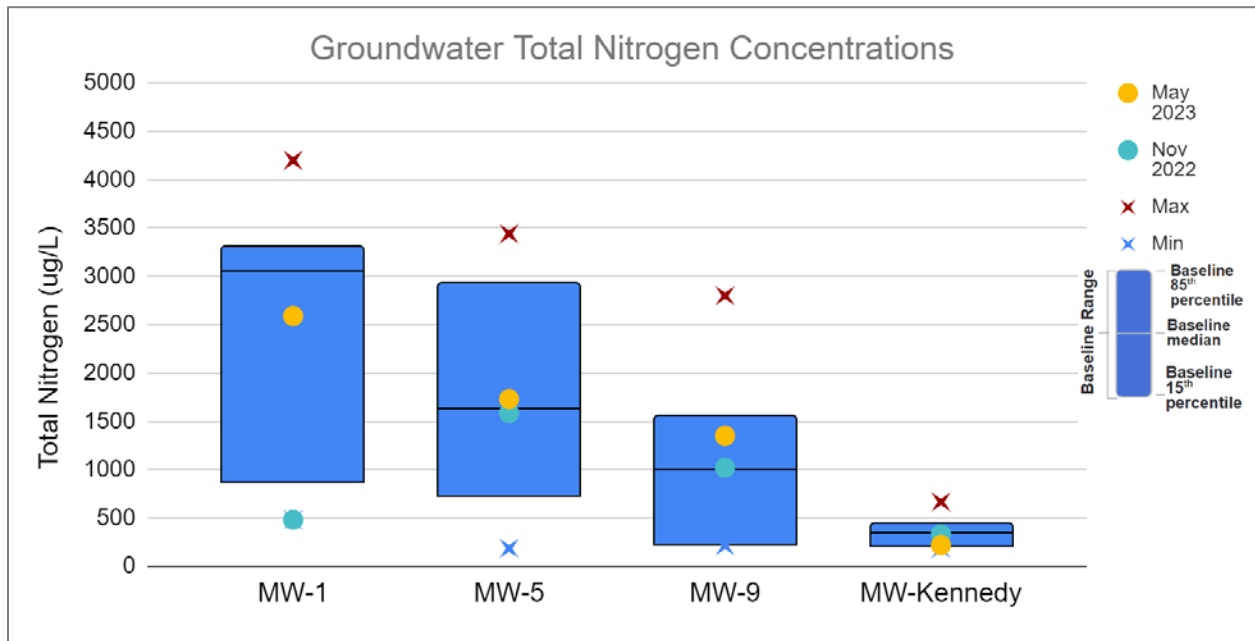


Figure 38. Groundwater Total Nitrogen Concentration Summary Statistics and WY 2023 values (Nov 2022 and May 2023)

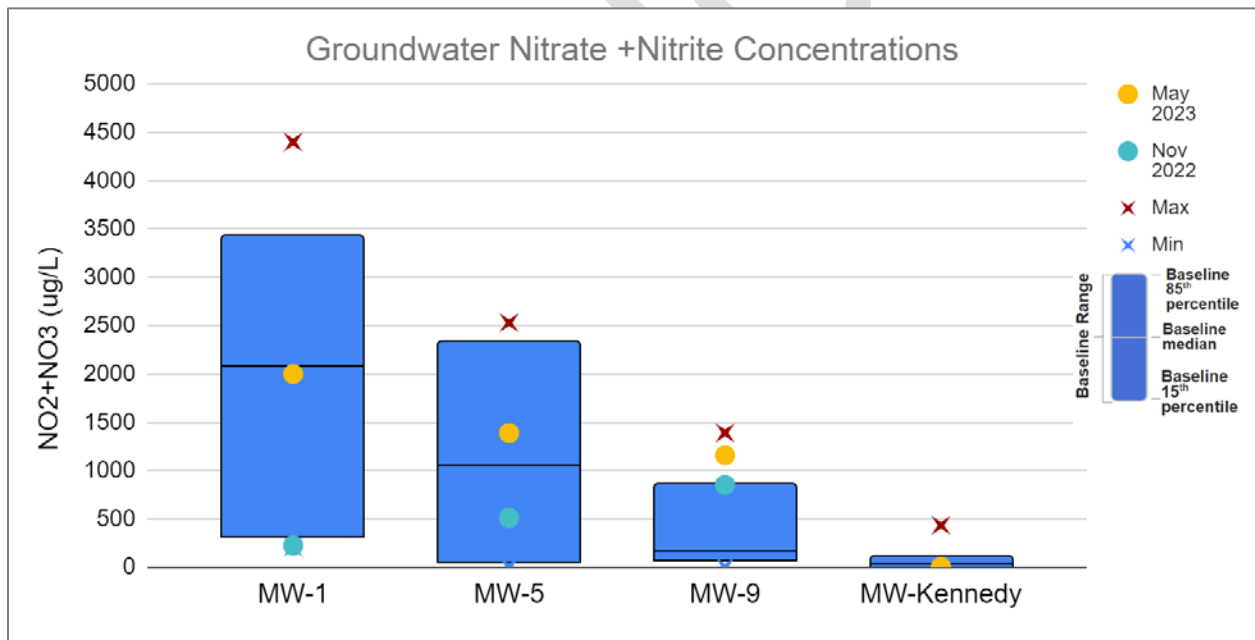


Figure 39. Groundwater Nitrate +Nitrite Concentration Summary Statistics (2013-2023), WY 2023 (November 2022 and May 2023).

The maximum and baseline median TN and NO₂+NO₃ concentrations decrease closer to and below the Reservoir. The concentrations of TN and NO₂+NO₃ were higher in May 2023 at all three sites upstream but TN was lower at the well (Kennedy) below the Reservoir. The WY 2023 concentrations of TN and NO₂+NO₃ were below the baseline median at MW-1, but were above the baseline median at MW-9 just upstream of the Reservoir. Ammonia has also been monitored in groundwater, but due to variability in detection limits, similar analysis provided limited information.

4.0 RESERVOIR MONITORING RESULTS

Reservoir monitoring focuses on data collection to support regulatory requirements and maintaining the beneficial uses of aquatic life, recreation, water supply, and agriculture. The primary concerns are nutrients, including all species of phosphorus and nitrogen, and chl α .

Three sites in the Reservoir are included in the monitoring program: CCR-1, CCR-2, and CCR-3 (Figure 38). CCR-1, also called the “Dam site”, is located in the northwest area within the Reservoir. CCR-2, called the “Swim Beach site”, is located in the northeast area within the Reservoir nearest the swim beach and Reservoir outlet. CCR-3 is referred to as the “Inlet site” and corresponds to the south area within the Reservoir closer to where the streams enter the Reservoir.

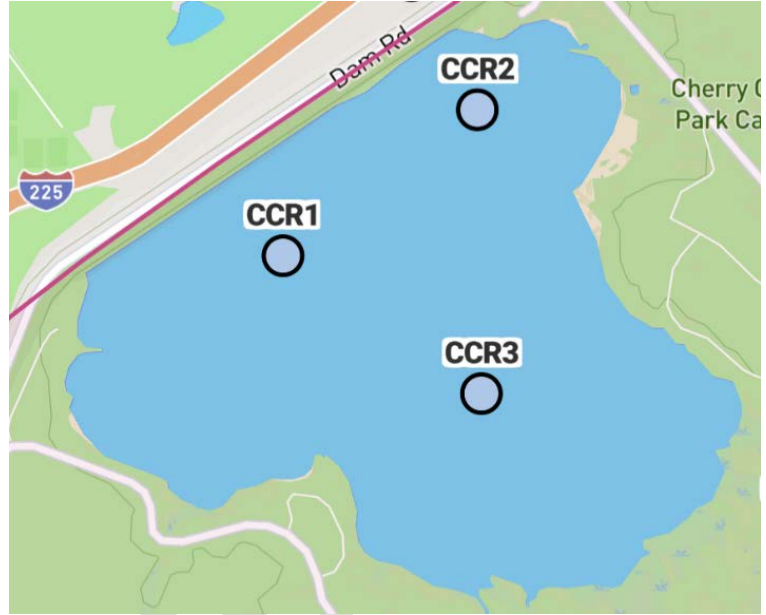


Figure 40. Cherry Creek Reservoir Monitoring Locations

Each site is sampled monthly though the year when ice-free conditions allow, and twice a month from May through September. Transparency, dissolved oxygen, temperature, and pH are included in the regular monitoring to support regulations protecting aquatic life and other beneficial uses.

Water quality samples are collected from the photic zone (0-3 m composite) at each site and from 4 m to the bottom at CCR-2. Physical parameters are measured at 1 m increments from the surface (0 m) to the bottom, which varied from 6.2 to almost 8 m during WY 2023.

In addition to the physical and chemical water quality monitoring, the analysis of reservoir plankton concentrations also helps determine the overall health of Cherry Creek Reservoir, the potential for environmental risks, and impacts on water quality. Plankton growth trends and population diversity through the seasons are analyzed through monthly sample collection throughout the year and twice a month through the summer months. Identification and enumeration are completed on all samples with biovolumes calculated on all phytoplankton samples and biomass calculated on all zooplankton samples.

4.1 USACE RESERVOIR GATE EXERCISE ACTIVITY

The USACE usually completes the annual gate operation activity at the outlet of Cherry Creek Reservoir in late May to verify the proper operation of the outlet gates. The activity was planned for May 24th 2023; however, due to active flood control operations at that time, the gates were operated and maintained at an average release rate of 250 cfs but no additional information was provided. It is assumed that this flushing exercise may release some of the nutrient rich water and sediments from the bottom of the Reservoir.

4.2 TRANSPARENCY

Water transparency, characterized by Secchi depth, is used as an indicator for lake and reservoir water quality because primary productivity (algae) and turbidity of the water column reduce the depth at which light can penetrate. In addition, the photic zone, characterized by 1% Light Transmittance, is a measure of the depth at which light can penetrate the water column and algae can complete photosynthesis. Both Secchi depth and the 99% light attenuation (1% Light Transmission) were measured at all three Reservoir sites during each monitoring event

Figure 41 illustrates the WY 2023 median Secchi depths along with the 1992 to 2023 POR summary statistics for each Reservoir sites. The Secchi depths are similar between the three Reservoir sites, and the WY 2023 median Secchi depth measurements were similar to the baseline medians. The Secchi Depth values in the Reservoir represent low transparency and eutrophic conditions.

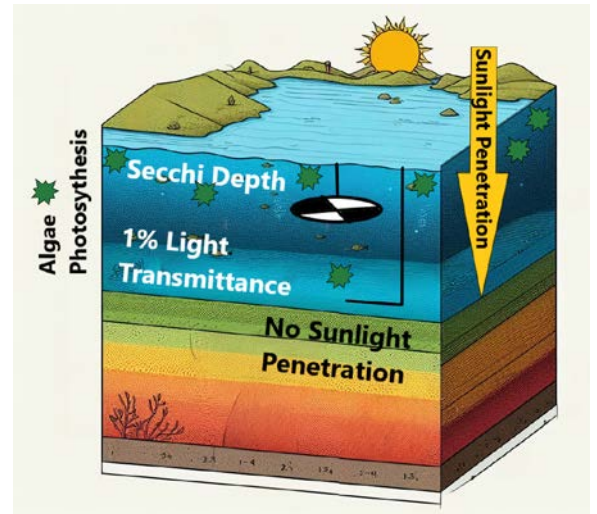


Image 1. Water Transparency - Secchi Depth and Photic Zone

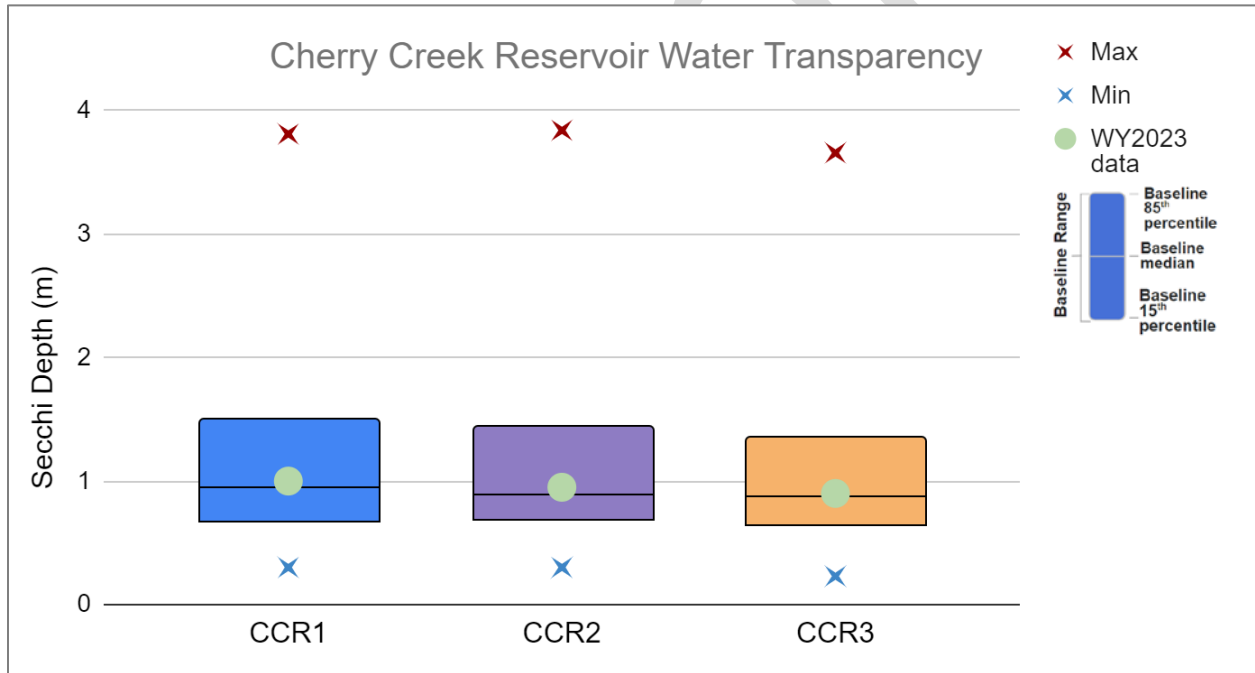


Figure 41. Cherry Creek Reservoir Water Transparency, Secchi Depth Summary Statistics and WY 2023 values.

Figure 42 shows monthly WY 2023 medians along with POR summary statistics. For the most part, the Secchi depth followed a similar seasonal pattern when compared to the historical monthly values. The Secchi depths were highest and above the baseline medians in May, June, and July 2023, which coincided with the period of above average precipitation. Storm events and periods of extended precipitation are responsible for reduced sunlight, increasing inflows to the reservoir, reducing water temperature, and likely assist with mixing, all of which reduce the potential for algae growth and increased water transparency.

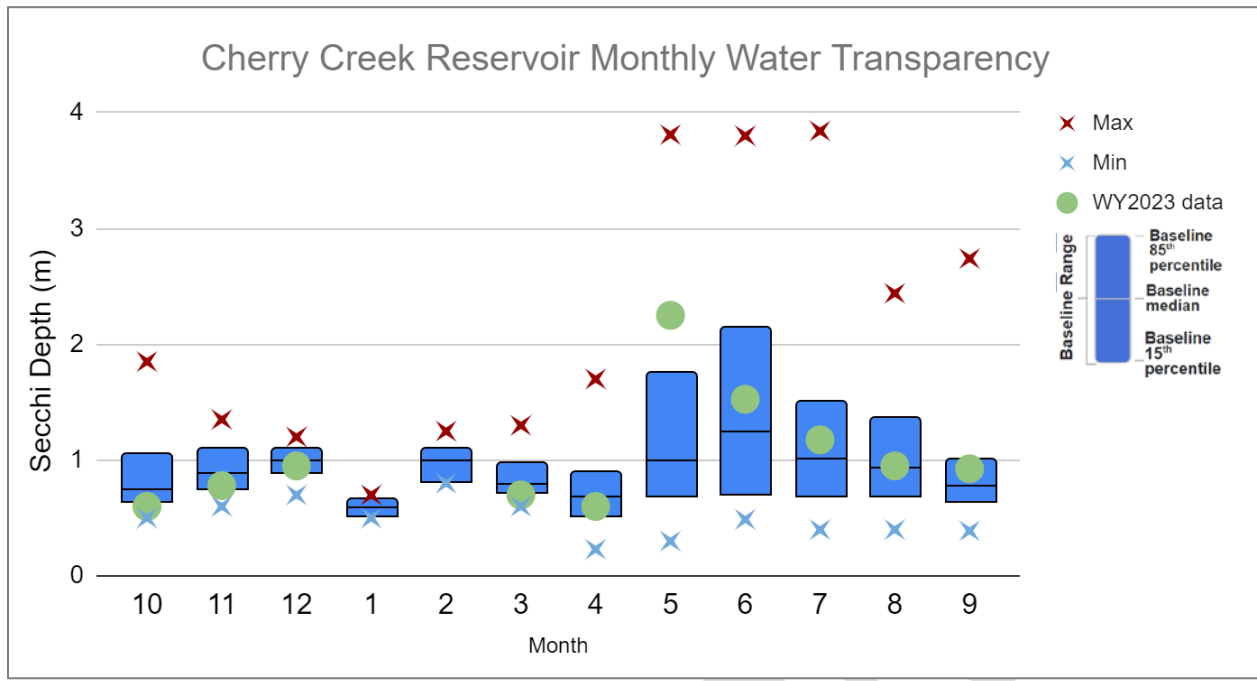


Figure 42. Monthly Median Secchi Depth in Cherry Creek Reservoir from 1992-2022, Summary Statistics and WY 2023 values.

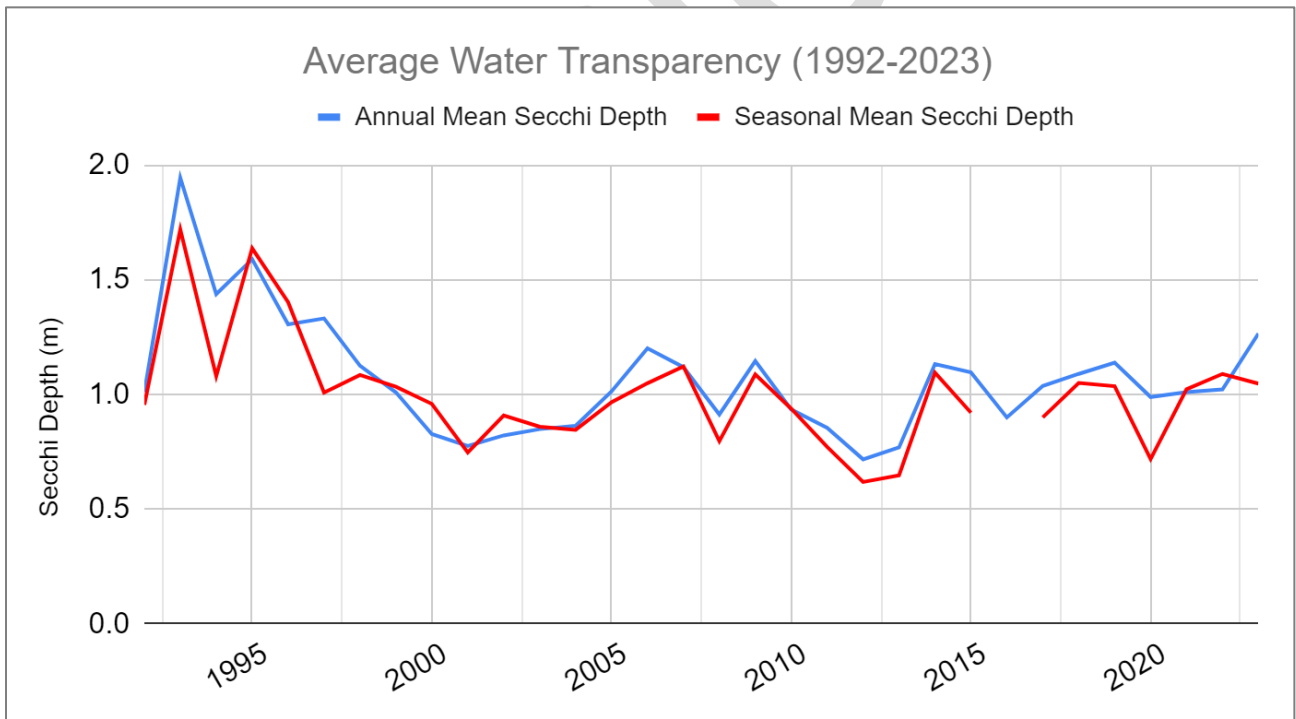


Figure 43. Annual and Seasonal Mean of Secchi Depth in Cherry Creek Reservoir from 1992-2023.

Figure 43 shows the historical annual and seasonal (July through September) mean Secchi depths for Cherry Creek Reservoir. From approximately 1998 to present, the annual mean Secchi depth has been in the eutrophic range, with all annual means less than 2 meters (See Section 4.15). A MK trend analysis indicates that there is no significant increase or decrease over time in either annual or seasonal measurements.

The depth 1% light transmittance is considered the photic zone, or the depth at which photosynthesis can occur; below that depth, primary productivity would be light limited. Like the Secchi depth measurements, the highest measurements of 1% light transmittance were observed in early spring and summer, decreasing through September (Figure 44). There is a clear relationship between the photic zone and water transparency; 1% light transmittance averages around three times the Secchi depth.

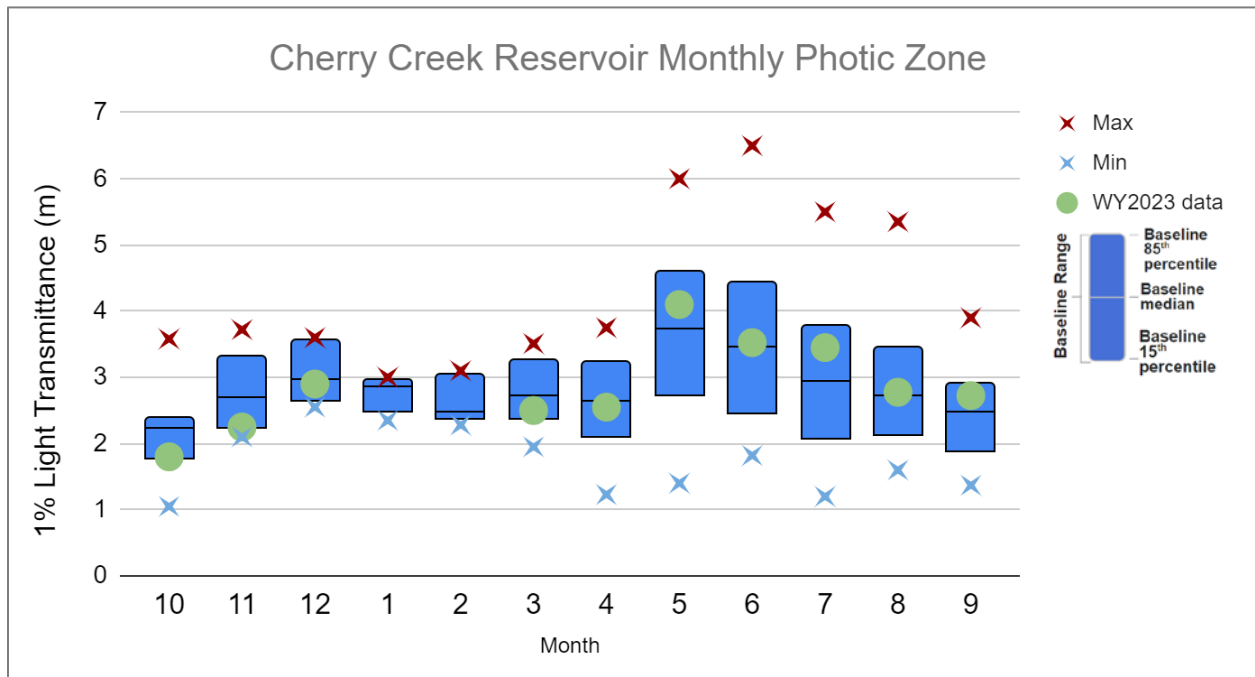


Figure 44. Cherry Creek Reservoir Monthly Photic Zone, Depth of 1% Light Transmittance Summary Statistics and WY 2023 median depths.

4.3 CHLOROPHYLL α

Cherry Creek Reservoir has a seasonal (July through September) chl α standard of 18 $\mu\text{g/L}$ as set by WQCC Reg 38. During each sampling event in WY 2023, chl α levels were measured from composite samples collected from 0, 1, 2, and 3 meters at all three monitoring sites in the Reservoir.

Figure 45 displays the chl α concentration summary statistics for 1992-2023 and the WY 2023 median values. The WY 2023 medians are similar to the baseline medians. Figure 46 illustrates the monthly chl α WY 2023 concentrations along with POR summary statistics. The WY 2023 seasonal chl α mean was 20.9 $\mu\text{g/L}$, which does not meet the Reg 38 standard of 18 $\mu\text{g/L}$ (Figure 47). The standard only allows an exceedance frequency of once in five years; four of the last five (4/5) and eight of the last ten (8/10) years have exceeded this value. The Reservoir is not meeting the chl α water quality standard.

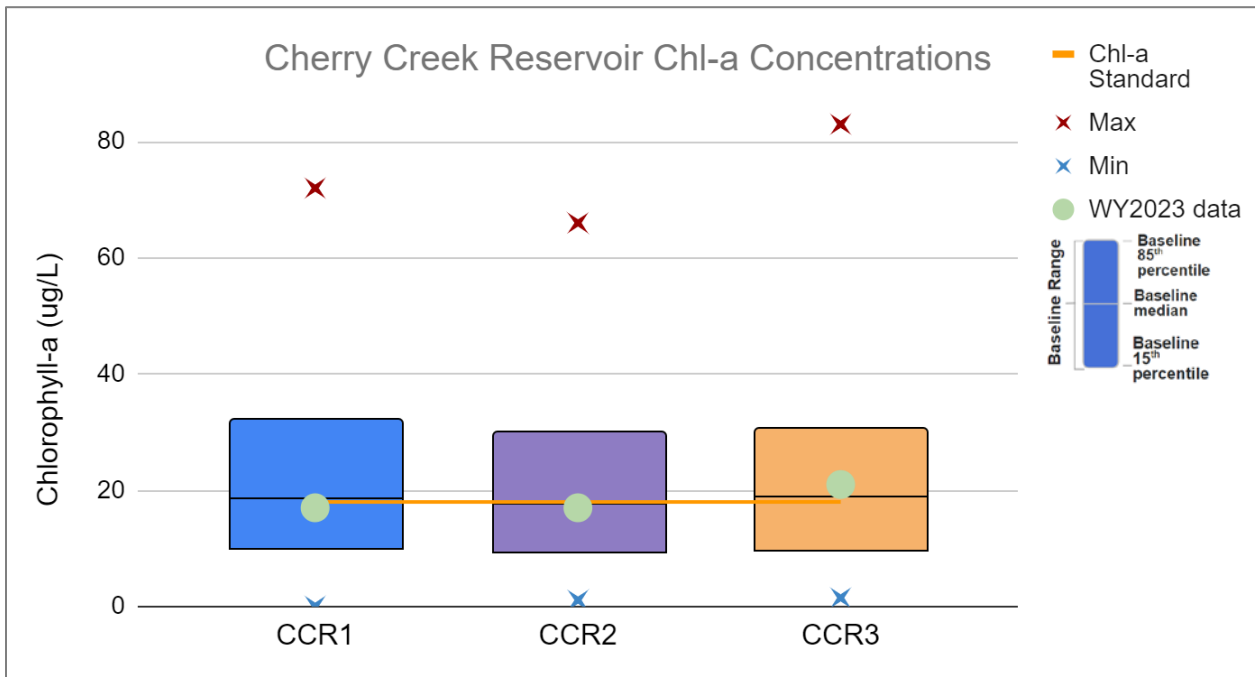


Figure 45. Cherry Creek Reservoir Chlorophyll α Concentrations, POR Summary Statistics and WY 2023 data.

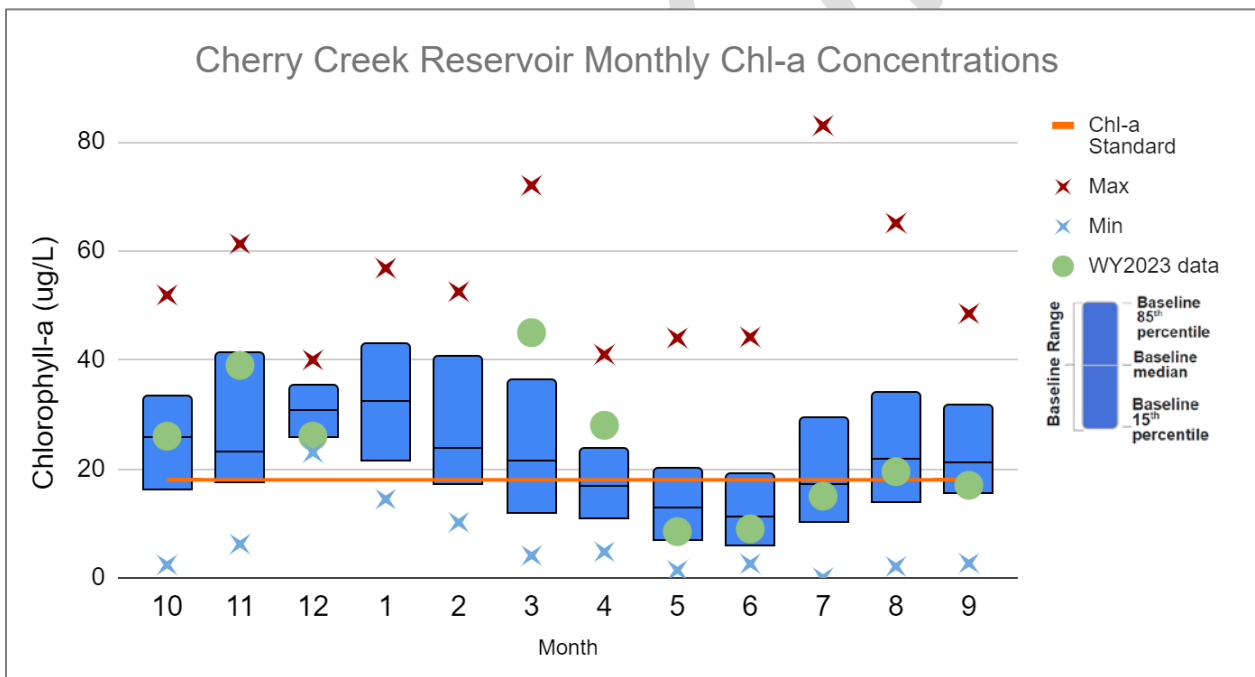


Figure 46. Monthly Median Chlorophyll α Concentrations in Cherry Creek Reservoir from 1992-2022, Summary Statistics and WY 2023 values.

The highest WY 2023 monthly median chl α concentrations were collected during the monitoring events in November and March and the lowest in May, June, and July, even though there was a bloom in late July. The low chl α values coincided with the highest water transparency in Cherry Creek Reservoir. However, as soon as the weather started to warm and the heavy precipitation from spring and early summer stopped, algae concentrations increased.

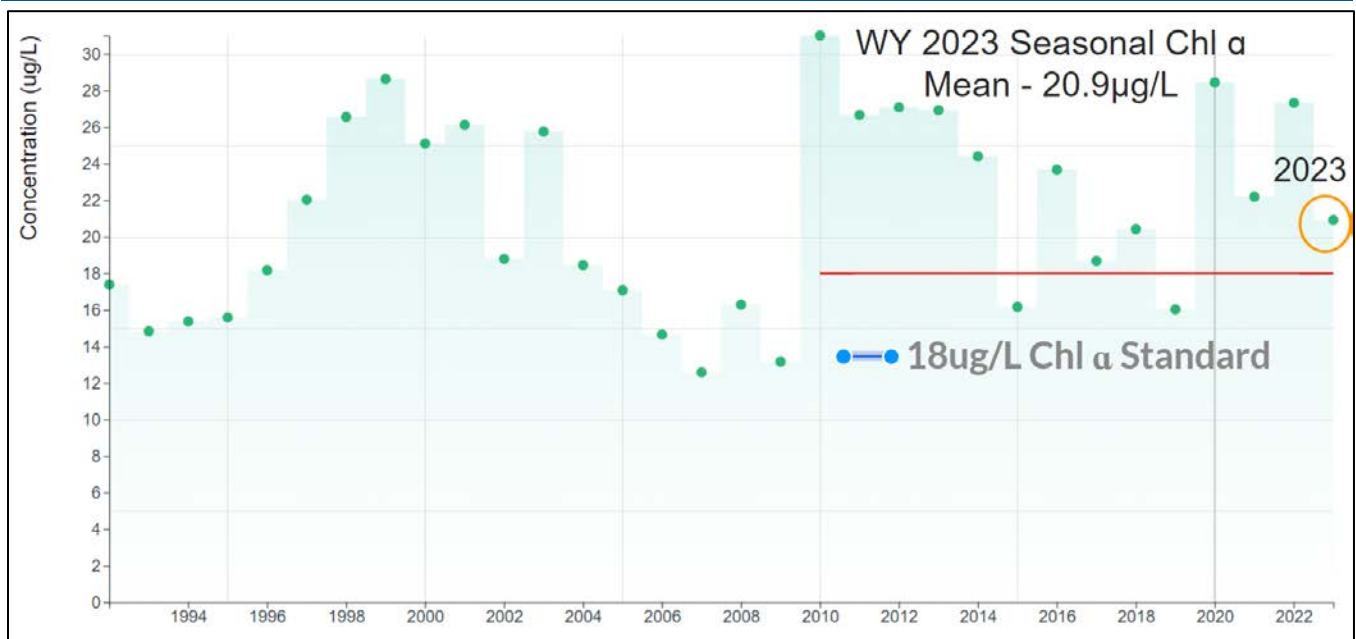


Figure 47. Seasonal Mean Chlorophyll *a* in Cherry Creek Reservoir WY 1991-2023.

Translating the impacts of chl *a* concentrations on water quality into terms that are meaningful to most recreational lake users is a complex task. Walmsley and Butty (1979) proposed some typical relationships between maximum chl *a* concentrations and observed impacts (Table 17) to describe perceptions of water quality by typical lake users.

Table 17. Impact of Chlorophyll *a* Concentrations on Perceived Water Quality

Chlorophyll <i>a</i> Concentration	Nuisance Value
0 to 10 µg/L	No problems evident
10 to 20 µg/L	Some algal scums evident
20 to 30 µg/L	Nuisance conditions encountered
Greater than 30 µg/L	Severe nuisance conditions encountered

The chl *a* concentrations in Cherry Creek Reservoir indicate that some algal scums to severe nuisance conditions are present throughout the year (Figure 46). When algal scums are evident, Colorado Parks and Wildlife monitors and tests for potential cyanobacteria toxins at multiple public areas.

On July 17th a cyanobacteria bloom was observed in the marina along the shoreline but had very low concentrations of toxin (0.5 µg/L), well below the recreational threshold for closure of 8 µg/L. “Caution” signs were posted in the area to inform the public. Ongoing monitoring detected that the toxin increased to >10 µg/L and a closure was implemented on July 28th in the vicinity of the bloom and “Danger” signs were posted. On July 31st the toxin levels had decreased to below the recreational threshold and the closure was lifted on Aug 4th following the results from laboratory analysis. By August 15th the bloom had dissipated, and no toxin was present.

The pattern of short duration cyanobacteria blooms is common when they are present in Cherry Creek Reservoir. There are many factors that drive and disrupt the blooms. Informing the public with appropriate signage in impacted areas is helpful to reduce risks associated with toxin.

4.4 TEMPERATURE

The Warm Water Aquatic Life classification for Cherry Creek Reservoir in Reg 38 has a chronic Maximum Weekly Average Temperature (MWAT) standard of 26.2°C (79.2°F) and an acute Daily Maximum (DM) standard of 29.3°C (84.6 °F). Both of these standards were met in Cherry Creek Reservoir in WY 2023.

Continuous temperature monitoring is completed annually near site CCR-2 in Cherry Creek Reservoir. The temperature loggers are placed in even increments from one (1) meter of depth to the bottom of the Reservoir and are mounted on a marker buoy. However, after removal of the thermistor chain from the reservoir in the fall of 2023, the chain and equipment could not be located so this data is not available for WY 2023.

During each monitoring event, temperature profiles were also collected during each monitoring event. Figure 46 illustrates the temperature profiles collected at Reservoir station CCR-2 during the routine monitoring events in WY 2023.

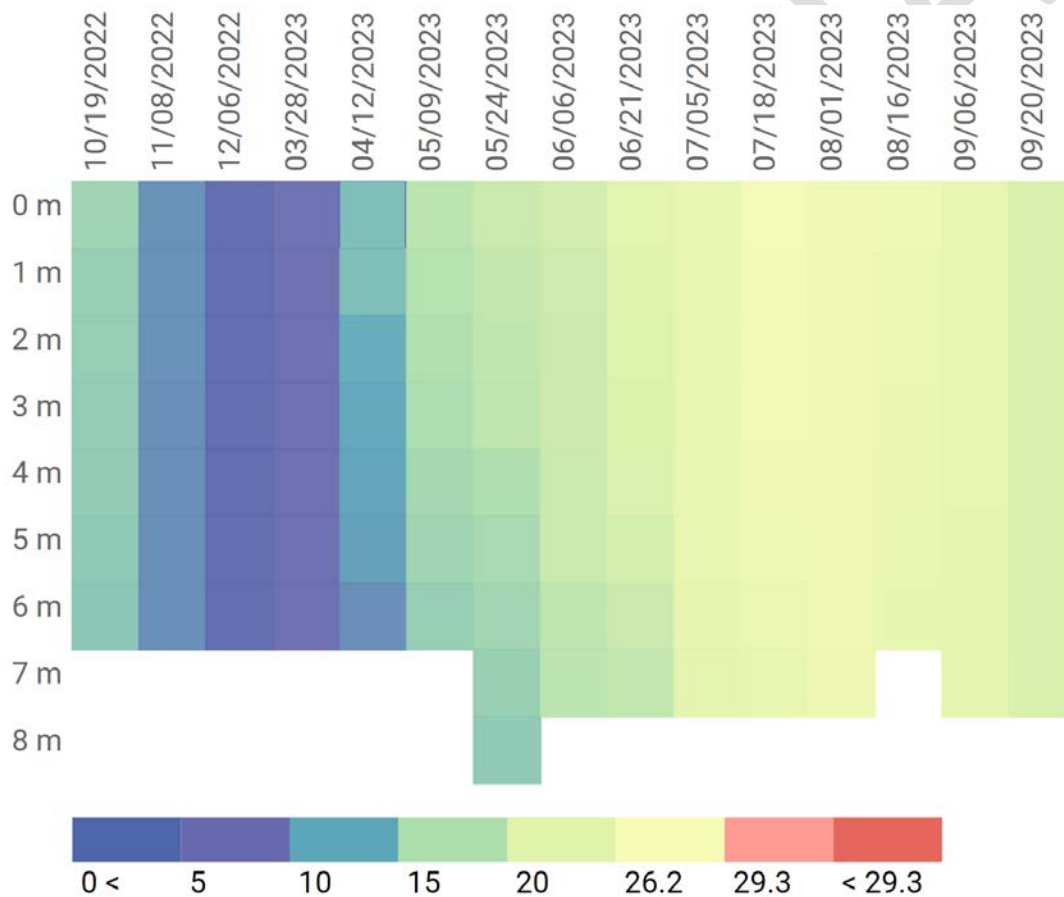


Figure 48. Temperature (°C) Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

The maximum temperature measured at the surface during the Reservoir monitoring events was 26.2 °C (79.16°F) at CCR-3 on July 18, 2023. On that same date, the temperature was 25.3 °C at CCR-2 and 24.9 °C at CCR-1. Cherry Creek Reservoir did not exceed the MWAT or DM standards in WY 2023 and therefore was in attainment. The biggest temperature range measured in the vertical profiles during the monitoring events was 4.5°C on July 18, 2023 (Figure 48).

Although Cherry Creek Reservoir has a destratification system, some of the characteristics of seasonal and mid-season turnover, or mixing events, still occur. However, it is difficult to determine the main turnover events since the Reservoir appears to be polymictic, or able to mix multiple times a season. There was some variability in temperature from the surface to the bottom, which was much more apparent during the warmer summer months of July and August, but during the rest of the year thermal stratification was limited in the Reservoir.

4.5 DISSOLVED OXYGEN

Reg 38 assigns a minimum chronic dissolved oxygen standard of 5.0 mg/L to the Reservoir. The standard requires DO to be at least 5.0 mg/L in the upper portion of a lake or reservoir and that if DO is below 5.0 mg/L, adequate refuge for aquatic life (with DO above 5.0 mg/L) needs to be available at other depths or locations in the Reservoir during the same time period. DO concentrations are measured at 1 m depth intervals throughout the water column during each monitoring event at each site. Cherry Creek Reservoir met the DO standard in WY 2023.

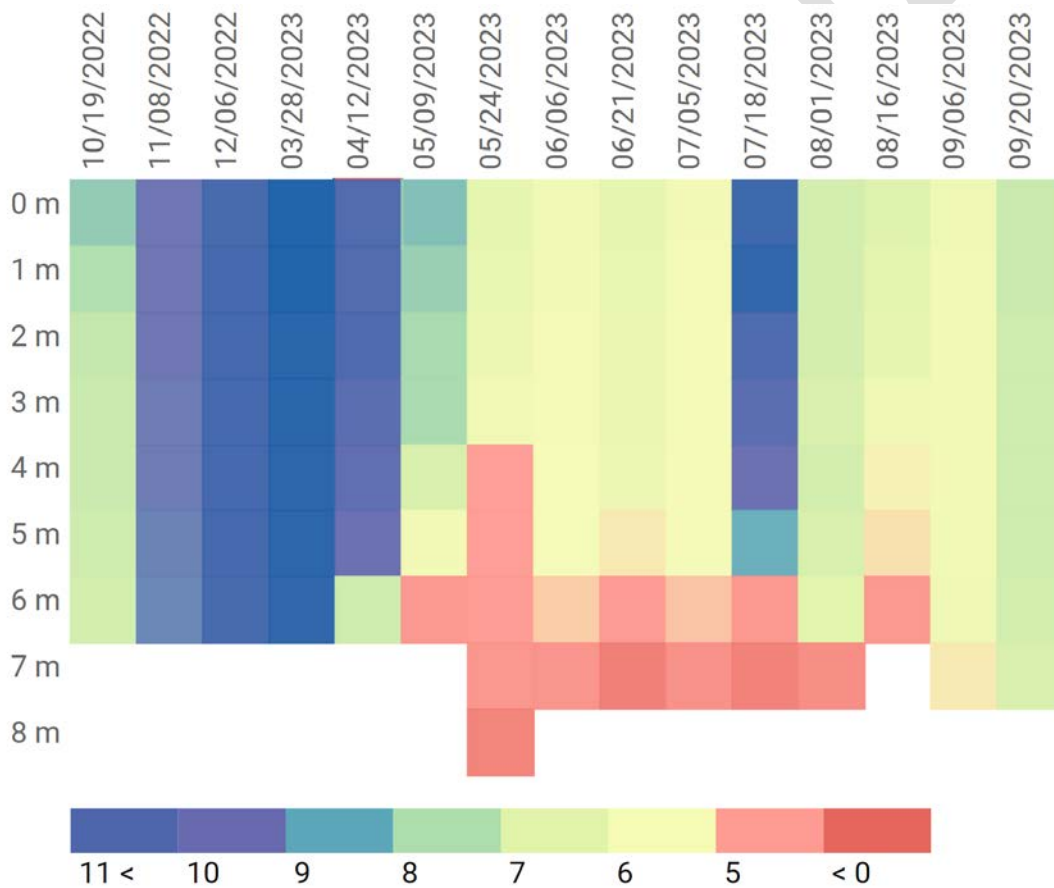


Figure 49. Dissolved Oxygen (mg/L) Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

Figure 49 illustrates the DO concentrations from the surface (0 m) to the bottom in the Reservoir at station CCR-2 during WY 2023. The profiles from the other two sites (CCR-1 and CCR-3) are available on the data portal. DO concentrations below 5.0 mg/L at or near the bottom of the reservoir during the warm summer months are likely due to high microbial activity or decomposition in the hypolimnion and sediments that reduce DO concentrations. During these periods of low DO in the bottom of the Reservoir internal loading of phosphorus from the sediments is likely. The internal loading patterns are affected by the stratification of the water column.

The epilimnion of a lake or reservoir is the mixed layer near the surface. This is the layer in which most photosynthesis occurs because of its higher relative temperature and sunlight penetration. Aquatic macrophytes or rooted plants grow in the littoral (near shore) zone, but most phytoplankton exist in the epilimnion layer. The hypolimnion, or bottom layer, is cooler and denser than the layers above. This layer is where suspended materials, dead algae and other aquatic organisms and plants settle to the bottom to decompose. During the decomposition process, bacterial oxygen consumption exceeds the concentrations in the water, so the DO levels decline. These anoxic conditions at the bottom of the Reservoir in the hypolimnion lead to internal loading of phosphorus from the sediments (Figure 50). When the reservoir mixes, either seasonally or due to high inflows or wind, these high phosphorus concentrations reach the epilimnion where warmer conditions and sunlight penetration drives algae growth.

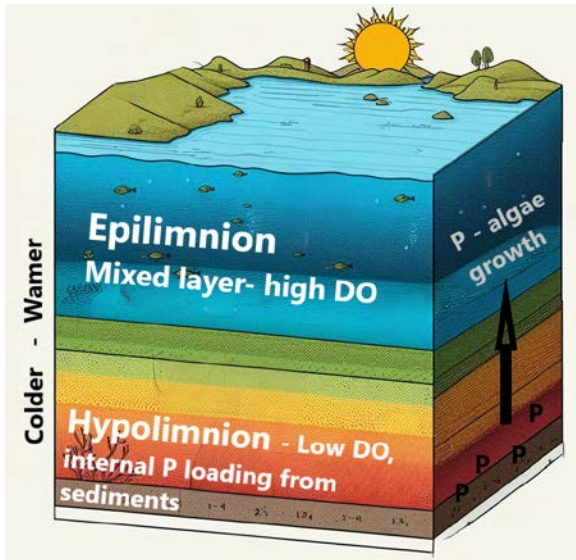


Figure 50. Stratification Layers and Internal Loading

The reservoir destratification system (RDS) at Cherry Creek Reservoir, which pumps air to the bottom of the reservoir through diffusers, helps to mix the water column and is most effective in the spring and fall when there is less thermal stratification.

4.6 pH

Reg 38 assigns a pH standard for Cherry Creek Reservoir based on the acceptable pH Range of 6.5 to 9.0 for protection of aquatic life.

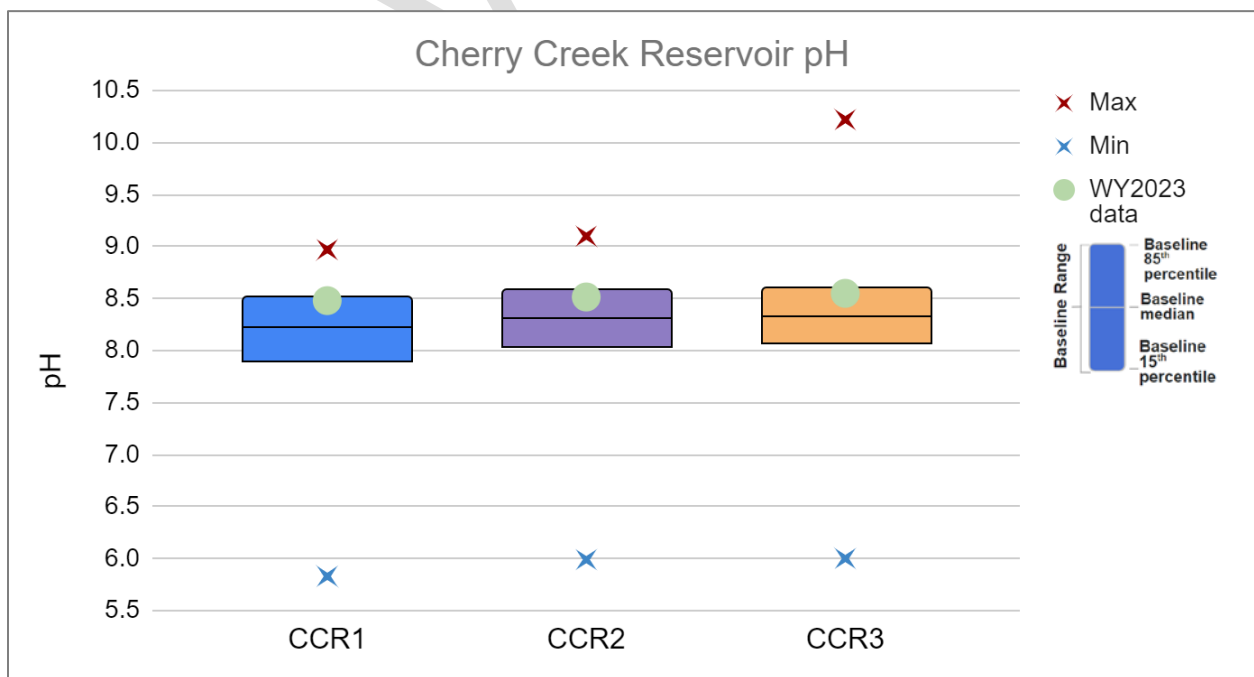


Figure 51. Cherry Creek Reservoir pH, Summary Statistics and WY 2023 medians.

Assessment of pH data is based on comparison of the 15th percentile of the data to a lower pH limit of 6.5 and comparison of the 85th percentile of the data to an upper pH limit of 9.0. Cherry Creek Reservoir attained the pH standard in WY 2023 although median values were above the baseline medians at each site (Figure 51).

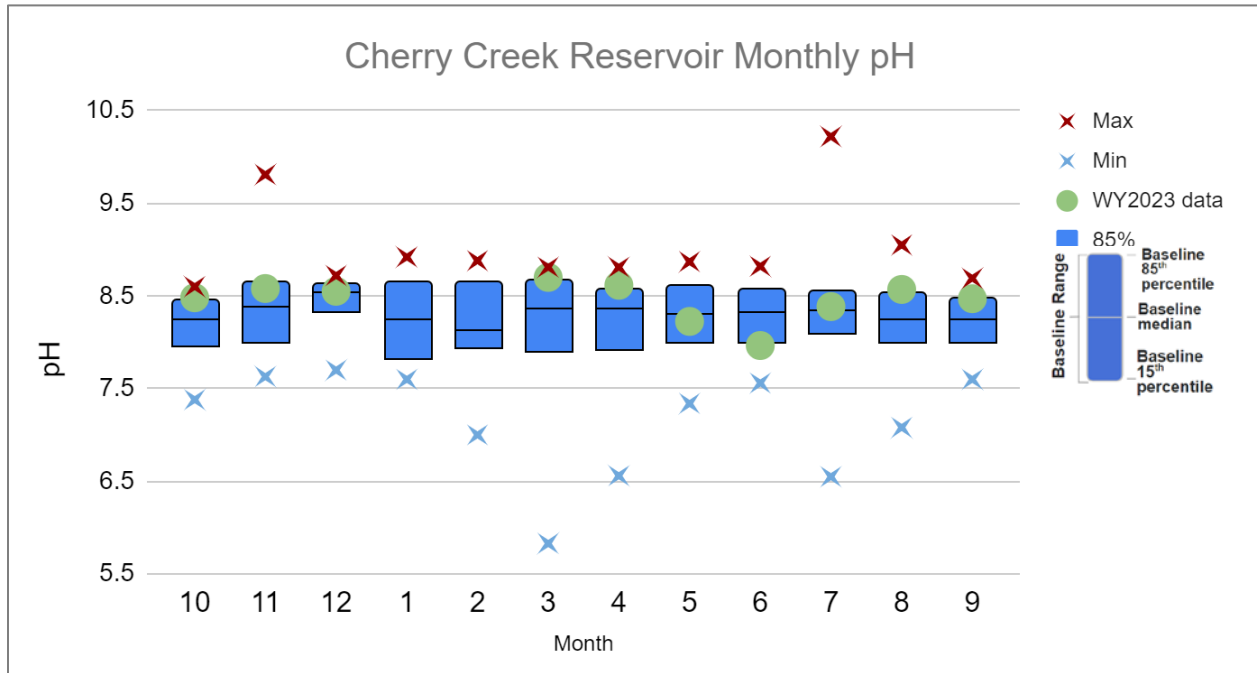


Figure 52. Cherry Creek Reservoir Monthly Median pH.

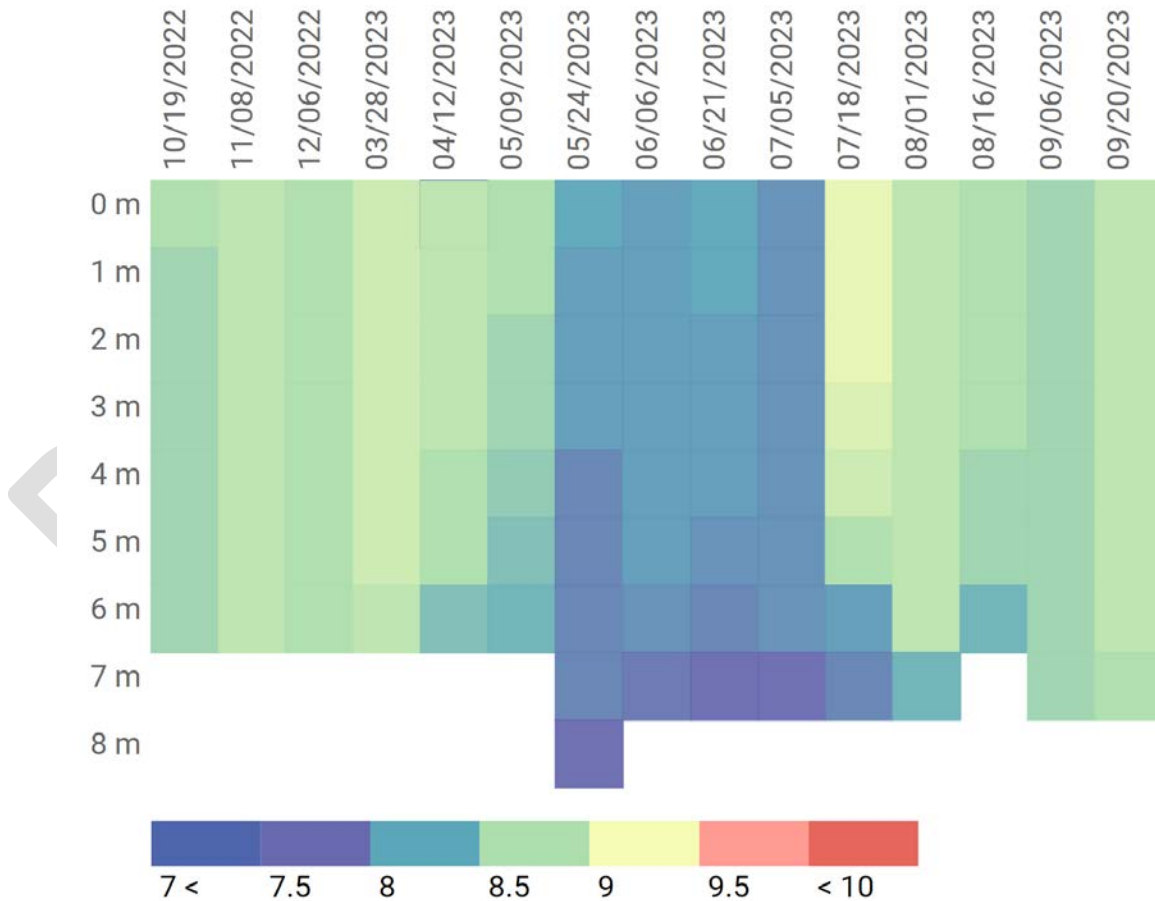


Figure 53. pH Depth Profile from CCR-2, Cherry Creek Reservoir, WY 2023.

The monthly median pH in WY 2023 was below the baseline median in May, June, and slightly below in July (Figure 52) but was near or above the baseline median in all other months. Figure 53 illustrates the pH depth profile for CCR-2. Profiles for the other two Reservoir sites are available on the data portal. The lowest pH values were recorded during the period of high precipitation in May through early July, but as algal productivity increased the pH values observed were also higher but never exceeded 9.0. Lower pH values were present at or near the bottom of the Reservoir which is typical.

Higher pH values are usually correlated with higher productivity and elevated chl α concentrations in the Reservoir. This occurs because photosynthesis removes carbon dioxide, a weak acid, from the water column. For example, the highest chl α concentration measured in WY 2023 was 38 $\mu\text{g/L}$ on July 18th, which coincided with the pH of 8.9 on the same date.

4.7 OXIDATION REDUCTION POTENTIAL

Figure 54 shows the Oxidation Reduction Potential (ORP) WY 2023 monitoring values from CCR-2. Higher ORP values indicate an oxidative state and increased potential to break down organic material, whereas low and negative values indicate a reducing environment.

During WY 2023, the ORP in the photic zone was lowest on July 18th, 2023, when there was a bloom present in the Reservoir. In late September ORP values were low through the water column. Lower ORP values indicate a reducing environment at the bottom of the Reservoir, which usually coincides with lower DO and lower pH measurements. These lower values are an indication of decomposition processes in the sediments and the sediment-water interface, as well as seasonal trends normally seen in the Reservoir. Higher ORP values, indicating an oxidizing environment, were present during periods with higher DO levels and colder water temperatures.

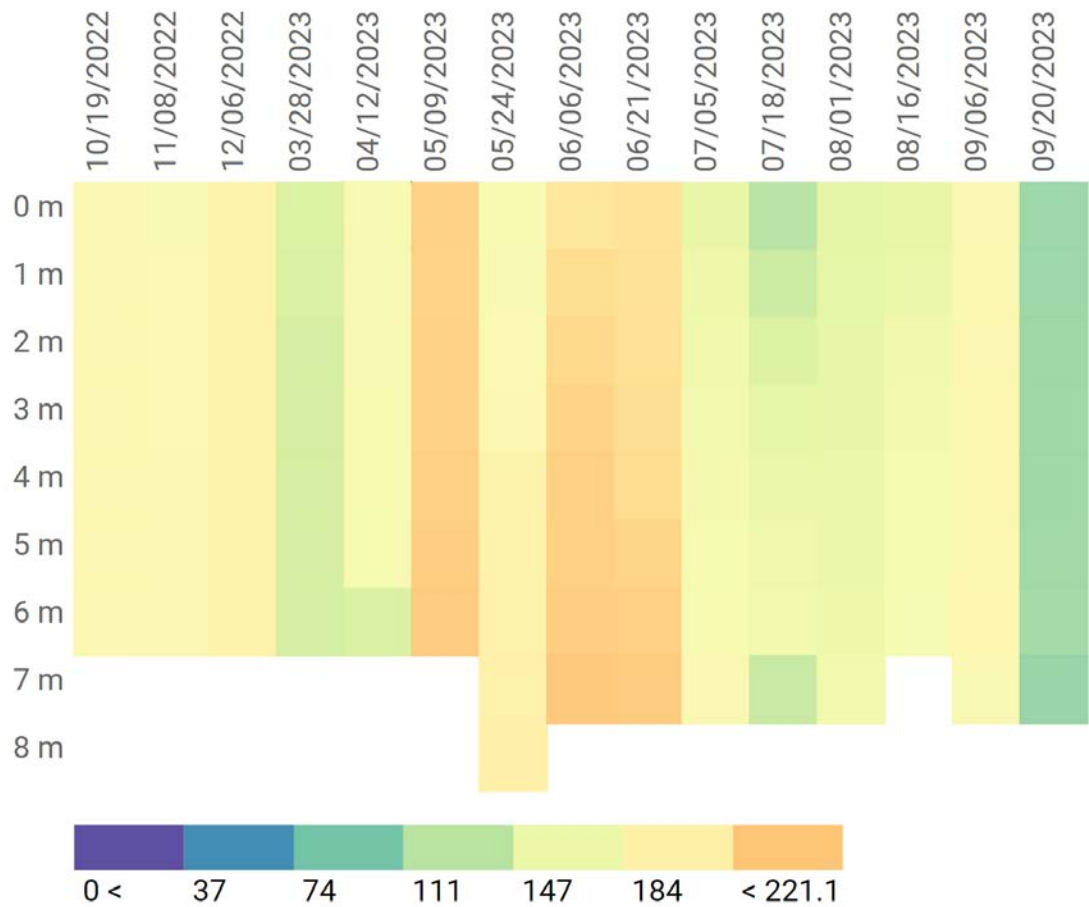


Figure 54. Oxidation Reduction Potential (mV) Depth Profile, CCR-2, Cherry Creek Reservoir, WY 2023.

4.8 CONDUCTIVITY

The specific conductance, or conductivity, is a representation of dissolved solids (e.g., salts, minerals) in Cherry Creek Reservoir. Although there is no water quality standard for conductivity, US EPA considers levels above 1,500 $\mu\text{S}/\text{cm}$ above average for most streams in the US. Figure 55 shows the annual median specific conductance WY 2023 values along with the POR statistics for the Reservoir monitoring sites compared to the EPA benchmark. Reservoir WY 2023 median conductivity values were similar to baseline values and below EPA benchmarks. Figure 56 illustrates monthly conductivity in the Reservoir. During WY 2023, the conductivity was above the baseline median until May, during the period of above average precipitation, and then increased slowly through September. (Although conductivity differed throughout the year, there was limited variability observed from the top to bottom of the Reservoir and between the three monitoring sites (Figure 57).

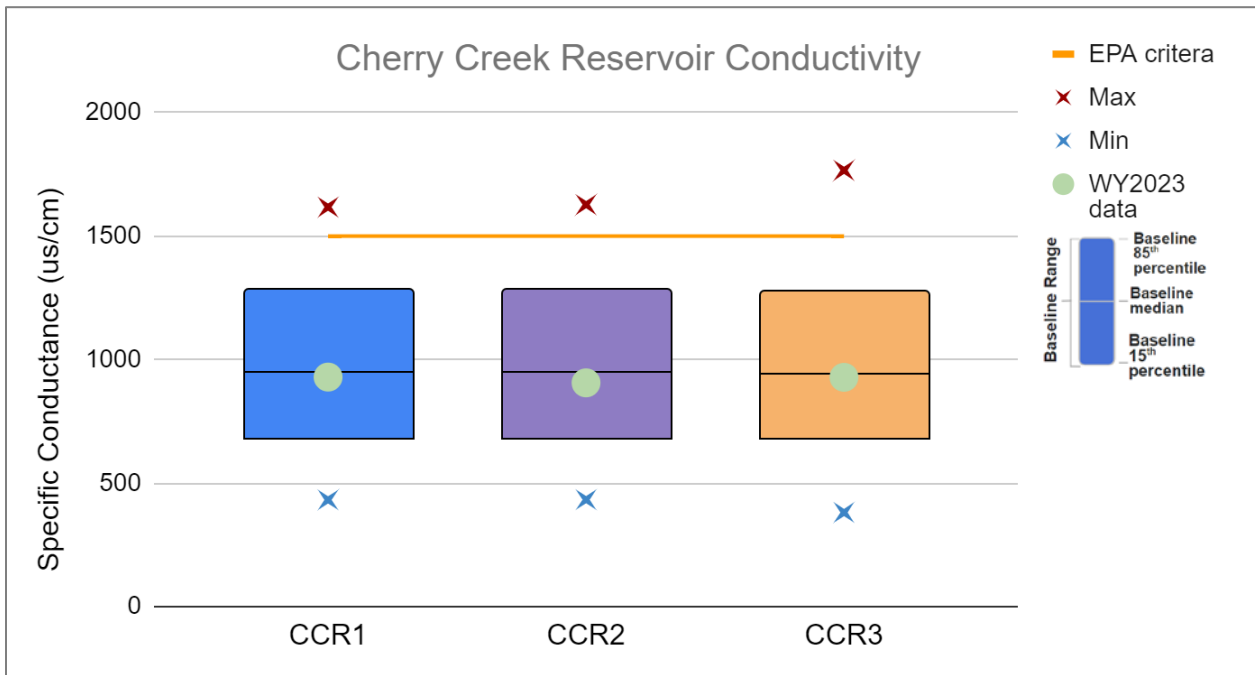


Figure 55. Cherry Creek Reservoir Conductivity, Summary Statistics (1999-2023), WY 2023 medians.

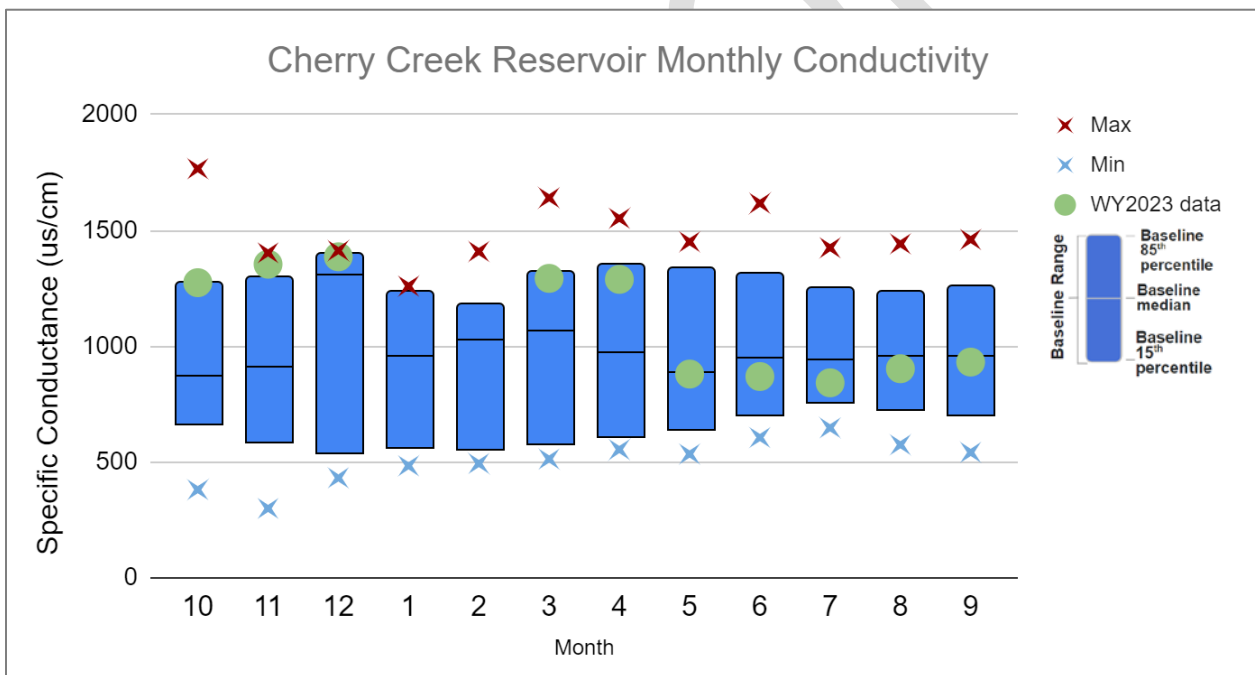


Figure 56. Monthly Conductivity in Cherry Creek Reservoir, Summary Statistics and WY 2023 medians.

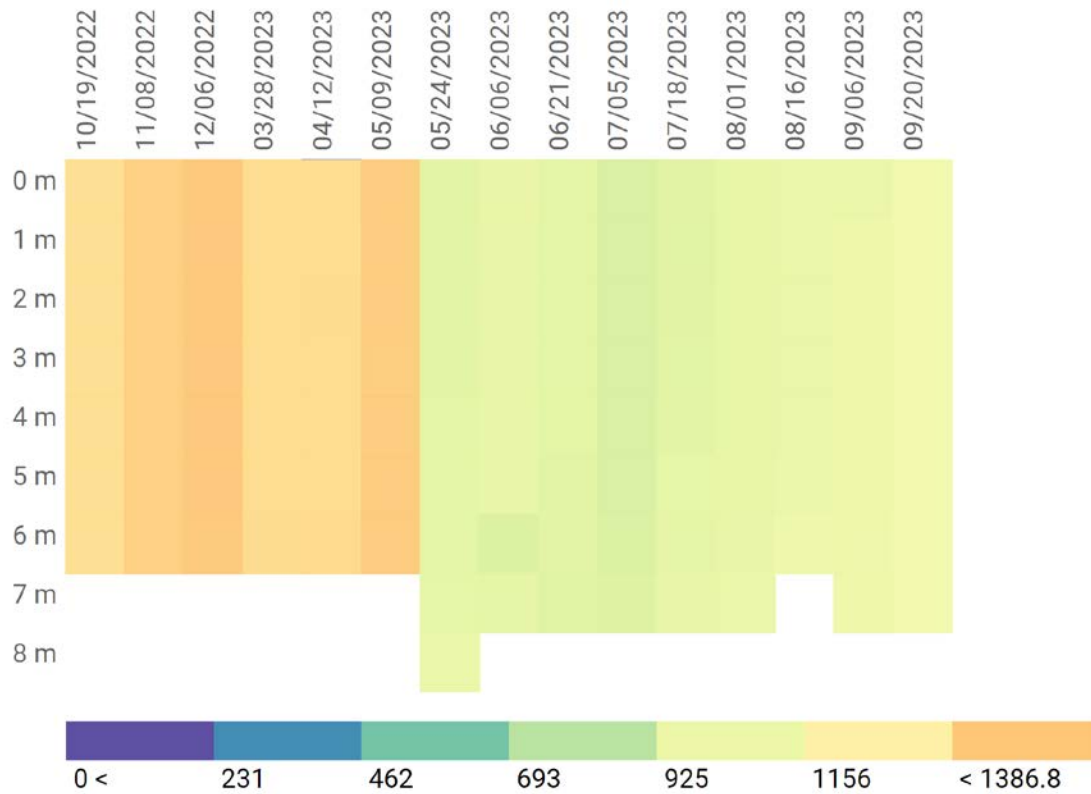


Figure 57. Conductivity (Specific Conductance $\mu\text{S/cm}$) Depth Profile, Cherry Creek Reservoir, CCR-2, WY 2023.

4.9 SUSPENDED SOLIDS

Total suspended solids (TSS) in a lake or reservoir represent all particles greater than $2\ \mu\text{m}$ in the water column such as sand silt, clay, and algae. The TSS concentrations in Cherry Creek Reservoir impact water clarity and can indirectly affect chl α concentrations due to changes in depth of sunlight penetration.

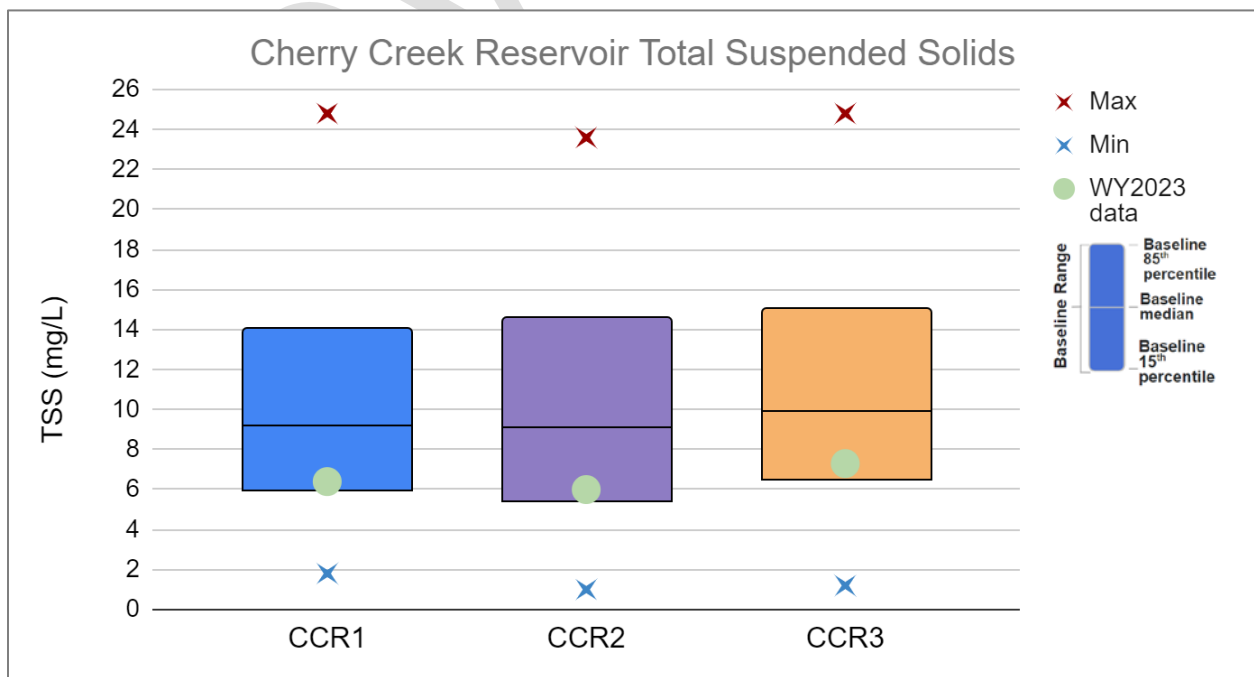


Figure 58. Total Suspended Solid Concentrations in Photic Zone, Cherry Creek Reservoir, Summary Statistics (1992-2023) and WY 2023 medians.

Although stormflows often have high TSS concentrations which can impact downstream lakes and reservoirs, the median concentrations in WY 2023 were below the baseline median (Figure 58). In addition, the monthly medians following the high spring inflows were lower than the baseline medians and below the 15th percentile in May and June (Figure 59.).

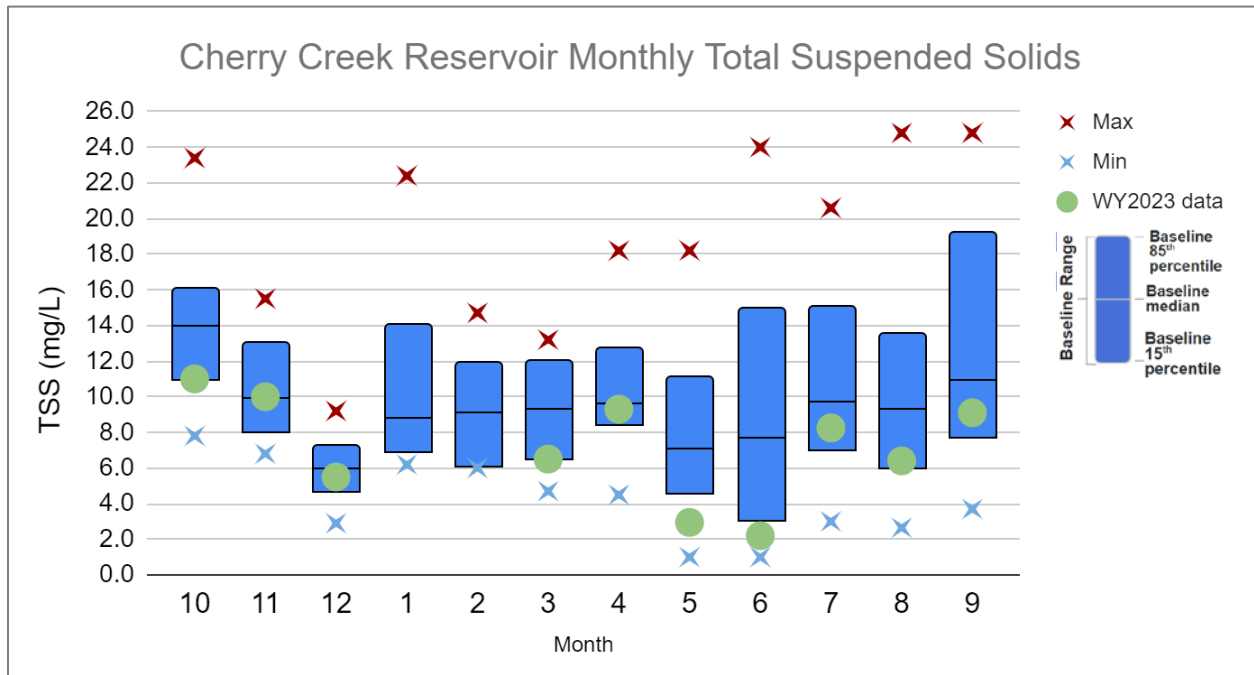


Figure 59. Monthly Total Suspended Solids in Cherry Creek Reservoir, Summary Statistics and WY 2023 medians.

4.10 TOTAL PHOSPHORUS

In many aquatic environments phosphorus limits primary productivity or algal growth, but in eutrophic or nutrient rich environments, like Cherry Creek Reservoir, phosphorus may not be limiting. Total phosphorus (TP) is made up of both particulate and dissolved phosphorus. Particulate phosphorus is what remains suspended in the water column instead of settling to the bottom of a lake or reservoir. It includes both inorganic material, such as soil particles and clay minerals, and organic phosphorus, which includes particulate forms such as algal cells and plant fragments.

Although there are no currently applicable standards for TP in Cherry Creek Reservoir, WQCC Regulation 31 (Reg 31) specifies interim nutrient criteria for warm water reservoirs greater than twenty-five (>25) acres. During the WQCC's April 2023 rulemaking hearing for lake nutrients, nutrient standards were adopted in all lakes and reservoirs upstream of domestic wastewater dischargers. For those lakes downstream of domestic wastewater dischargers, like Cherry Creek Reservoir, the standards were adopted with a delayed effective date of December 31, 2027. On the effective date the standards will become effective in Cherry Creek Reservoir unless a site specific standard is developed and adopted by the WQCC. The 2012 warm water TP criterion for large warm reservoirs is 83 µg/L TP as a summer (July 1-September 30) average in the mixed layer (median of multiple depths), with an allowable exceedance frequency of one-in-five years. The WQCC TP standard will be 47 µg/L in 2027. Figure 60 shows the historical seasonal (July to September) median concentration and the WY 2023 median and mean for the three sites in the photic zone (0-3 m) plotted against the 2012 criteria represented by the orange line and the 2027 standards represented by the purple line. The WY 2023 seasonal mean of 135.9 µg/L is much higher than the last two years and the highest seasonal TP concentration observed since 2011 and

2012. The long term median seasonal phosphorus concentrations average 92 ug/L between the three sites in Cherry Creek Reservoir (Figure 61).

In WY 2023 the monthly median concentrations were below the baseline median in October through December 2022, but at or above the baseline median for the rest of the year and above the 85th percentile from May through August 2023 (Figure 62). The WY 2023 data suggests that the elevated TP concentrations in the Reservoir throughout the year are contributing to the eutrophic conditions.

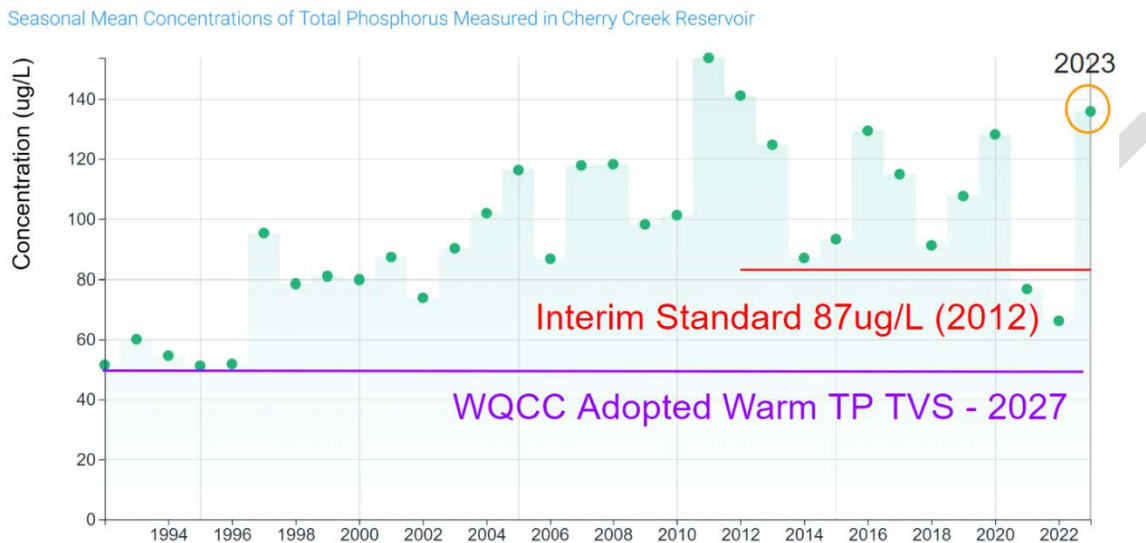


Figure 60. Seasonal Mean Total Phosphorus Concentrations in Cherry Creek Reservoir.

Cherry Creek Reservoir Seasonal Total Phosphorus Concentrations

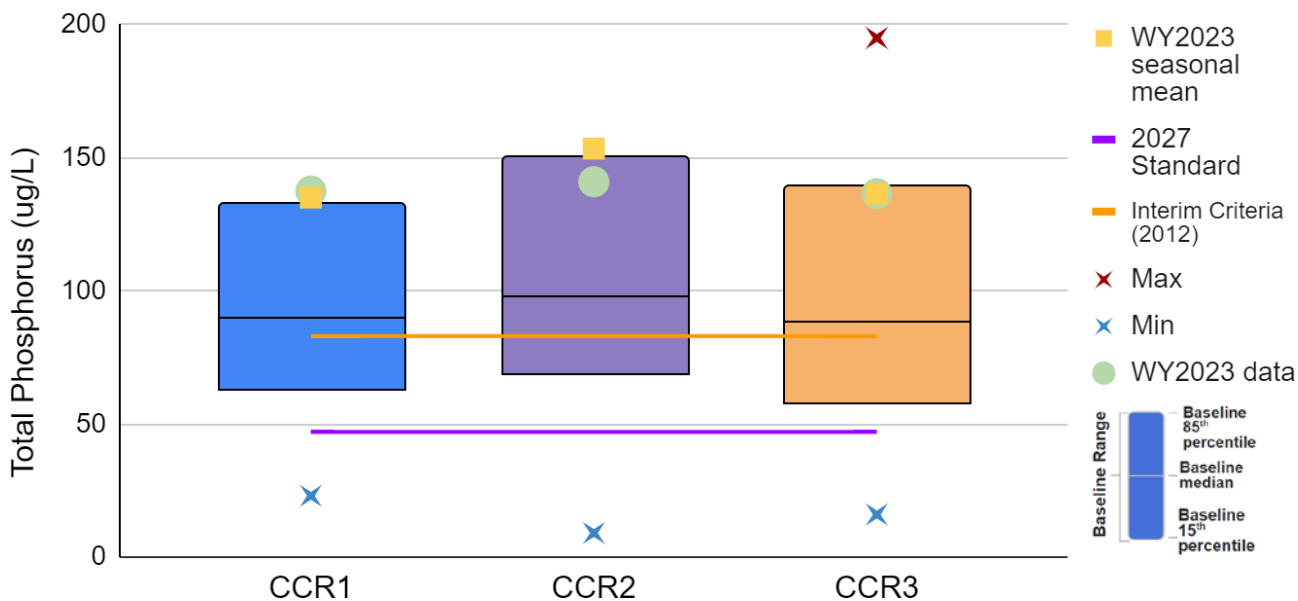


Figure 61. Seasonal TP Concentrations in Photic Zone, Cherry Creek Reservoir, Summary Statistics (1992-2023), WY 2023 medians and means.

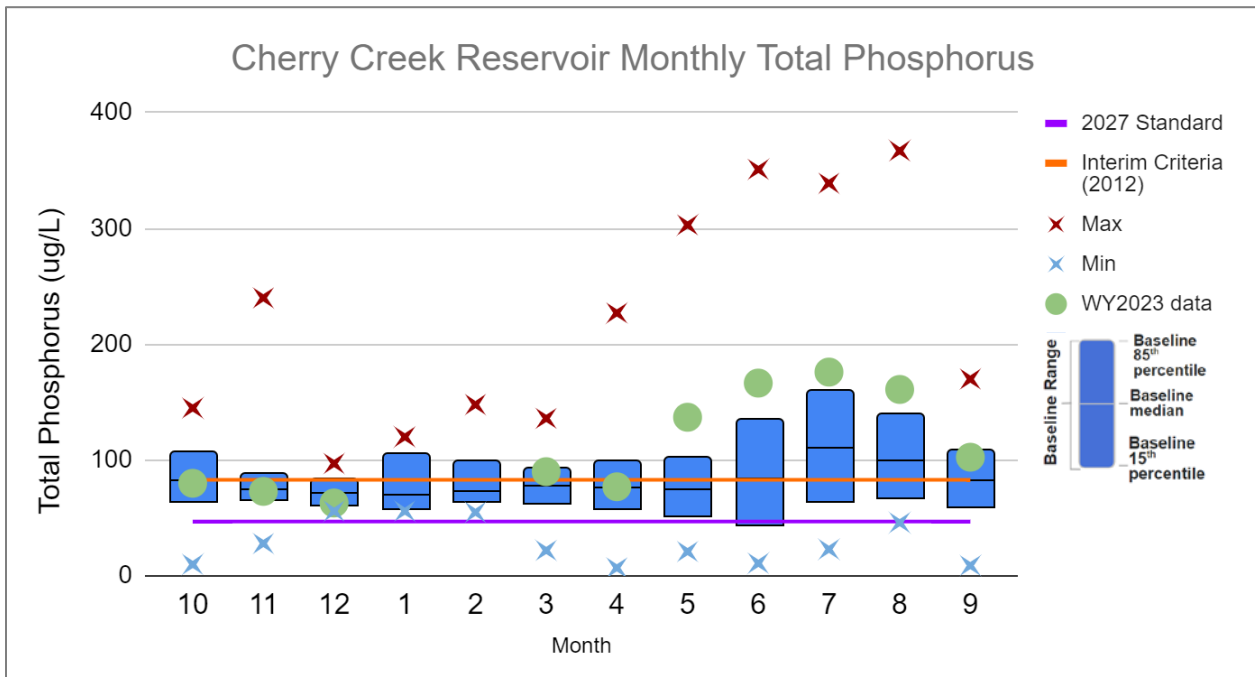


Figure 62. Monthly Median Total Phosphorus in Cherry Creek Reservoir, Summary Statistics and WY 2023 medians.

Figure 63 displays the TP concentrations depth variability through WY 2023 in Cherry Creek Reservoir. The highest concentrations in the photic zone (0-3 m) were seen during the late spring and summer of 2023. The samples from below the photic zone had TP concentrations generally increased with depth and were highest in bottom samples from late May through September. The TP depth profiles at Reservoir monitoring station CCR-2, and the concentrations from the photic zone composite at CCR-1 and CCR-3, available on the data portal, show similar results.

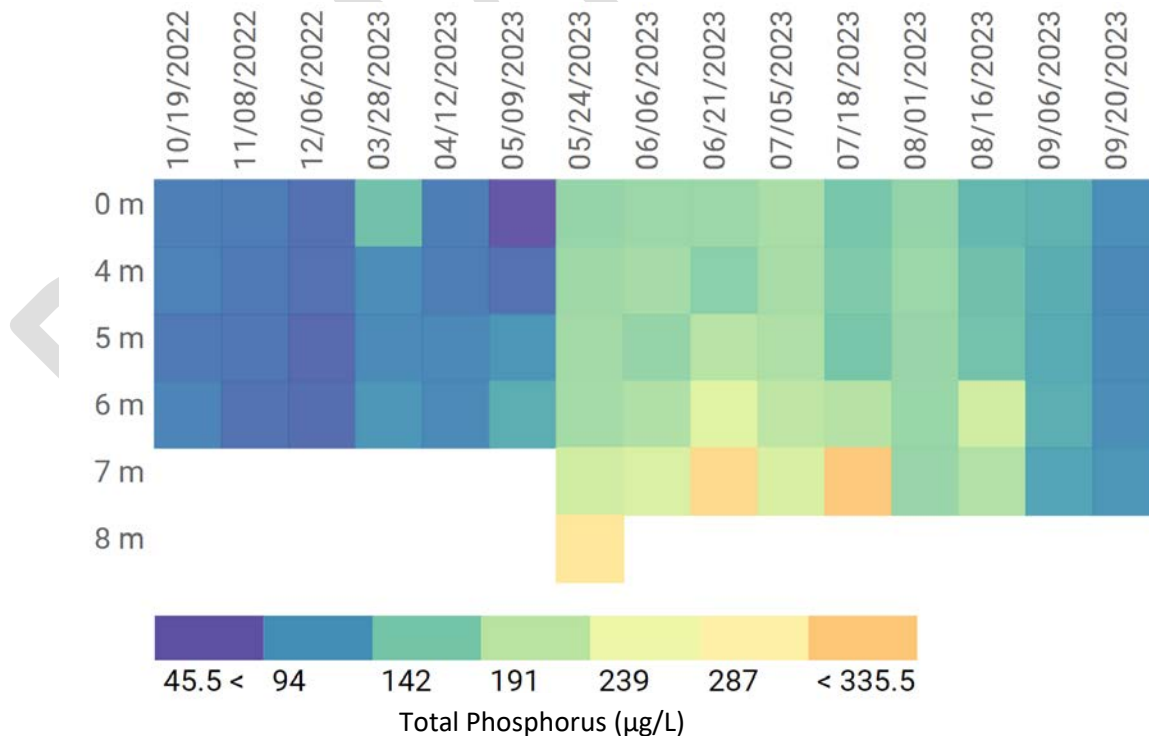


Figure 63. Total Phosphorus Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

Phosphorus increases in the hypolimnion can be caused by internal legacy sediment loading or result from the decomposition of algal cells and other organic matter settling from higher levels in the water column. Inflows of cold runoff water, which has a higher density than warmer surface waters and sinks to the bottom as it enters a lake, can also directly increase hypolimnetic nutrient concentrations. In years with limited stormflows, the higher nutrient concentrations at depth are more likely due to organic deposition and decomposition or internal loading.

4.11 DISSOLVED AND SOLUBLE REACTIVE PHOSPHORUS

Total dissolved phosphorus (TDP includes dissolved organic and inorganic material. Dissolved inorganic phosphorus is usually reported as soluble reactive phosphorus (SRP), which represents the bioavailable form of phosphorus that is readily available for uptake by algae.

Figure 64 and Figure 65 depict the profiles of TDP and SRP from site CCR-2 during WY 2023. Monthly median TDP concentrations average approximately 30% of the total phosphorus concentrations and SRP averages approximately 15%.

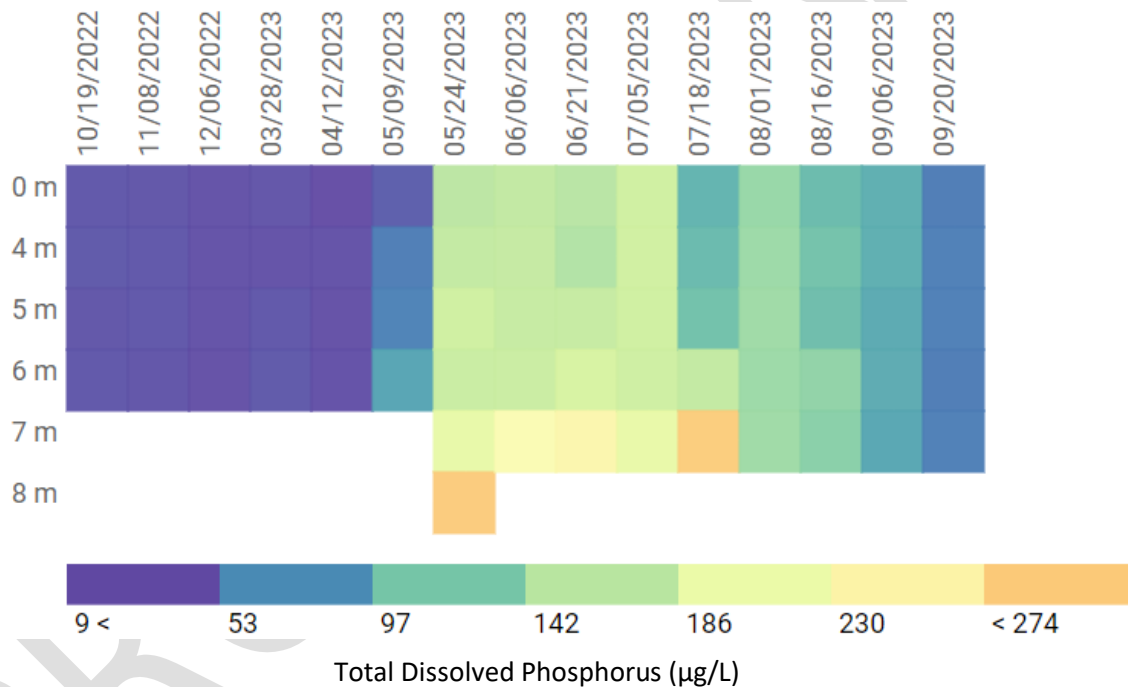


Figure 64. Total Dissolved Phosphorus Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

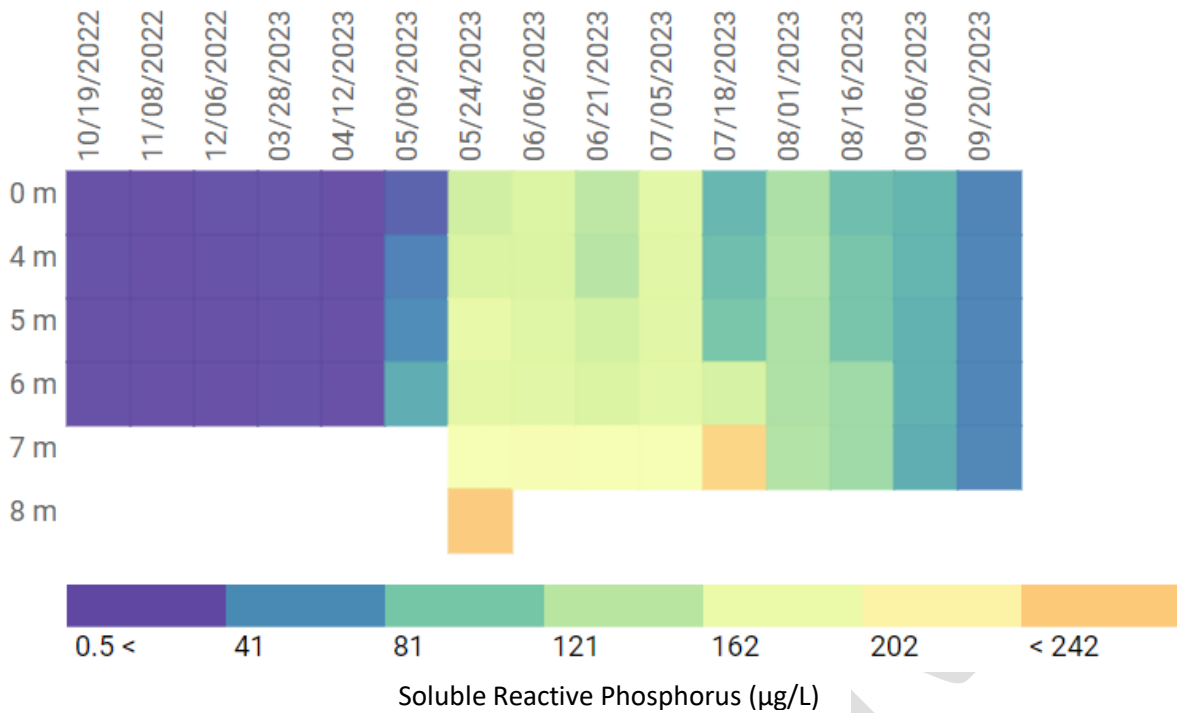


Figure 65. Soluble Reactive Phosphorus Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

During WY 2023, both TDP and SRP remained relatively constant through late fall and winter, but levels throughout the water column show much more variability as the temperatures warm and the season progresses. Since SRP is the bioavailable form of phosphorus, it is typical to see decreases in SRP concentrations in the photic zone through the summer months as productivity increases and phytoplankton and other organisms incorporate SRP into cell material. There was an association of lower levels of TDP and SRP during events when DO levels were low and pH was elevated. Similar patterns of internal loading are observed with these forms of phosphorus during the warmer summer month when DO concentrations are low at the bottom of the Reservoir. As the season progressed, primary productivity in the photic zone was utilizing the available forms of phosphorus as they were released and mixed throughout the water column.

4.12 TOTAL NITROGEN

Nitrogen in aquatic systems comes from many possible natural and anthropogenic sources, including fertilizers, animal and human waste, organic plant matter, and even the air. Nitrogen is often abundant in lakes and reservoirs but when limited, cyanobacteria can utilize (or “fix”) nitrogen gas diffused in the water that provides a competitive advantage over other algae species.

Although there are no currently applicable standards for TN in Cherry Creek Reservoir, WQCC Regulation 31 specifies interim nutrient criteria for warm water reservoirs greater than twenty-five (>25) acres. Like TP, TN standards were adopted in all lakes and reservoirs upstream of domestic wastewater dischargers. After December 31, 2027 standards adopted will become effective in Cherry Creek Reservoir unless site specific standards are developed and adopted by the WQCC. The 2012 warm water total nitrogen criterion for large reservoirs is 910 µg/L TP as a summer (July 1-September 30) average in the mixed layer (median of multiple depths), with an allowable exceedance frequency of one-in-five years. The WQCC 2027 standard for TN will be 640 µg/L in 2027.

Figure 66 shows the historical seasonal mean (July to September) TN concentration from the three sites in the photic zone (0-3 m) plotted against the 2012 criteria represented by the red line and the 2027 standard represented by the purple line. The WY 2023 seasonal mean of 801.8 µg/L is lower than the last three years.

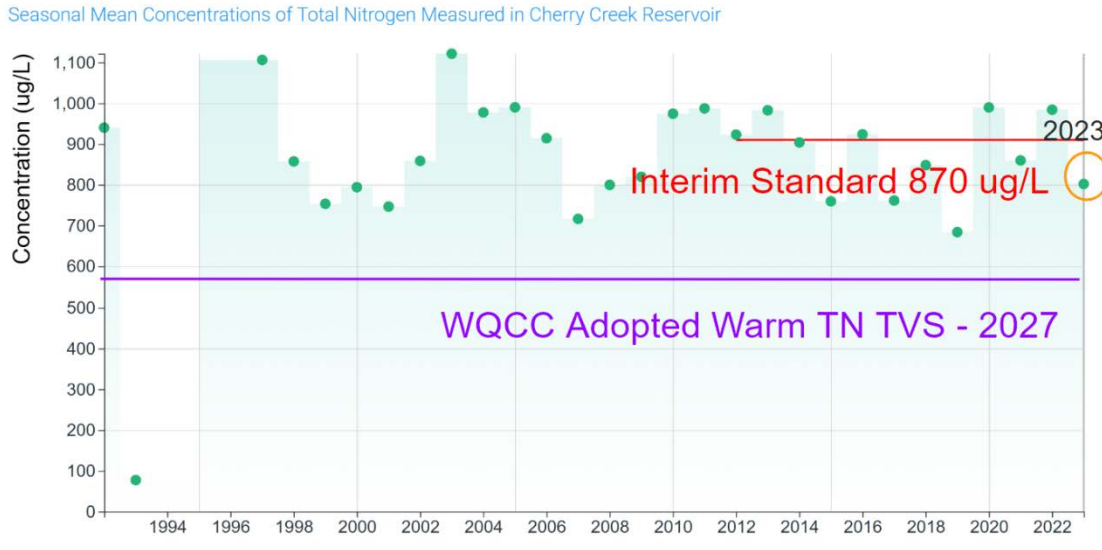


Figure 66. Seasonal Mean Total Nitrogen Concentrations in Cherry Creek Reservoir.

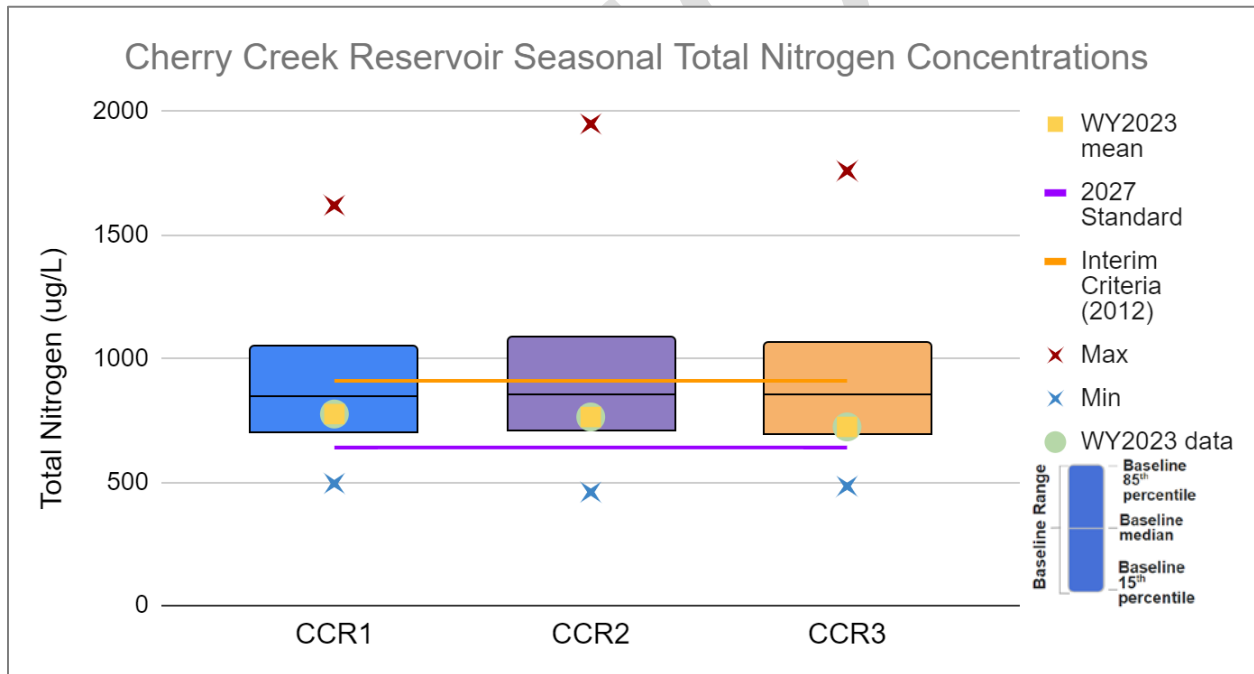


Figure 67. Seasonal Total Nitrogen Concentrations in the Photic Zone, Cherry Creek Reservoir, Summary Statistics (1992-2023), WY 2023 medians and means.

During WY 2023, the monthly median TN concentrations varied and were near or above the baseline monthly medians in October through December 2022 and March through April 2023 (Figure 68). However, concentrations were much lower in early May and then increased to well above the baseline median in June then decreased as the season progressed. When evaluating TN with depth from the samples collected at CCR-2 during WY 2023 (Figure 69), the seasonal changes concentrations observed were consistent throughout the water column. The data from the other two monitoring sites from the photic zone are available on the data portal.

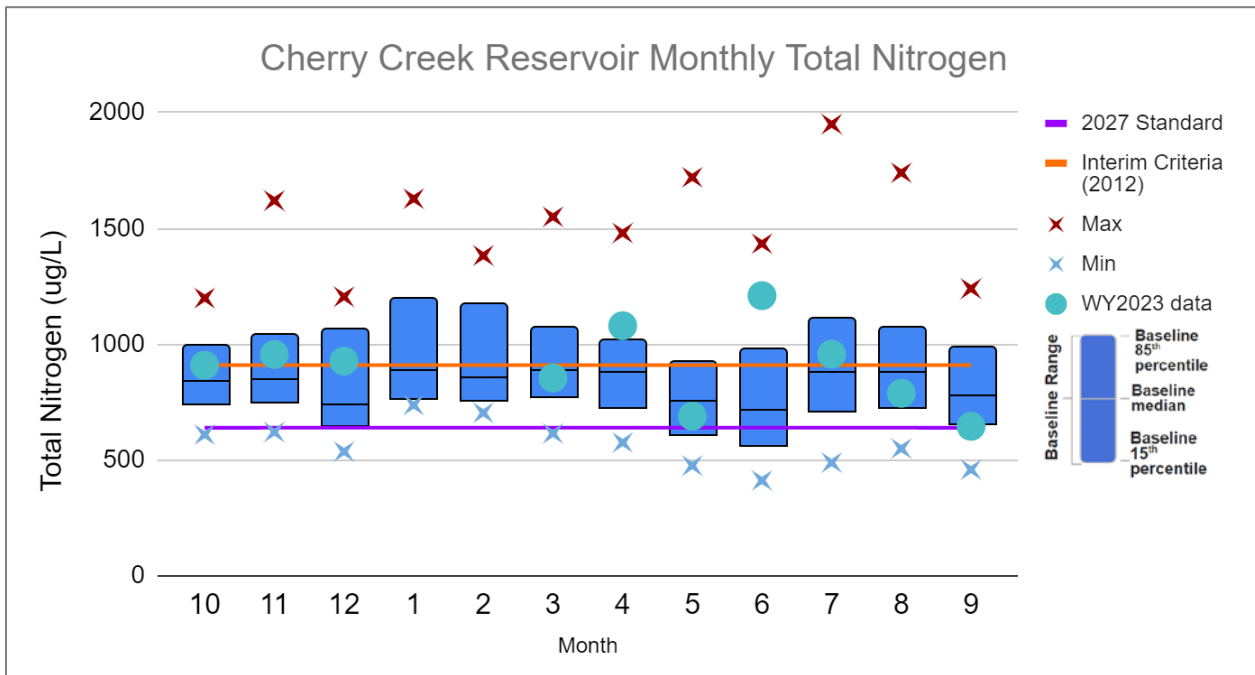


Figure 68. Monthly Total Nitrogen Concentrations, Summary Statistics and WY 2023 medians.

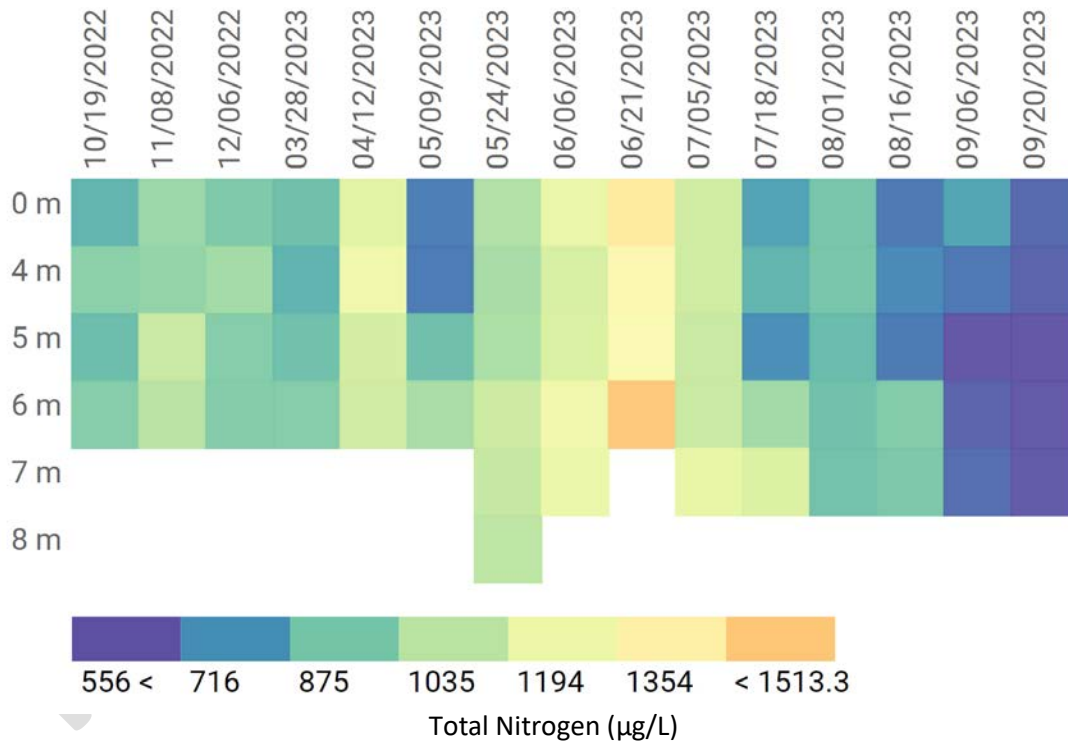


Figure 69. Total Nitrogen Depth Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

4.13 TOTAL INORGANIC NITROGEN

Total Inorganic Nitrogen (TIN) is calculated as the sum of nitrate-nitrite-N ($\text{NO}_3+\text{NO}_2\text{-N}$) and ammonia-N ($\text{NH}_3\text{-N}$) concentrations and represents the forms of nitrogen that are immediately available for algal growth. Figure 70

and Figure 71 illustrate $\text{NO}_3+\text{NO}_2\text{-N}$ and $\text{NH}_3\text{-N}$ concentrations separately, but both were very low and often below the detection limit during WY 2023. TIN concentrations were elevated in June and July at the deeper sampling sites. Possible reasons for the high TIN concentrations in the hypolimnion are decomposition processes and internal nitrogen loading.

Nitrate is the predominant form of inorganic nitrogen when oxygen is present, and ammonia is the predominant form in the absence of oxygen. Phytoplankton can incorporate ammonia directly into cellular material but readily convert nitrate to ammonia when nitrate dominates.

Nitrates were generally low in the photic zone of Cherry Creek Reservoir throughout WY 2023 except for June and early July. On 11 of the 15 monitoring events in WY 2023, $\text{NO}_3+\text{NO}_2\text{-N}$ concentrations were below the detection limit of $5 \mu\text{g/L}$ in the photic zone (0-3 m) at CCR-2. When $\text{NO}_3+\text{NO}_2\text{-N}$ concentrations are low, it is an indicator that algal growth in the Reservoir is limited by nitrogen concentrations.

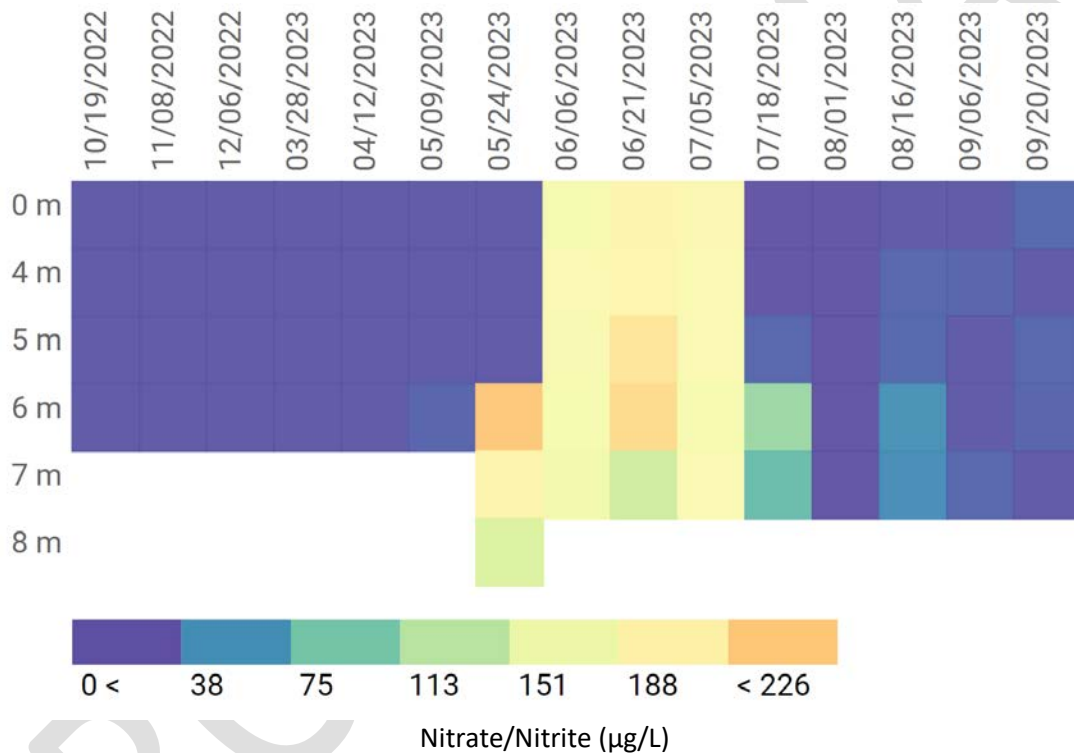


Figure 70. Nitrate/Nitrite Depth Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

Ammonia concentrations, shown as $\text{NH}_3\text{-N}$ (Figure 71) were elevated at depth from May through July, but lower in surface water on most dates. This is an indication of a highly productive reservoir. Ammonia, like nitrate, is a readily available form of nitrogen for algal growth.

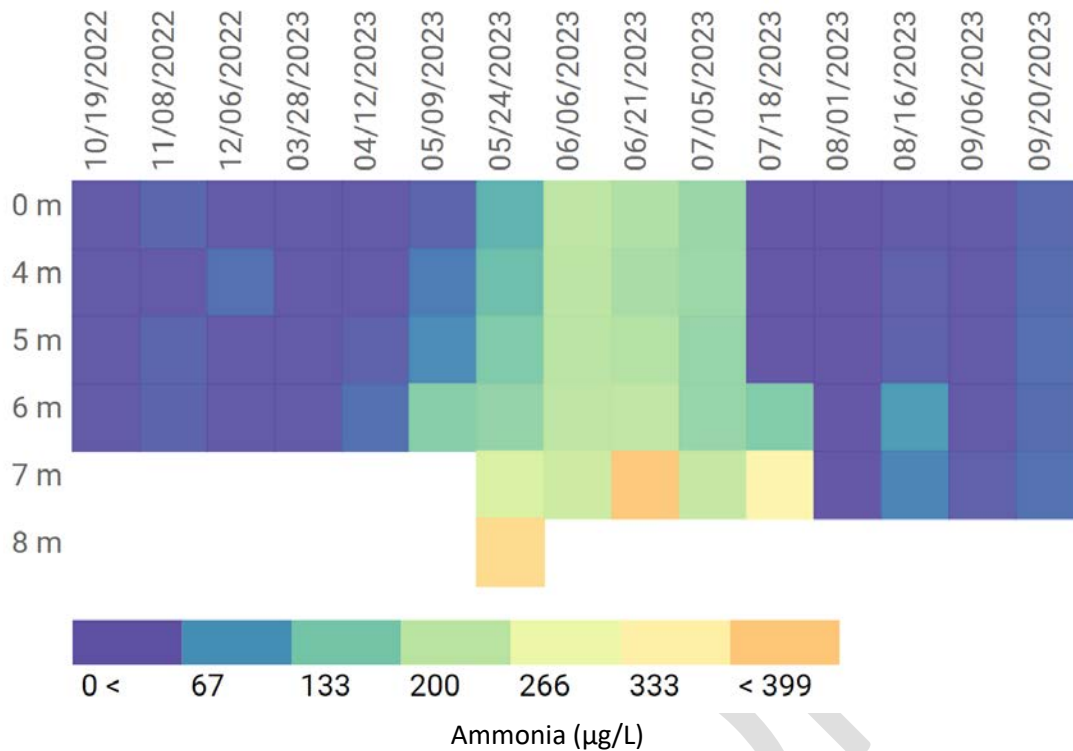


Figure 71. Ammonia Depth Profile at CCR-2, Cherry Creek Reservoir, WY 2023.

On 8 of the 15 monitoring events in WY 2023, $\text{NH}_3\text{-N}$ concentrations were below the detection limit of 5 $\mu\text{g/L}$ in the photic zone (0-3 m) at CCR-2 and 11 of 13 concentrations were below 20 $\mu\text{g/L}$. The increases in ammonia concentrations in the deeper layers also correlated to the periods of lower oxygen at the bottom of the Reservoir. These elevated ammonia values also corresponded to the dates of the lower chl α concentrations. These concentrations are likely due to the release of ammonia from phytoplankton as the bloom that was present died off following the extended period of precipitation.

4.14 LIMITING NUTRIENT

Nitrogen and phosphorus are the nutrients that usually limit algal growth in natural waters. Both the relative concentrations of nitrogen and phosphorus and the absolute concentrations of these nutrients play important roles in structuring phytoplankton communities (Schindler, 1977; Reynolds, 1986). The average nitrogen to phosphorus (N:P) ratio of healthy, growing algal cells is about 7 to 1 by weight (or about 16 to 1 by molar ratio). This value, known as the Redfield ratio, is generally assumed to be the ratio in which these nutrients are ultimately required by algal cells (Reynolds, 1986). Generally, large N:P ratios (>7) indicate that the growth of the phytoplankton community will be limited by the concentration of phosphorus present, while small N:P ratios (<7) indicate that growth will be limited by nitrogen concentrations (Schindler, 1977). The ratios of total inorganic nitrogen (TIN = nitrate + nitrite-N + ammonia-N) to SRP may be more meaningful than the ratio of TN to TP because the inorganic nutrient forms are more directly available to support the growth of aquatic organisms. The potential for cyanobacteria to fix atmospheric nitrogen may be one of the main factors leading to a phytoplankton community dominated by cyanobacteria (see Section 5.1). In lakes and reservoirs with nitrogen limitation, cyanobacteria populations have an advantage over other types of algae and can easily dominate populations and limit diversity.

Figure 72 plots the nutrient mass ratios of TN:TP (in blue), TDN:TDP (in green), and TIN:SRP (in orange). The lines indicate the mass ratio of nitrogen to phosphorus indicating whether nitrogen or phosphorus is limiting. Chl α is plotted on the secondary axis in a red dotted line and the point of limitation is the purple dotted line.

The graph shows that for almost all of the growing season all forms of nitrogen were limited in Cherry Creek Reservoir. Although there was some variability, the concentrations of chl α had relatively higher values following limitation of one or more forms of nitrogen. (See Phytoplankton Section 4.15).

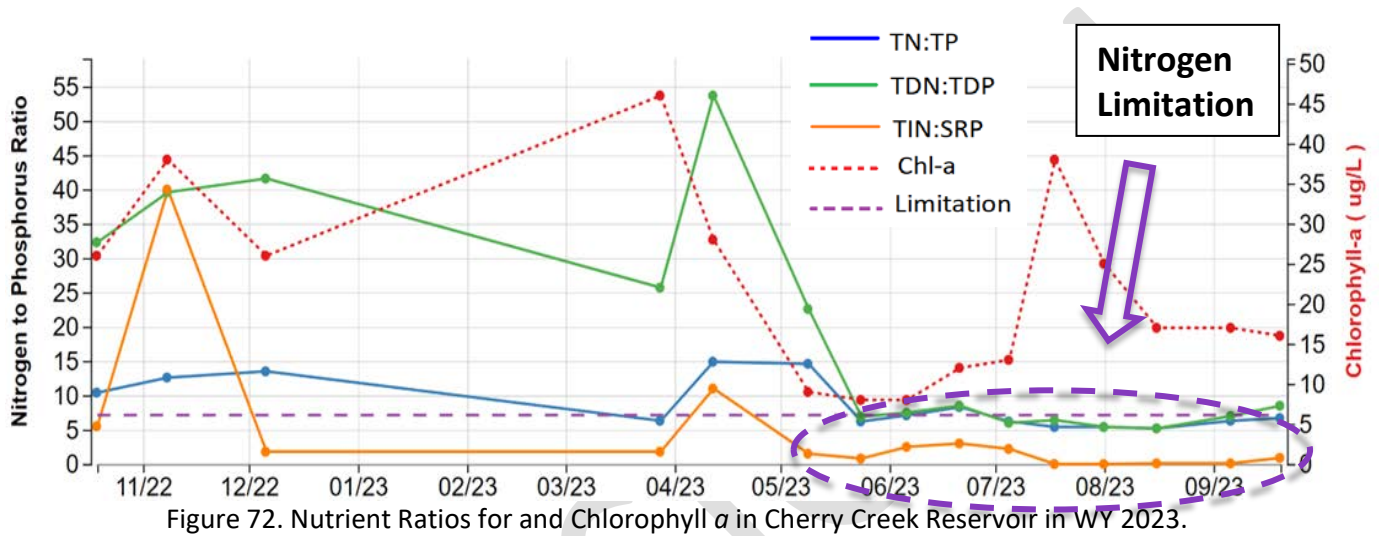


Figure 72. Nutrient Ratios for and Chlorophyll α in Cherry Creek Reservoir in WY 2023.

4.15 TROPHIC STATE ANALYSIS

The trophic state of a lake is a relative expression of the biological productivity of a lake. Two approaches to TSI are presented below, one based on the Carlson index and one based on EPA criteria.

Carlson Index

The Trophic State Index (TSI) developed by Carlson (1977) is among the most commonly used indicators of lake trophic state. This index is expressed as three separate indices based on observations of TP concentrations, chl α concentrations, and Secchi depths from a variety of lakes. TP is used in the index because phosphorus is often the nutrient limiting algal growth in lakes. Chl α is a plant pigment present in all algae and is used to provide an indication of the algal biomass in a lake. Secchi depth is a common measure of the transparency of lake water. The three are related in many lakes because transparency is often limited by algal growth and algal growth can be limited by phosphorus in productive lakes. However, the high phosphorus concentrations in Cherry Creek Reservoir often indicate nitrogen limiting conditions.

Mean values of TP, chl α , and Secchi depth for an individual lake are logarithmically converted to a scale of relative trophic state ranging from 1 to 100. Elevated values for the TSI are indicative of higher productivity. A TSI of less than 35 indicates oligotrophic conditions, a TSI between 35 and 50 indicates mesotrophic conditions, and a TSI greater than 50 indicates eutrophic conditions. Hypereutrophic, or excessively productive lakes, have TSI values greater than 70. Higher numbers are associated with increased probabilities of encountering nuisance conditions, such as algal scum.

Trophic state indices for Cherry Creek Reservoir from WY 2023 are presented in Table 18. These values were calculated using the average of the photic zone (0-3 m) composite samples collected at stations CCR-1, CCR-2,

and CCR-3 during the months of May through September because Carlson (1977) suggested that summer average values may produce the most meaningful results.

Table 18. Trophic State Indices for Cherry Creek Reservoir WY 2018-2023.

Year	Trophic State Index (TSI)		
	Total P	Secchi Depth	Chlorophyll α
2023	76	55	58
Trophic State	Hypereutrophic	Eutrophic	Eutrophic

Figure 73 displays the historical TSI for Cherry Creek Reservoir for each of the parameters for the May-September averages for TP, Secchi depth, and chl α from 2002 to 2023. Based on this index, Cherry Creek Reservoir is considered eutrophic for Secchi depth and chl α , and ranges between eutrophic and hypereutrophic based on TP concentrations. Although the TSI has shown variability over time, the TSI for TP in WY 2023 was the highest observed since 2002. This high TSI value for TP can be attributed to the high concentrations of phosphorus in the stream inflows during the large storm events in WY 2023.

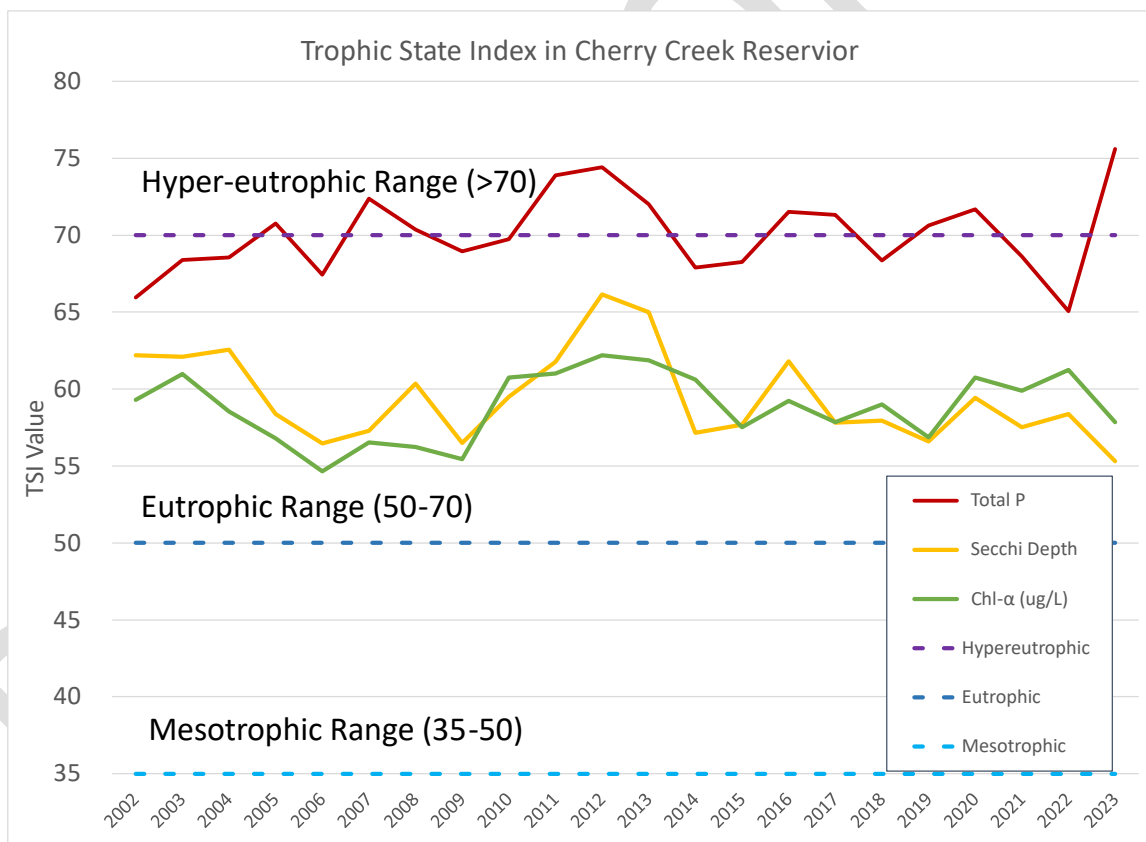


Figure 73. Trophic State Index for Cherry Creek Reservoir (2002-2023).

EPA Trophic State Criteria

Trophic state can also be assessed by comparing monitoring data to trophic state criteria, such as those developed by the U.S. EPA (1980). Table 19 presents a comparison of Cherry Creek Reservoir monitoring data from WY 2023

(May-September) to EPA trophic state criteria. Values for the various parameters were the same averages used to calculate the trophic state indices.

Table 19. Comparison of Cherry Creek Reservoir Monitoring Data to EPA Trophic State Criteria WY 2023.

Trophic State	Characteristic			
	Total P (mg/L)	Chlorophyll α ($\mu\text{g/L}$)	Secchi Depth (m)	Relative Productivity
Oligotrophic	< 0.005	< 2.0	> 8	Low
Mesotrophic	0.005 - 0.030	2.0 - 6.0	4 – 8	Moderate
Eutrophic	0.030 - 0.100	6.0 - 40.0	2 – 4	High
Hypereutrophic	> 0.100	> 40.0	< 2	Excessive
Cherry Creek Reservoir	0.142	16.16	1.38	High

The trophic state criteria in Table 19, like calculated trophic state indices, are based on somewhat arbitrary concentrations that are typically found when the average lake user perceives that water quality problems exist. Comparison of monitoring data from Cherry Creek Reservoir to the EPA trophic state criteria indicate that conditions in Cherry Creek Reservoir are in the eutrophic range for chl α concentrations and hyper-eutrophic for TP and Secchi depth.

The trophic state based on the EPA criteria is slightly different than the Carlson index calculations. It is important to consider that sometimes the trophic state related to Secchi depth alone can be misleading since conventional trophic state criteria assume that Secchi depth is related primarily to algal turbidity. Inorganic turbidity can be a more important factor in determining water clarity for many reservoirs, where Secchi depth does not always provide a good indication of trophic state since these measurements cannot distinguish between algal productivity and inorganic suspended sediment. Inorganic turbidity plays a role in water transparency and associated Secchi depths in Cherry Creek Reservoir as well.

Although these two methods use slightly different calculations and ranges, both the Carson Index and EPA criteria indicate eutrophic to hypereutrophic conditions of Cherry Creek Reservoir for each of the individual parameters evaluated.

4.16 PRECIPITATION

The rain that falls on the Reservoir also serves as a nutrient source and is considered an inflow in the nutrient balance. The TP and TN baseline median, summary statistics and median concentrations for the samples collected from the storms in WY 2023 are displayed in Figure 74. The baseline median is used to calculate the TP and TN added to the Reservoir based on daily precipitation and the surface area.

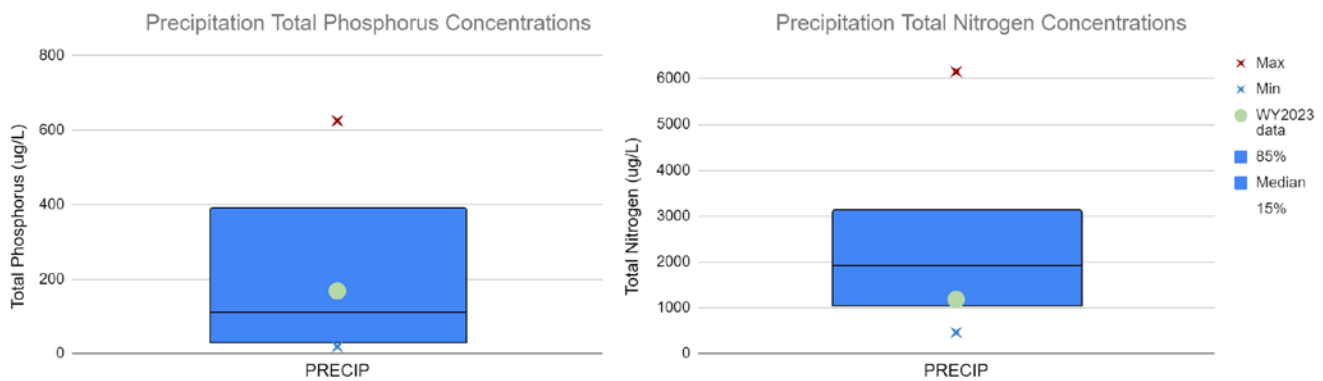


Figure 74. Total Phosphorus and Nitrogen in Precipitation, Summary Statistics and WY 2023.

4.17 PLANKTON DYNAMICS

Analyses of phytoplankton and zooplankton samples were used to assess biological conditions in Cherry Creek Reservoir during WY 2023. Both numbers of individuals (cells/mL for phytoplankton and animals/L for zooplankton) and biovolume ($\mu\text{m}^3/\text{mL}$ for phytoplankton) or biomass ($\mu\text{g}/\text{L}$ for zooplankton) were reported.

4.17.1 PHYTOPLANKTON

Phytoplankton are photosynthetic organisms that are the primary producers in aquatic systems. They form the base of aquatic food chains and are grazed upon by zooplankton and herbivorous fish. A healthy lake should support a diverse assemblage of phytoplankton, in which many algal groups are represented.

In many environmental instances, algal numbers (cells/mL) and algal biovolume ($\mu\text{m}^3/\text{mL}$) closely correlate with one another, but that is not always the case. It is possible, and a common occurrence, for a phytoplankton community to have a large number of very small-sized algal cells, particularly in systems, such as Cherry Creek Reservoir, that have high numbers of cyanobacteria (Cyanophyta), commonly referred to as blue-green algae. At other times, the phytoplankton community can be dominated by a few algal species that are very large in size.

Phytoplankton samples were collected at site CCR-2 from the photic zone (0-3 m composite sample) and analyzed to identify and quantify the populations present on each sampling date. The results from WY 2023 indicate high productivity with diverse populations.

Due to factors outside of the CCBWQA's control, some of the phytoplankton data from the end of WY 2023 is not available. As soon as it can be analyzed, the phytoplankton chapter will be completed and provided in the amended final report.

4.17.2 ZOOPLANKTON

Zooplankton are microscopic animals that consume algae and bacteria in the water column. Some types of zooplankton feed on algae, some on other zooplankton, and some take in both plant and animal particles. Monitoring populations is important because larger zooplankton can exert significant grazing pressure on algal cells; however, they are also subject to predation as they are a food source for larger crustaceans, aquatic insects, and fish. Zooplankton populations in lakes vary with temperature, food supply, and other environmental factors, with reported populations ranging from a few to several hundred individuals per liter (Hutchinson, 1967). Very little detailed information is available on zooplankton dynamics and populations in reservoirs,

although turbidity, increased flow, and other factors probably reduce their numbers to below those observed in natural lakes (Marzolf, 1990).

Most freshwater zooplankton are part of only three phyla: *Arthropoda*, which includes cladocerans, copepods, and ostracods; *Rotifera*; and *Protozoa*. Cladocerans and copepods are microscopic crustaceans that feed primarily on phytoplankton, while ostracods are omnivores and eat both small phytoplankton and other organic material. Larger organisms in these groups can be an important food source for fish and can also exert grazing pressure on phytoplankton populations when present in high enough numbers. Rotifers are microscopic animals that feed on detritus and smaller organisms, such as bacteria. They can also serve as a food source for larger zooplankton. Protozoans are single-celled organisms that feed on other microorganisms, organic matter, and debris.

Zooplankton samples were collected as vertical tows from a depth of 6 m to the surface at station CCR-2 on each sampling date. Zooplankton numbers and diversity were both low compared to average phytoplankton populations in freshwater lakes.

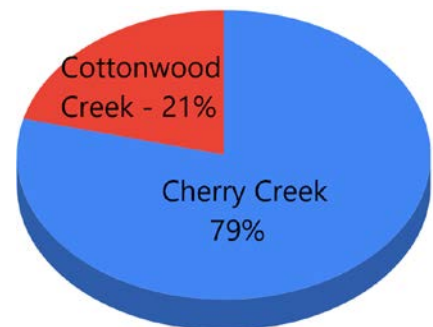
Due to factors outside of the CCBWQA's control, some of the zooplankton data from the end of WY 2023 is not available. As soon as it can be analyzed the zooplankton chapter will be completed and provided in the amended final report.

5.0 WATER BALANCE

Due to circumstances outside of the control of the CCBWQA, some of the inflow data required for the calculations in the water balance is not available. As an alternative, the relative inflow discharge ratio of Cherry Creek to Cottonwood Creek from 2016-2022, along with the inflow, outflow and reservoir storage provided by the USACE will be used. However, the storage information provided by the USACE is also not available due to a discrepancy in the elevation datum shift. This discrepancy should be fixed by the end of January at which time the storage information will be provided, and the required calculations can be completed.

In order to represent the relative inflow contributions for Cherry Creek and Cottonwood as accurately as possible during the periods of time when no data was available the average of the historical values from 2016-2022 were used.

On average, Cherry Creek contributes 79% and Cottonwood Creek contributes 21% of the total surface water flows to Cherry Creek Reservoir.



6.0 FLOW WEIGHTED NUTRIENT CONCENTRATIONS

The nutrient concentrations of the inflows and the outflow of Cherry Creek Reservoir are used to calculate the mass storage on an annual basis. The flow-weighted influent phosphorus goal, derived as part of the 2009 Reg 38 rulemaking process, to achieve the 18 µg/L chl α standard, is 200 µg/L. Flow-weighted nutrient concentrations and mass storage in the Reservoir for WY 2023 will be provided after the water balance has been completed.

8.0 NUTRIENT MASS BALANCES

Following the water balance and flow-weighted nutrient concentrations the mass storage calculations for the Reservoir will be completed following the information provided by the USACE in early 2024.

9.0 WY 2023 WATER QUALITY SUMMARY

The results obtained from the CCBWQA's comprehensive monitoring program documents water quality within the watershed over time. Key findings from monitoring conducted during WY 2023 include:

- Cherry Creek Reservoir did not meet the chl α seasonal standard for WY 2023, but it did meet the Reg 38 standards for temperature, pH, and dissolved oxygen to support the Class 1 Warm Water Aquatic Life classification.
- Cherry Creek Reservoir continues to remain eutrophic to hypereutrophic in regard to total phosphorus, chl α , and transparency of the water. There was a cyanobacteria bloom in 2023 resulting in posting of signage to inform the public or closures to recreational users of the Reservoir due to risk or presence of toxin.
- Surface water flows are the main contributor of nutrient concentrations in the inflows and nutrient loading of the reservoir. The WY 2023 weather and precipitation in the watershed directly impacted the water quantity and quality of Reservoir inflows, internal Reservoir dynamics, and the overall exchange rate.
- The WY 2023 Reservoir conditions due to above average inflows and precipitation resulted in higher water levels, and reduced residence time. However, the high phosphorus concentrations from the flood events increased the potential for algae growth, cyanobacteria blooms, and high chl α concentrations that were present shortly after the rain slowed and the temperatures warmed.
- There continues to be notable differences in water quality between Cherry Creek, Cottonwood Creek, and Piney Creek. Cherry Creek has much higher concentrations of phosphorus, and Cottonwood Creek has higher concentrations of nitrogen. Piney Creek continues to demonstrate lower concentrations of nutrients and suspended solids when compared to Cherry Creek during baseflow conditions. Stream characteristics vary in terms of stream channel morphology, flow patterns, wetlands, vegetation growth patterns, effects of storm events, watershed development, number of permitted wastewater treatment facility discharge outfalls, and differences in runoff from the watersheds all play a role in water quality.
- Conductivity in the streams and groundwater is significantly increasing over time, which impacts Reservoir water quality and dynamics.
- In WY 2023, the constructed wetland PRF ponds on Cottonwood Creek functioned effectively to remove phosphorus and suspended solids during stormflow conditions. In addition, the PRF Ponds on Cottonwood Creek have been functioning effectively when evaluating upstream to downstream concentrations on a long-term basis.

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APPENDICES

APPENDIX A - Cherry Creek Basin Water Quality Authority Monitoring Data, WY 2023.

APPENDIX B – WY 2023 Cherry Creek Reservoir Daily Inflow and Outflow Data and Monthly Summary Information (will be included with amended report)

PROVISIONAL

Appendix A – Cherry Creek Basin Water Quality Authority Monitoring Data, WY 2023

Table 1. Cherry Creek Reservoir, Physical Parameters, WY 2023

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	10/19/2022	11/8/2022	12/6/2022	3/28/2023	4/12/2023	5/9/2023	5/24/2023	6/6/2023	6/21/2023	7/5/2023	7/18/2023	8/1/2023	8/16/2023	9/6/2023	9/20/2023
Light Transmittance [99% Attenuation]	m	CCR-1	1.8	2.3	3.0	2.5	2.7	4.0	4.1	4.4	2.7	3.4	3.4	2.8	2.8	3.1	2.5
Light Transmittance [99% Attenuation]	m	CCR-2	1.9	2.4	2.8	2.4	2.6	3.8	4.1	5.0	2.6	3.6	4.0	2.6	2.8	2.9	2.6
Light Transmittance [99% Attenuation]	m	CCR-3	1.6	2.3	2.9	2.5	2.5	4.1	4.3	4.6	2.5	2.8	3.5	2.4	2.8	2.9	2.5
Light Transmittance [Secchi Depth]	m	CCR-1	0.6	0.8	1.0	0.8	0.7	1.5	2.8	2.5	0.7	1.5	1.2	1.1	1.0	1.1	0.9
Light Transmittance [Secchi Depth]	m	CCR-2	0.7	0.8	1.0	0.7	0.6	1.5	2.9	2.8	0.6	1.6	1.1	1.1	0.9	1.0	0.8
Light Transmittance [Secchi Depth]	m	CCR-3	0.6	0.8	0.9	0.7	0.6	1.7	2.8	2.4	0.6	1.1	1.2	1.0	0.8	1.0	0.8
Oxygen Dissolved	mg/L	CCR-1	8.1	9.9	11.4	12.8	11.1	8.5	7.0	6.4	6.7	6.3	11.8	7.1	7.1	6.0	6.7
Oxygen Dissolved	mg/L	CCR-2	8.4	9.9	11.6	13.1		8.6	6.9	6.3	6.8	6.2	12.0	7.4	7.1	6.4	7.5
Oxygen Dissolved	mg/L	CCR-3	8.4	10.3	11.3	13.1	10.7	8.4	6.8	6.7	6.8	6.2	12.0	7.6	7.5	6.5	7.5
pH	None	CCR-1	8.5	8.6	8.6	8.8	8.6	8.5	8.0	7.9	8.0	7.8	8.9	8.5	8.5	8.4	8.5
pH	None	CCR-2	8.5	8.6	8.5	8.7		8.5	8.0	7.9	8.0	7.9	8.9	8.6	8.5	8.4	8.6
pH	None	CCR-3	8.5	8.6	8.6	8.7	8.6	8.5	8.0	8.0	8.0	7.9	9.0	8.7	8.6	8.4	8.6
Specific Conductance	uS/cm	CCR-1	1,272	1,354	1,413	1,295	1,286	1,369	860	903	869	815	848	888	904	929	959
Specific Conductance	uS/cm	CCR-2	1,276	1,353	1,387	1,284		1,370	859	900	870	813	851	889	904	908	960
Specific Conductance	uS/cm	CCR-3	1,277	1,352	1,406	1,292	1,284	1,372	867	902	869	819	851	892	905	928	960
Temperature Water	deg C	CCR-1	14.0	7.8	2.7	5.0	10.6	15.7	17.8	18.0	19.7	21.9	24.9	24.0	23.2	21.9	18.8
Temperature Water	deg C	CCR-2	13.9	7.9	2.7	5.3		16.2	17.5	18.4	20.3	21.9	25.3	24.2	23.6	21.9	19.1
Temperature Water	deg C	CCR-3	14.2	7.9	3.0	5.0	10.3	16.6	17.8	18.8	19.8	21.6	26.2	24.2	23.9	22.0	19.5

Table 2. Cherry Creek Reservoir Nutrients and Chemical Parameters, WY 2023

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	10/19/2022	11/8/2022	12/6/2022	3/28/2023	4/12/2023	5/9/2023	5/24/2023	6/6/2023	6/21/2023	7/5/2023	7/18/2023	8/1/2023	8/16/2023	9/6/2023	9/20/2023
Chlorophyll-a	ug/L	CCR-1	25	39	26	42	26	11	6	8	8	14	37	16	17	17	17
Chlorophyll-a	ug/L	CCR-2	26	38	26	46	28	9	8	8	12	13	38	25	17	17	16
Chlorophyll-a	ug/L	CCR-3	27	43	27	45	31	11	8	9	10	15	38	24	22	14	20
Phaeo-a	ug/L	CCR-1	4	50	50	50	50	50	50	50	50	50	2	50	1	10	7
Phaeo-a	ug/L	CCR-2	1	50	50	50	3	50	50	50	50	50	50	50	50	12	7
Phaeo-a	ug/L	CCR-3	4	50	50	3	50	50	50	50	50	50	50	50	50	11	7
Soluble Reactive Phosphorus as P	ug/L	CCR-1	535	538	514	344	499	492	1,010	1,050	1,230	1,050	523	637	489	506	346
Soluble Reactive Phosphorus as P	ug/L	CCR-2	484	554	499	334	483	451	975	1,100	1,180	935	509	638	448	534	350
Soluble Reactive Phosphorus as P	ug/L	CCR-3	522	553	494	348	489	474	940	1,050	1,120	1,020	503	637	440	436	365
Dissolved Phosphorus	ug/L	CCR-1	2	1	5	4	1	8	132	137	116	133	62	101	71	66	36
Dissolved Phosphorus	ug/L	CCR-2	1	1	3	3	1	13	135	145	123	152	67	111	73	66	33
Dissolved Phosphorus	ug/L	CCR-3	1	1	2	5	1	14	133	139	118	144	63	108	73	60	32
Total Phosphorus	ug/L	CCR-1	17	14	11	12	10	19	139	138	140	151	82	116	90	75	46
Total Phosphorus	ug/L	CCR-2	15	14	12	13	9	20	142	147	140	157	80	118	86	76	41
Total Phosphorus	ug/L	CCR-3	15	12	12	13	11	23	139	140	129	153	75	114	92	69	47
Nitrate + Nitrite as N	ug/L	CCR-1	77	72	62	90	77	50	159	157	149	167	139	155	136	123	91
Nitrate + Nitrite as N	ug/L	CCR-2	79	72	64	107	81	69	193	178	197	186	204	164	153	116	90
Nitrate + Nitrite as N	ug/L	CCR-3	82	79	62	83	76	60	157	153	149	177	137	162	136	112	95
Total Ammonia as N	ug/L	CCR-1	3	3	3	3	12	3	3	162	192	191	3	3	11	3	3
Total Ammonia as N	ug/L	CCR-2	3	3	3	3	3	3	3	163	184	176	3	3	3	3	12
Total Ammonia as N	ug/L	CCR-3	3	3	12	3	12	3	3	164	188	184	3	3	11	3	3
Dissolved Nitrogen as N	ug/L	CCR-1	3	19	21	3	20	14	105	194	182	157	3	3	13	3	22
Dissolved Nitrogen as N	ug/L	CCR-2	3	17	3	3	3	16	105	203	188	165	3	3	3	3	19
Dissolved Nitrogen as N	ug/L	CCR-3	3	3	3	3	17	20	86	185	171	154	3	3	3	3	14

Total Nitrogen as N	ug/L	CCR-1	535	538	514	344	499	492	1,010	1,050	1,230	1,050	523	637	489	506	346
Total Nitrogen as N	ug/L	CCR-2	484	554	499	334	483	451	975	1,100	1,180	935	509	638	448	534	350
Total Nitrogen as N	ug/L	CCR-3	522	553	494	348	489	474	940	1,050	1,120	1,020	503	637	440	436	365
Total Organic Carbon	mg/L	CCR-2	7.3	7.2	7.3	6.6	6.4	6.8	7.1	6.9	6.8	7.6	7.3	7.0	6.8	6.9	7.3
Dissolved Organic Carbon	mg/L	CCR-2	7.2	7.1	6.7	6.3	6.0	6.4	6.6	6.5	6.1	6.9	7.1	6.4	5.9	6.7	6.3
Total Suspended Solids	mg/L	CCR-1	11	8	6	7	9	4	2	2	3	5	9	7	6	7	6
Total Suspended Solids	mg/L	CCR-2	11	10	6	6	11	4	1	1	2	3	9	6	6	9	9
Total Suspended Solids	mg/L	CCR-3	11	11	3	7	9	4	1	2	2	7	10	10	6	9	9
Total Volatile Suspended Solids	mg/L	CCR-1	6	4	3	6	5	2	1	2	2	1	3	3	1	3	2
Total Volatile Suspended Solids	mg/L	CCR-2	7	5	5	5	9	2	1	1	2	1	4	3	2	4	3
Total Volatile Suspended Solids	mg/L	CCR-3	7	5	3	6	6	2	1	1	1	1	3	4	1	3	3

Table 3. Cherry Creek Watershed Streams Sites Physical Parameters, WY 2023.

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	10/19/2022	11/7-8/2022	12/6/2022	1/10/2023	2/14/2023	3/14/2023	4/12/2023	5/3-4/2023	6/15/2023	7/10/2023	8/9/2023	9/13/2023
Conductivity	umhos/cm	CC-1		394						374				
Conductivity	umhos/cm	CC-2		1170						689				
Conductivity	umhos/cm	CC-4		1060						866				
Conductivity	umhos/cm	CC-5		951						893				
Conductivity	umhos/cm	CC-6		1004						966				
Conductivity	umhos/cm	CC-7		1101						1023				
Conductivity	umhos/cm	CC-8		1118						1057				
Conductivity	umhos/cm	CC-9		1324						1278				
Conductivity	umhos/cm	CC-10		1339				1205		1298				

Conductivity	umhos/cm	CC-Out		1384						1451				
Conductivity	umhos/cm	CC-USGSFRANKTOWN		247						289				
Conductivity	umhos/cm	CC-USGSPARKER		798						797				
Conductivity	umhos/cm	CT-1		1665						1854				
Conductivity	umhos/cm	CT-2		1714						1920				
Conductivity	umhos/cm	CT-P1		2580				4270		3000				
Conductivity	umhos/cm	CT-P2		2450						2970				
Conductivity	umhos/cm	PC-1		2150				2340		2190				
Dissolved Oxygen	mg/L	CC-1		9						8				
Dissolved Oxygen	mg/L	CC-2		6						8				
Dissolved Oxygen	mg/L	CC-4		6						9				
Dissolved Oxygen	mg/L	CC-5		9						8				
Dissolved Oxygen	mg/L	CC-6		11						9				
Dissolved Oxygen	mg/L	CC-7	9	10	10	12	11	10	8	9	8	8	8	8
Dissolved Oxygen	mg/L	CC-8		11						10				
Dissolved Oxygen	mg/L	CC-9		9						7				
Dissolved Oxygen	mg/L	CC-10	8	10	11	11	11	11	9	8	8	7	7	8
Dissolved Oxygen	mg/L	CC-Out	9	10	11	11	11	10	10	9	8	7	7	7
Dissolved Oxygen	mg/L	CC-USGSFRANKTOWN		11						9				
Dissolved Oxygen	mg/L	CC-USGSPARKER		7						7				
Dissolved Oxygen	mg/L	CT-1	11	11	11	10	11	11	11	10	8	7	8	9
Dissolved Oxygen	mg/L	CT-2	10	10	11	11	10	10	9	8	7	7	6	6
Dissolved Oxygen	mg/L	CT-P1	9	10	11	12	12	11	10	11	8	7	8	8
Dissolved Oxygen	mg/L	CT-P2	11	11	12	11	11	11	8	10	7	7	8	8
Dissolved Oxygen	mg/L	PC-1	10	15	13	13	14	12	11	13	8	9	9	9
Dissolved Oxygen, Saturation	%	CC-1		84						93				
Dissolved Oxygen, Saturation	%	CC-2		63						88				
Dissolved Oxygen, Saturation	%	CC-4		68						105				
Dissolved Oxygen, Saturation	%	CC-5		96						104				

Dissolved Oxygen, Saturation	%	CC-6		114						115				
Dissolved Oxygen, Saturation	%	CC-7	99	108	96	112	103	100	91	119	96	98	96	94
Dissolved Oxygen, Saturation	%	CC-8		114						130				
Dissolved Oxygen, Saturation	%	CC-9		93						95				
Dissolved Oxygen, Saturation	%	CC-10	88	103	106	98	112	111	112	104	96	94	92	98
Dissolved Oxygen, Saturation	%	CC-Out	99	104	102	100	101	96	106	96	103	100	101	101
Dissolved Oxygen, Saturation	%	CC-USGSFRANKTOWN		103						100				
Dissolved Oxygen, Saturation	%	CC-USGSPARKER		92						92				
Dissolved Oxygen, Saturation	%	CT-1	123	121	100	85	100	115	150	126	110	96	105	117
Dissolved Oxygen, Saturation	%	CT-2	113	107	99	91	96	97	112	105	85	102	84	71
Dissolved Oxygen, Saturation	%	CT-P1	103	106	104	106	108	101	127	132	101	95	119	104
Dissolved Oxygen, Saturation	%	CT-P2	120	117	116	96	99	110	87	118	86	100	114	109
Dissolved Oxygen, Saturation	%	PC-1	99	159	133	128	128	114	120	161	99	123	111	111
Dissolved Oxygen, Saturation	%	CC-2		63						88				
pH		CC-1		8						8				
pH		CC-2		7						8				
pH		CC-4		7						8				
pH		CC-5		8						8				
pH		CC-6		8						8				
pH		CC-7	8	8	8	8	8	7	8	8	8	8	8	8
pH		CC-8		8						8				
pH		CC-9		8						8				
pH		CC-10	8	8	8	8	8	7	8	8	8	8	8	8

pH		CC-Out	8	9	9	8	8	7	9	8	8	8	9	8
pH		CC-USGSFRANKTOWN		8						8				
pH		CC-USGSPARKER		8						8				
pH		CT-1	8	8	8	8	8	7	8	8	8	8	8	8
pH		CT-2	8	8	8	8	8	7	8	8	8	8	8	8
pH		CT-P1	8	8	8	8	8	7	8	8	8	8	8	8
pH		CT-P2	8	8	8	8	8	7	8	8	8	8	8	8
pH		PC-1	8	8	8	8	8	7	8	8	8	8	8	8
Specific Conductance	uS/cm	CC-1		375						346				
Specific Conductance	uS/cm	CC-2		1095						637				
Specific Conductance	uS/cm	CC-4		997						799				
Specific Conductance	uS/cm	CC-5		893						829				
Specific Conductance	uS/cm	CC-6		941						907				
Specific Conductance	uS/cm	CC-7	1214	1047	1110	1177	1062	918	944	958	763	973	594	935
Specific Conductance	uS/cm	CC-8		1069						996				
Specific Conductance	uS/cm	CC-9		1260						1214				
Specific Conductance	uS/cm	CC-10	1007	1275	1298	1394	1270	1110	1147	1223	964	1181	882	1132
Specific Conductance	uS/cm	CC-Out	1272	1353	1392	1496	1507	1467	1293	1368	867	855	896	942
Specific Conductance	uS/cm	CC-USGSFRANKTOWN		234						270				
Specific Conductance	uS/cm	CC-USGSPARKER		757						741				
Specific Conductance	uS/cm	CT-1	1755	1635	1997	2682	2049	1684	1830	1757	1374	1426	1595	1458
Specific Conductance	deg C	CT-2	1751	1680	2084	2512	2045	1801	1810	1804	1252	1409	1503	1347
Specific Conductance	deg C	CT-P1	2440	2472	3419	4034	3327	3904	3308	2751	1875	1548	2013	1739
Specific Conductance	deg C	CT-P2	2456	2347	3456	3922	3137	3746	3257	2719	1847	1547	2004	1761
Specific Conductance	deg C	PC-1	1746	2015	2105	2525	2239	2122	2117	2037	1318	1585	1102	1651
Water Temperature	deg C	CC-1		5						10				
Water Temperature	deg C	CC-2		9						13				
Water Temperature	deg C	CC-4		11						15				
Water Temperature	deg C	CC-5		10						15				
Water Temperature	deg C	CC-6		9						15				
Water Temperature	deg C	CC-7	9	8	4	5	4	5	10	16	15	18	17	15
Water Temperature	deg C	CC-8		8						17				

Water Temperature	deg C	CC-9		8						18				
Water Temperature	deg C	CC-10	8	7	3	2	5	7	14	17	17	21	17	17
Water Temperature	deg C	CC-Out	13	8	2	3	4	4	10	11	17	22	23	21
Water Temperature	deg C	CC-USGFSFRANKTOWN		4						10				
Water Temperature	deg C	CC-USGSPARKER		18						16				
Water Temperature	deg C	CT-1	11	9	2	1	4	7	19	16	19	22	20	21
Water Temperature		CT-2	10	8	2	0	3	6	17	16	18	22	20	17
Water Temperature		CT-P1	13	8	5	1	2	4	18	14	18	22	22	21
Water Temperature		CT-P2	12	9	3	1	2	6	11	14	17	23	22	20
Water Temperature		PC-1	7	9	6	5	3	4	9	15	14	19	18	14

Table 4. Cherry Creek Watershed Streams Sites Nutrients and Chemical Parameter Concentrations, WY 2023, Baseflow.

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	10/19/2022	11/7-8/2022	12/6/2022	1/10/2023	2/14/2023	3/14/2023	4/12/2023	5/3-4/2023	6/15/2023	7/10/2023	8/9/2023	9/13/2023
Nitrate + Nitrite as N	ug/L	CC-10	313	277	684	915	827	345	259	361	*	352	516	502
Nitrate + Nitrite as N	ug/L	CC-7	765	710	1,430	1,190	1,290	490	471	572		512	675	839
Nitrate + Nitrite as N	ug/L	CC-Out	3	3	18	123	165	16	3	3		171	3	16
Nitrate + Nitrite as N	ug/L	CT-1	1,410	1,450	2,810	1,680	2,030	2,950	1,010	1,182		844	1,410	1,580
Nitrate + Nitrite as N	ug/L	CT-2	1,250	1,250	2,530	1,770	2,070	2,820	472	491		495	459	679
Nitrate + Nitrite as N	ug/L	CT-P1	368	326	495	620	508	331	152	173		185	422	439
Nitrate + Nitrite as N	ug/L	CT-P2	544	402	599	708	536	518	327	259		270	618	590
Nitrate + Nitrite as N	ug/L	MCM-1	309		555		709		539				334	
Nitrate + Nitrite as N	ug/L	MCM-2	3		126		306		13				36	
Nitrate + Nitrite as N	ug/L	PC-1	152	77	197	304	151	200	28	92		355	404	405
Total Ammonia as N	ug/L	CC-10	3	3	3	18	3	22	3	23	*	21	16	26
Total Ammonia as N	ug/L	CC-7	10	10	3	19	13	22	3	34		11	22	35
Total Ammonia as N	ug/L	CC-Out	15	36	9	248	361	167	3	3		253	3	62
Total Ammonia as N	ug/L	CT-1	13	15	24	33	374	88	101	46		38	17	28

Total Ammonia as N	ug/L	CT-2	56	50	19	33	276	99	42	37		42	42	68
Total Ammonia as N	ug/L	CT-P1	22	14	14	21	13	66	22	3		31	3	43
Total Ammonia as N	ug/L	CT-P2	24	3	3	12	3	39	22	3		15	37	53
Total Ammonia as N	ug/L	MCM-1	3		3		3		14				3	
Total Ammonia as N	ug/L	MCM-2	3		3		3		3				3	
Total Ammonia as N	ug/L	PC-1	3	3	3	13	3	20	3	3			3	43
Dissolved Nitrogen as N	ug/L	CC-10			1,220	1,570	1,420	800	900		*	1,010	1,320	810
Dissolved Nitrogen as N	ug/L	CC-7			2,140	2,530	2,050	587	1,310			1,300	1,500	1,270
Dissolved Nitrogen as N	ug/L	CC-Out			553	1,130	1,390	710	790			1,070	720	320
Dissolved Nitrogen as N	ug/L	CT-1			3,810	3,690	3,280	3,240	2,210			1,950	2,720	2,250
Dissolved Nitrogen as N	ug/L	CT-2			3,680	3,730	3,240	2,820	1,380			1,430	1,400	1,230
Dissolved Nitrogen as N	ug/L	CT-P1			1,060	1,290	1,150	890	910			920	1,250	850
Dissolved Nitrogen as N	ug/L	CT-P2			1,200	1,500	1,120	1,050	1,130			1,060	1,560	1,010
Dissolved Nitrogen as N	ug/L	PC-1			658	930	810	480	540			1,143	1,280	557
Total Nitrogen as N	ug/L	CC-10	814	776	1,320	1,680	1,460	810	970	768	1,480	1,160	1,580	960
Total Nitrogen as N	ug/L	CC-7	1,430	1,657	2,220	2,600	2,120	920	1,350	1,250	1,960	1,400	1,750	1,350
Total Nitrogen as N	ug/L	CC-Out	841	1,060	924	1,310	1,610	980	1,270	748	950	1,280	920	560
Total Nitrogen as N	ug/L	CT-1	2,770	3,560	3,970	3,780	3,640	3,430	2,500	2,090	1,900	2,280	2,770	2,520
Total Nitrogen as N	ug/L	CT-2	2,520	2,940	3,860	3,830	3,400	2,990	1,710	1,190	1,510	1,570	1,580	1,350
Total Nitrogen as N	ug/L	CT-P1	1,060	1,130	1,200	1,510	1,230	970	960	660	1,380	1,140	1,390	980
Total Nitrogen as N	ug/L	CT-P2	1,210	1,220	1,280	1,510	1,410	1,250	1,300	840	1,800	1,170	1,700	1,200
Total Nitrogen as N	ug/L	MCM-1	737		983		1,230		1,170		900		1,020	
Total Nitrogen as N	ug/L	MCM-2	293		418		810		520		815		705	
Total Nitrogen as N	ug/L	PC-1	669	506	766	1,070	860	500	750	530	1,120	1,250	1,400	930
Soluble Reactive Phosphorus as P	ug/L	CC-10	128	126	92	87	73	83	95	170	*	218	196	155
Soluble Reactive Phosphorus as P	ug/L	CC-7	71	78	52	51	41	42	49	124		164	168	114
Soluble Reactive Phosphorus as P	ug/L	CC-Out	1	5	2	31	65	15	1	24		162	90	49
Soluble Reactive Phosphorus as P	ug/L	CT-1	3	3	3	2	4	5	5	8		29	20	10
Soluble Reactive Phosphorus as P	ug/L	CT-2	5	5	3	2	3	4	4	7		39	31	20
Soluble Reactive Phosphorus as P	ug/L	CT-P1	6	3	4	3	4	6	5	3		35	23	26
Soluble Reactive Phosphorus as P	ug/L	CT-P2	4	2	3	2	3	4	4	8		31	45	21
Soluble Reactive Phosphorus as P	ug/L	MCM-1	387		261		193		198				418	

Soluble Reactive Phosphorus as P	ug/L	MCM-2	248		145		157		134				366	
Soluble Reactive Phosphorus as P	ug/L	PC-1	43	42	45	47	37	31	34	54		119	116	101
Dissolved Phosphorus	ug/L	CC-10	136	129	93	91	83	85	96	174	*	224	205	156
Dissolved Phosphorus	ug/L	CC-7	82	82	58	57	48	44	56	126		174	169	117
Dissolved Phosphorus	ug/L	CC-Out	13	16	13	41	77	26	8	30		178	111	61
Dissolved Phosphorus	ug/L	CT-1	9	9	9	12	12	12	12	13		54	30	20
Dissolved Phosphorus	ug/L	CT-2	13	12	8	13	10	11	10	13		52	41	30
Dissolved Phosphorus	ug/L	CT-P1	9	9	8	11	10	10	8	8		50	33	34
Dissolved Phosphorus	ug/L	CT-P2	7	8	6	10	9	8	7	15		44	54	30
Dissolved Phosphorus	ug/L	MCM-1	411		274		193		210				419	
Dissolved Phosphorus	ug/L	MCM-2	270		171		182		140				380	
Dissolved Phosphorus	ug/L	PC-1	49	46	46	49	47	33	36	61		126	129	103
Total Phosphorus	ug/L	CC-10	172	150	112	102	102	105	121	194	245	297	370	250
Total Phosphorus	ug/L	CC-7	102	98	71	72	71	61	76	162	241	218	287	149
Total Phosphorus	ug/L	CC-Out	85	91	64	97	116	99	68	118	178	228	158	106
Total Phosphorus	ug/L	CT-1	41	36	32	35	45	43	49	52	79	85	77	66
Total Phosphorus	ug/L	CT-2	34	46	23	25	31	39	48	44	86	74	76	54
Total Phosphorus	ug/L	CT-P1	64	42	40	31	36	41	30	45	87	95	72	64
Total Phosphorus	ug/L	CT-P2	38	29	19	27	43	41	41	62	83	91	102	70
Total Phosphorus	ug/L	MCM-1	429		333		196		241		337		446	
Total Phosphorus	ug/L	MCM-2	291		180		194		142		345		411	
Total Phosphorus	ug/L	PC-1	66	46	70	61	77	45	50	87	146	151	174	122
Total Alkalinity	mg/L	CC-10						228						232
Total Alkalinity	mg/L	CT-2												165
Total Alkalinity	mg/L	CT-P1						281						201
Calcium	mg/L	CC-10						114						121
Calcium	mg/L	CT-2												121
Calcium	mg/L	CT-P1						293						157
Magnesium	mg/L	CC-10						17						17
Magnesium	mg/L	CT-2												25
Magnesium	mg/L	CT-P1						68						35
Potassium	mg/L	CC-10						8						8

Potassium	mg/L	CT-2							7					6
Potassium	mg/L	CT-P1							8					8
Sodium	mg/L	CC-10							109					97
Sodium	mg/L	CT-2												133
Sodium	mg/L	CT-P1							492					167
Total Chloride	mg/L	CC-10							166					160
Total Chloride	mg/L	CT-2												224
Total Chloride	mg/L	CT-P1							904					295
Total Organic Carbon	mg/L	CC-10	5	4	4	5	5	4	4	5	7		6	6
Total Organic Carbon	mg/L	CT-2	7	7	6	7	6	7	7	7	7		8	9
Dissolved Organic Carbon	mg/L	CC-10	4	4	4	4	4	4	4	4	6		6	6
Dissolved Organic Carbon	mg/L	CT-2	7	7	6	7	6	7	7	7	6		8	9
Total Sulfate as SO4	mg/L	CC-10							122					123
Total Sulfate as SO4	mg/L	CT-2												160
Total Sulfate as SO4	mg/L	CT-P1							529					250
Total Suspended Solids	mg/L	CC-10	16	4	3	4	6	5	9	12	34	29	68	45
Total Suspended Solids	mg/L	CC-7	4	2	2	3	2	3	3	7	22	14	49	8
Total Suspended Solids	mg/L	CC-Out	13	12	6	3	3	5	7	7	10	15	14	17
Total Suspended Solids	mg/L	CT-1	11	6	11	13	13	12	14	8	8	18	17	18
Total Suspended Solids	mg/L	CT-2	13	5	7	6	7	10	9	8	6	5	4	4
Total Suspended Solids	mg/L	CT-P1	20	18	17	10	7	9	6		12	5	8	10
Total Suspended Solids	mg/L	CT-P2	12	8	5	8	6	9	12	10	10	9	14	15
Total Suspended Solids	mg/L	MCM-1	0		1		1		13		9		5	
Total Suspended Solids	mg/L	MCM-2	0		0		4		1		13		5	
Total Suspended Solids	mg/L	PC-1	5	3	11	6	5	4	1	4	8	3	19	5
Total Volatile Suspended Solids	mg/L	CC-10	1	1	1	1	1	1	1	3	6	5	6	7
Total Volatile Suspended Solids	mg/L	CC-7	0	0	1	0	1	1	1	2	4	3	9	2
Total Volatile Suspended Solids	mg/L	CC-Out	5	6	4	3	2	3	6	4	3	4	3	3
Total Volatile Suspended Solids	mg/L	CT-1	1	1	1	2	2	3	3	2	3	4	4	8
Total Volatile Suspended Solids	mg/L	CT-2	3	1	2	1	1	2	2	4	2	4	1	2
Total Volatile Suspended Solids	mg/L	CT-P1	3	4	5	2	2	3	2		4	3	2	2
Total Volatile Suspended Solids	mg/L	CT-P2	2	2	1	2	1	2	4	4	3	4	3	0

Total Volatile Suspended Solids	mg/L	MCM-1	0		1		1		3		2		1	
Total Volatile Suspended Solids	mg/L	MCM-2	0		0		1		0		2		1	
Total Volatile Suspended Solids	mg/L	PC-1	0	1	2	1	1	1	1	2	1	2	4	2

Table 5. Cherry Creek Watershed Streams Sites Nutrients and Chemical Parameter Concentrations, WY 2023, Stormflow.

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	10/2/2022	4/28/2023	5/11/2023	6/5/2023	6/12/2023	6/22/2023	7/5/2023	7/21/2023
Nitrate + Nitrite as N	ug/L	CC-10				249	409		369	585
Nitrate + Nitrite as N	ug/L	CC-7				429	2450		476	795
Nitrate + Nitrite as N	ug/L	CT-1	558			172	328		453	573
Nitrate + Nitrite as N	ug/L	CT-2	958			190	670		820	364
Nitrate + Nitrite as N	ug/L	CT-P1	413			138	346		311	467
Nitrate + Nitrite as N	ug/L	CT-P2	822			185	462		309	388
Nitrate + Nitrite as N	ug/L	PC-1				162	259		222	50
Total Ammonia as N	ug/L	CC-10				22			3	3
Total Ammonia as N	ug/L	CC-7				25	15		3	3
Total Ammonia as N	ug/L	CT-1	3			21			24	84
Total Ammonia as N	ug/L	CT-2	3			73	19		55	3
Total Ammonia as N	ug/L	CT-P1	20				57		30	3
Total Ammonia as N	ug/L	CT-P2	109			48			12	68
Total Ammonia as N	ug/L	PC-1 - Piney Creek							3	3
Dissolved Nitrogen as N	ug/L	CC-10							1100	
Dissolved Nitrogen as N	ug/L	CC-7							1400	
Dissolved Nitrogen as N	ug/L	CT-1							1200	
Dissolved Nitrogen as N	ug/L	CT-2							1800	
Dissolved Nitrogen as N	ug/L	CT-P1							1200	
Dissolved Nitrogen as N	ug/L	CT-P2							1000	
Dissolved Nitrogen as N	ug/L	PC-1							1000	

Total Nitrogen as N	ug/L	CC-10		1920	2980	1050	2380		1360	1800
Total Nitrogen as N	ug/L	CC-7		1460	3420	1350	2550		1400	2150
Total Nitrogen as N	ug/L	CT-1	1370	2300	4120	863	1230		1490	1900
Total Nitrogen as N	ug/L	CT-2	1810	2610	3180	891	1320		2220	1980
Total Nitrogen as N	ug/L	CT-P1	1210	990	3050	1030	1060		1270	1360
Total Nitrogen as N	ug/L	CT-P2	2000	890	4070	979	1080		1240	1750
Total Nitrogen as N	ug/L	PC-1			3420	1050	1120		1220	1860
Soluble Reactive Phosphorus as P	ug/L	CC-10				177	213		170	217
Soluble Reactive Phosphorus as P	ug/L	CC-7				110	86		155	92
Soluble Reactive Phosphorus as P	ug/L	CT-1	10			46	15		50	4
Soluble Reactive Phosphorus as P	ug/L	CT-2	13			44	32		53	53
Soluble Reactive Phosphorus as P	ug/L	CT-P1	21			4	40		30	58
Soluble Reactive Phosphorus as P	ug/L	CT-P2	20			73	26		24	38
Soluble Reactive Phosphorus as P	ug/L	PC-1				164	22		154	6
Dissolved Phosphorus	ug/L	CC-10				183	216		179	218
Dissolved Phosphorus	ug/L	CC-7				119	92		164	102
Dissolved Phosphorus	ug/L	CT-1	28			52	20		65	18
Dissolved Phosphorus	ug/L	CT-2	23			50	33		68	69
Dissolved Phosphorus	ug/L	CT-P1	27			14	45		44	73
Dissolved Phosphorus	ug/L	CT-P2	30			79	30		37	50
Dissolved Phosphorus	ug/L	PC-1 - Piney Creek				168	22		180	20
Total Phosphorus	ug/L	CC-10		412	602	189	1040	590	368	525
Total Phosphorus	ug/L	CC-7		225	1050	191	617	529	378	382
Total Phosphorus	ug/L	CT-1	86	604	731	103	207	183	117	271
Total Phosphorus	ug/L	CT-2	78	96	102	84	70	97	104	126
Total Phosphorus	ug/L	CT-P1	98	141	348	141	94	939	141	152
Total Phosphorus	ug/L	CT-P2	154	115	333	96	87	952	119	146
Total Phosphorus	ug/L	PC-1 - Piney Creek			2250	193	160	379	259	493
Total Suspended Solids	mg/L	CC-10		246	265	53	930	330	107	138
Total Suspended Solids	mg/L	CC-7		52	717	31	460	240	118	84
Total Suspended Solids	mg/L	CT-1	18	460	240	16	64	77	21	151

Total Suspended Solids	mg/L	CT-2	9	21	15	8	9	6	12	7
Total Suspended Solids	mg/L	CT-P1	24	23	150	36	14	870	18	17
Total Suspended Solids	mg/L	CT-P2	26	18	155	19	12	280	16	22
Total Suspended Solids	mg/L	PC-1			685	21	32	109	33	235
Total Volatile Suspended Solids	mg/L	CC-10		28	35	6	120	55	7	15
Total Volatile Suspended Solids	mg/L	PC-1			685	21	32	109	33	235
Total Volatile Suspended Solids	mg/L	CC-10		28	35	6	120	55	7	15
Total Volatile Suspended Solids	mg/L	CC-7		6	117	4	70	30	12	12
Total Volatile Suspended Solids	mg/L	CC-Out	4	65	53	3	11	14	2	25
Total Volatile Suspended Solids	mg/L	CT-1	3	7	4	2	3	3	3	1
Total Volatile Suspended Solids	mg/L	CT-2	6	6	33	9	3	130	5	2
Total Volatile Suspended Solids	mg/L	CT-P1	8	5	45	5	3	55	3	7
Total Volatile Suspended Solids	mg/L	CT-P2			55	5	9	23	7	55

Table 6. Cherry Creek Watershed Groundwater Monitoring Data, WY 2023.

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	November 2022	May 2023
Dissolved Oxygen	mg/L	Kennedy Station	6.5	5.7
Dissolved Oxygen	mg/L	MW-1 - Monitoring Well 1	2.4	5.0
Dissolved Oxygen	mg/L	MW-5 - Monitoring Well 5	0.6	1.1
Dissolved Oxygen	mg/L	MW-9 - Monitoring Well 9	0.6	1.1
Dissolved Oxygen, Saturation	%	Kennedy Station	74	66
Dissolved Oxygen, Saturation	%	MW-1 - Monitoring Well 1	28	57
Dissolved Oxygen, Saturation	%	MW-5 - Monitoring Well 5	8	12
Dissolved Oxygen, Saturation	%	MW-9 - Monitoring Well 9	7	12
pH		Kennedy Station	7.3	7.3
pH		MW-1 - Monitoring Well 1	6.9	6.5
pH		MW-5 - Monitoring Well 5	6.9	6.9
pH		MW-9 - Monitoring Well 9	7.1	7.1
Specific Conductance	uS/cm	Kennedy Station	1,294	1,262
Specific Conductance	uS/cm	MW-1 - Monitoring Well 1	505	1,160
Specific Conductance	uS/cm	MW-5 - Monitoring Well 5	1,135	1,281
Specific Conductance	uS/cm	MW-9 - Monitoring Well 9	1,346	1,459
Water Temperature	deg C	Kennedy Station	12	12

Water Temperature	deg C	MW-1 - Monitoring Well 1	12	11
Water Temperature	deg C	MW-5 - Monitoring Well 5	16	11
Water Temperature	deg C	MW-9 - Monitoring Well 9	10	11
Soluble Reactive Phosphorus as P	ug/L	Kennedy Station	110	117
Soluble Reactive Phosphorus as P	ug/L	MW-1 - Monitoring Well 1	240	165
Soluble Reactive Phosphorus as P	ug/L	MW-5 - Monitoring Well 5	215	150
Soluble Reactive Phosphorus as P	ug/L	MW-9 - Monitoring Well 9	253	196
Dissolved Phosphorus	ug/L	Kennedy Station	111	127
Dissolved Phosphorus	ug/L	MW-1 - Monitoring Well 1	246	174
Dissolved Phosphorus	ug/L	MW-5 - Monitoring Well 5	215	154
Dissolved Phosphorus	ug/L	MW-9 - Monitoring Well 9	256	209
Total Phosphorus	ug/L	Kennedy Station	234	220
Total Phosphorus	ug/L	MW-1 - Monitoring Well 1	331	199
Total Phosphorus	ug/L	MW-5 - Monitoring Well 5	224	173
Total Phosphorus	ug/L	MW-9 - Monitoring Well 9	257	261
Nitrate + Nitrite as N	ug/L	Kennedy Station	3	3
Nitrate + Nitrite as N	ug/L	MW-1 - Monitoring Well 1	226	2,000
Nitrate + Nitrite as N	ug/L	MW-5 - Monitoring Well 5	512	1,390
Nitrate + Nitrite as N	ug/L	MW-9 - Monitoring Well 9	853	1,160
Total Ammonia as N	ug/L	Kennedy Station	109	55
Total Ammonia as N	ug/L	MW-1 - Monitoring Well 1	2.5	13.0
Total Ammonia as N	ug/L	MW-5 - Monitoring Well 5	2.5	2.5
Total Ammonia as N	ug/L	MW-9 - Monitoring Well 9	2.5	2.5
Total Nitrogen as N	ug/L	Kennedy Station	334	220
Total Nitrogen as N	ug/L	MW-1 - Monitoring Well 1	481	2,590
Total Nitrogen as N	ug/L	MW-5 - Monitoring Well 5	1,585	1,730
Total Nitrogen as N	ug/L	MW-9 - Monitoring Well 9	1,020	1,350
Total Organic Carbon	mg/L	Kennedy Station	3.1	4.4
Total Organic Carbon	mg/L	MW-1 - Monitoring Well 1	4.1	2.8
Total Organic Carbon	mg/L	MW-5 - Monitoring Well 5	3.4	4.2
Total Organic Carbon	mg/L	MW-9 - Monitoring Well 9	2.8	3.6
Dissolved Organic Carbon	mg/L	Kennedy Station	2.8	3.2
Dissolved Organic Carbon	mg/L	MW-1 - Monitoring Well 1	3.9	2.3
Dissolved Organic Carbon	mg/L	MW-5 - Monitoring Well 5	3.2	3.9
Dissolved Organic Carbon	mg/L	MW-9 - Monitoring Well 9	2.5	2.9
Conductivity	umhos/cm	Kennedy Station	1,302	1,356
Conductivity	umhos/cm	MW-1 - Monitoring Well 1	505	1,248

Conductivity	umhos/cm	MW-5 - Monitoring Well 5	1,186	1,376
Conductivity	umhos/cm	MW-9 - Monitoring Well 9	1,414	1,560
Total Chloride	mg/L	Kennedy Station	197	181
Total Chloride	mg/L	MW-1 - Monitoring Well 1	46	156
Total Chloride	mg/L	MW-5 - Monitoring Well 5	152	186
Total Chloride	mg/L	MW-9 - Monitoring Well 9	153	165
Total Sulfate as SO4	mg/L	Kennedy Station	130	132
Total Sulfate as SO4	mg/L	MW-1 - Monitoring Well 1	18	62
Total Sulfate as SO4	mg/L	MW-5 - Monitoring Well 5	107	111
Total Sulfate as SO4	mg/L	MW-9 - Monitoring Well 9	210	248

Table 6. Cherry Creek Watershed Precipitation Nutrient Concentrations, WY 2023.

<i>Constituent</i>	<i>Units</i>	<i>Location Name</i>	4/28/2023	5/11/2023	6/5/2023	6/12/2023	6/22/2023	7/5/2023
Total Nitrogen as N	ug/L	Rain Sampler	836	851	1980	1180		2200
Total Phosphorus	ug/L	Rain Sampler	42	19	338	140	465	196



EXTERNAL MEMORANDUM

To: Cherry Creek Basin Water Quality Authority
Modeling Sub-committee

From: Alan J. Leak, P.E.
Principal
RESPEC
720 South Colorado Blvd., Suite 410 S
Denver, CO 80246

Date: December 27, 2023

Subject: Additional Watershed Model Scenarios and Scenario Approaches

Two additional watershed scenarios were completed using the Cherry Creek 2030 Future Development HSPF model:

1. Reduced WWTF TN concentration (Scenario 12).
2. Scenario 12 plus improved water quality treatment for all developed areas (Scenario 13).

Scenarios 8 through 13 were also rerun using an alternative approach where water quality efficiencies were adjusted using the flow efficiency. This method essentially applies efficiency factors to concentrations after flow reductions instead of directly to loads at the edge of the stream. Details regarding the additional scenarios and new efficiency factor methodology are provided in the following sections. Results are briefly described in this memo with the full results provided in the accompanying Excel document.

ADDITIONAL SCENARIOS

Scenario 12 was developed using Scenario 11 (full 2030 buildout) as the base model. The WWTF TN concentrations were capped at 6 mg/l during the summer (April – September) and 8 mg/l during the winter (October – March). The Pinery, Parker, and Stonegate facilities exceed the seasonal limits 100%, 58%, and 23% of the simulation time-period, respectively. During these periods, the total nitrate-nitrite concentrations were reduced until there were no more exceedances. The Arapaho County Water and Wastewater Authority (ACWWA) facility never exceeded the seasonal TN limits, so those associated time series remained unchanged.

Relative to Scenario 11, the inflow TN loads and concentrations to Cherry Creek Reservoir for Scenario 12 were reduced by 5% and 3%, respectively. The TN load is still 51% higher than the base model, but the concentration is 1% lower. Inflow volume, TSS, and TP remained the same as Scenario 11.

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Scenario 13 was developed using Scenario 12 as the base model. Efficiency factors for flow (0.4), TSS (0.5), TN (0.1), and TP (0.25) that were applied to new development in Scenarios 10 – 12 were applied to all developed model landuse categories.

As expected, Scenario 13 resulted in a reduction in inflow volume and water quality loads to the reservoir relative to Scenarios 11 and 12. However, there was a 10% and 15% increase in inflow TP and TN concentrations, respectively, relative to Scenario 11. These increases are likely due to the enrichment that occurs when the flow efficiency is larger than water quality efficiencies (e.g., when more volume than TN load is removed at the edge of the stream, an increase in inflow concentration is expected). Although applying the efficiencies directly to loads is acceptable, it can produce counterintuitive results, so an alternative methodology was tested.

ALTERNATIVE EFFICIENCY FACTOR APPROACH

The equation below was used to adjust the water quality efficiency factors as a function of the flow efficiency factor for Scenarios 8 – 13.

$$NewEff_{WQ} = (1 - Eff_{Flow}) \times OriEff_{WQ} + Eff_{Flow}$$

where:

NewEff_{WQ} = adjusted water quality efficiency factor

Eff_{Flow} = efficiency factor for flow

OriEff_{WQ} = original water quality efficiency factor

This methodology prevents enrichment in water quality pollutants when the flow efficiency is higher than a water quality efficiency. It also preserves runoff concentrations for parameters that have zero efficiency by setting the efficiency to that of flow. For example, the efficiency for BOD was zero, so the concentrations in runoff actually increased using original method even though the load remained the same. The original and new efficiency factors are summarized in Table 1.

Table 1. Summary of Original and New Efficiency Factors.

Parameter	Original Efficiency	New Efficiency	Original Efficiency	New Efficiency
	Scen 8 – 9	Scen 8 – 9	Scen 10 – 13	Scen 10 – 13
Flow	0.20	0.20	0.40	0.40
TSS	0.50	0.60	0.50	0.70
TP	0.25	0.40	0.25	0.55
TN	0.10	0.28	0.10	0.46
Temperature	0.00	0.20	0.00	0.40
DO	0.00	0.20	0.00	0.40
BOD	0.00	0.20	0.00	0.40
Carbon	0.00	0.20	0.00	0.40









The updates were easily applied and run because changes only involved copying the adjusted factors to existing scenario model files. The new methodology resulted in lower loads and concentrations relative to the same scenarios using the original methodology. Overall, the narrative remains the same regarding inflow to the reservoir where TSS, TP, and TN loads are still substantially larger than the base condition for Scenarios 8-12. Scenario 13 resulted in no change in TP load and slight increase in TSS and TN load relative to the base condition. Furthermore, all inflow concentrations for Scenario 13 were lower than the base results. We recommend using the results from the new methodology because it is a better representation of how BMPs operate in the real world.

Enclosure

Track Methods for Cherry Creek 2030 buildout scenario

C. Lupo - Feb, 2019

Tabs

	Summary of 2030 model runs using direct load efficiency factors (applied at edge of stream)
	Summary of 2030 model runs using treatment train efficiency factors (applied to BMP outflow)
	BMP efficiency/reduction summaries and changes made to the MASS-LINK to represent
	Reach work to represent new PRFs
	SCHEMATIC update summaries and QAQC
	FTABLE extrapolation of Reach 210 to handle higher flows (model error required it)

Scenario descriptions

Base_v2	Base model used - point source TP representation corrected from v1
Scen004	Base model with only the new SCHEMATIC areas/development applied - Base_v2 uses NLCD 2011 while the new one uses 2016 on top of new development conversion
Scen005	Base model with only the new WWTF flows/loads applied
Scen006	Combination of 004 & 005 (new SCHEMATIC with new WWTF flows/loads)
Scen007	Scen006 with PRF parameter changes applied - considered rate reductions of silt/clay scour (M), and settling of BOD (KODSET), phytoplankton (PHYSET), and refractory organics (REFSET)
Scen008	Scen007 with LID/BMPs applied to newly developed areas - see "MASS-LINKS" tab for assumptions
Scen009	Scen004 with LID/BMPs applied to newly developed areas (no stream PRF or WWTF simulation)
Scen010	Scen008 with efficiency factor for Flow/Runoff increased from 0.2 to 0.4
Scen011	Scen010 with Parker WWTF discharge and loads reverted to the Base conditions
Scen012	Scen011 with TN capped at seasonal targets (6 & 8 mg/L) for all WWTFs
Scen013	Scen012 with efficiency factors for new development applied to all existing development
Scen0XXv2	Scenarios 8-13 with treatment train methodology applied rather than straight load treatment

Processing

SCHEMATIC Development:	...\Modeling\Schematic2030
WWTF Flows/Loads:	...\Schematic2030\WWTP_FlowsLoads.xlsx
Model Files:	...\HSPF_Models\CCW_Scen - contains UCIs, WDMs, and full result files

Landuse Key

1	Dev. Open Space
2	Dev. Low Intensity
3	Dev. Med Intensity
4	Dev. High Intensity
5	Grass/Shrub/Barren/P
6	Agriculture
7	Wetlands
8	Forest

Load and Concentration Results	Base_v2 Model				Scen004 Model				Scen005 Model				Scen006 Model				Scen007 Model				Scen008 Model				Scen009 Model				Scen010 Model				Scen011 Model				Scen012 Model				Scen013 Model											
	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN								
Source	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR
Cherry Creek Surface Flow	14473	2845	9447	43356	27991	15730	21495	80410	17815	3146	10367	65972	31706	16238	22682	106792	31706	14413	21252	102499	28756	5721	15883	95788	25101	6738	15973	73909	25929	5145	14903	91773	23680	5037	14321	75123	23680	5037	14321	75123	23680	5037	14321	75123	23680	5037	14321	75123	23680	5037	14321	75123
Cottonwood Creek Surface Flow	4340	280	839	18568	5195	395	1132	20374	4647	281	853	21561	5503	396	1147	23377	5503	396	1136	23280	5353	354	1084	23166	5046	353	1080	20257	5203	349	1068	23057	5203	349	1068	23057	5203	349	1068	23057	5203	349	1068	23057	4034	180	821	22191				
Other Surface Inflow	679	122	560	3520	903	123	685	4260	679	122	561	3525	935	123	703	4367	935	123	703	4367	906	123	687	4273	873	123	668	4159	852	123	657	4094	830	123	645	4020	830	123	645	4020	830	123	645	4020	422	63	490	3281				
Total Inflow	19491	3247	10846	65444	34090	16249	23312	105043	23141	3549	11781	91058	38144	16757	24532	134535	38144	14932	23092	130146	35015	6198	17655	123227	31019	7214	17721	98326	31984	5617	16628	118924	29713	5509	16034	102200	29713	5509	16034	102200	29713	5509	16034	102200	23629	4173	14015	92447				
FWMC	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L				
Cherry Creek Surface Flow	20.0	145	0.240	1.10	38.7	413	0.282	1.06	24.6	130	0.214	1.36	43.80	377	0.263	1.24	43.80	334	0.246	1.19	39.72	146	0.203	1.22	34.67	197	0.234	1.08	35.81	146	0.211	1.30	32.71	156	0.222	1.17	32.71	156	0.222	1.17	26.48	151	0.244	1.28								
Cottonwood Creek Surface Flow	5.99	47.4	0.071	1.57	7.18	55.9	0.080	1.44	6.42	44.4	0.068	1.71	7.60	53	0.077	1.56	7.60	53	0.076	1.56	7.39	49	0.074	1.59	6.97	51	0.079	1.48	7.19	49	0.076	1.63	7.19	49	0.076	1.63	7.19	49	0.076	1.63	5.57	33	0.075	2.02								
Other Surface Inflow	0.937	133	0.303	1.91	1.248	100	0.279	1.73	0.937	133	0.304	1.91	1.29	97	0.277	1.72	1.29	97	0.277	1.72	1.25	100	0.279	1.73	1.21	104	0.281	1.75	1.18	106	0.284	1.77	1.15	109	0.286	1.78	1.15	109	0.286	1.78	0.58	110	0.427	2.86								
Total Inflow	26.9	123	0.205	1.23	47.1	351	0.251	1.13	32.0	113	0.187	1.45	53	323.1	0.237	1.30	53	287.9	0.223	1.25	48	130.2	0.185	1.29	43	171.0	0.210	1.17	44	129.2	0.191	1.37	41	136.4	0.198	1.26	41	136.4	0.199	1.23	33	129.9	0.218	1.44								

Change Relative to Base_v2	Scen004 - SCH only				Scen005 - WWTF only				Scen006 - SCH & WWTF				Scen007 - SCH, WWTF, & PRF				Scen008 - SCH, WWTF, PRF, & LID				Scen009 - SCH & LID				Scen010 - 008 with Flow eff X 2				Scen011 - 010 w/ Base Parker WWTF				Scen012 - 011 w/ WWTF TN Capped				Scen013 - 012 w/ Eff Fac on all Dev					
	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN		
Source	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ		
Cherry Creek Surface Flow	93	453	128	85	23	11	10	52	119	471	140	146	119	407	125	136	99	101	68	121	73	137	69	70	79	81	58	112	64	77	52	73	64	77	52	66	32	38	34	54		
Cottonwood Creek Surface Flow	20	41	35	10	7	0	2	16	27	42	37	26	27	42	35	25	23	26	29	25	16	26	29	9	20	25	27	24	20	25	27	24	20	25	27	24	-7	-35	-2	20		
Other Surface Inflow	33	1	22	21	0	0	0	0	38	1	26	24	38	1	26	24	34	1	23	21	29	1	19	18	26	1	17	16	22	1	15	14	22	1	15	14	-38	-48	-12	-7		
Total Inflow	75	400	115	61	19	9	39	96	416	126	106	96	360	113	99	80	91	63	88	59	122	63	50	64	73	53	82	52	70	48	56	52	70	48	51	21	29	29	29	41		
FWMC	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ		
Cherry Creek Surface Flow	93	186	18	-4	23	-10	-11	24	119	161	10	12	119	131	3	8	99	1	-15	11	73	37	-3	-2	79	1	-12	18	64	8	-7	6	64	8	-7	1	32	4	2	17		
Cottonwood Creek Surface Flow	20	18	13	-8	7	-6	-5	8	27	12	8	-1	27	12	7	-1	23	2	5	1	16	8	11	-6	20	4	6	4	20	4	6	4	20	4	6	4	6	4	-7	-31	5	29
Other Surface Inflow	33	-24	-8	-9	0	0	0	0	38	-27	-9	-10	38	-27	-9	-10	34	-25	-8	-9	29	-22	-7	-8	26	-20	-6	-7	22	-18	-6	-7	22	-18	-6	-7	-38	-17	41	50		
Total Inflow	75	186	23	-8	19	-8	-9	17	96	164	16	5	96	135	9	2	80	6	-9	5	59	40	3	-6	64	5	-7	11	52	11	-3	2	52	11	-3	-1	21	6	7	17		

Load and Concentration Results	Base_v2 Model				Scen004 Model				Scen005 Model				Scen006 Model				Scen007 Model				Scen008 Model				Scen009 Model				Scen010 Model				Scen011 Model				Scen012 Model				Scen013 Model											
	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN								
Source	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR	AF/YR	TON/YR	LB/YR	LB/YR
Cherry Creek Surface Flow	14473	2845	9447	43356	27991	15730	21495	80410	17815	3146	10367	65972	31706	16238	22682	106792	31706	14413	21252	102499	28756	5552	15261	92211	25101	6570	15323	70119	25929	4805	13694	84859	23680	4697	13141	68397	23680	4697	13141	68397	19173	3278	9967	49777								
Cottonwood Creek Surface Flow	4340	280	839	18568	5195	395	1132	20374	4647	281	853	21561	5503	396	1147	23377	5503	396	1136	23280	5353	346	1053	22956	5046	345	1049	20050	5203	334	1007	22651	5203	334	1007	22651	4034	112	537	20224												
Other Surface Inflow	679	122	560	3520	903	123	685	4260	679	122	561	3525	935	123	703	4367	935	123	703	4367	906	123	687	4273	873	123	668	4159	852	123	657	4094	830	123	645	4020	830	123	645	4020	422	42	310	2000								
Total Inflow	19491	3247	10846	65444	34090	16249	23312	105043	23141	3549	11781	91058	38144	16757	24532	134535	38144	14932	23092	130146	35015	6022	17001	119440	31019	7039	17040	94328	31984	5262	15358	111604	29713	5155	14793	95068	29713	5155	14793	95068	23629	3432	10815	72001								
FWMC	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L	cfs	mg/L	mg/L	mg/L								
Cherry Creek Surface Flow	20.0	145	0.240	1.10	38.7	413	0.282	1.06	24.6	130	0.214	1.36	43.80	377	0.263	1.24	43.80	334	0.246	1.19	39.72	142	0.195	1.18	34.67	193	0.224	1.03	35.81	136	0.194	1.20	32.71	146	0.204	1.06	32.71	146	0.204	1.01	26.48	126	0.191	0.95								
Cottonwood Creek Surface Flow	5.99	47.4	0.071	1.57	7.18	55.9	0.080	1.44	6.42	44.4	0.068	1.71	7.60	53	0.077	1.56	7.60	53	0.076	1.56	7.39	48	0.072	1.58	6.97	50	0.076	1.46	7.19	47	0.071	1.60	7.19	47	0.071	1.60	5.57	20	0.049	1.84												
Other Surface Inflow	0.937	133	0.303	1.91	1.248	100	0.279	1.73	0.937	133	0.304	1.91	1.29	97	0.277	1.72	1.29	97	0.277	1.72	1.25	100	0.279	1.73	1.21	104	0.281	1.75	1.18	106	0.284	1.77	1.15	109	0.286	1.78	1.15	109	0.286	1.78	0.58	73	0.271	1.74								
Total Inflow	26.9	123	0.205	1.23	47.1	351	0.251	1.13	32.0	113	0.187	1.45	53	323.1	0.237	1.30	53	287.9	0.223	1.25	48	126.5	0.179	1.25	43	166.9	0.202	1.12	44	121.0	0.177	1.28	41	127.6	0.183	1.18	41	127.6	0.183	1.14	33	106.8	0.168	1.12								

Change Relative to Base_v2	Scen004 - SCH only				Scen005 - WWTF only				Scen006 - SCH & WWTF				Scen007 - SCH, WWTF, & PRF				Scen008 - SCH, WWTF, PRF, & LID				Scen009 - SCH & LID				Scen010 - 008 with Flow eff X 2				Scen011 - 010 w/ Base Parker WWTF				Scen012 - 011 w/ WWTF TN Capped				Scen013 - 012 w/ Eff Fac on all Dev			
	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN	Flow	TSS	TP	TN
Source	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ
Cherry Creek Surface Flow	93	453	128	85	23	11	10	52	119	471	140	146	119	407	125	136	99	95	62	113	73	131	62	82	79	69	45	96	64	65	39	58	64	65	39	58	32	15	6	15
Cottonwood Creek Surface Flow	20	41	35	10	7	0	2	16	27	42	37	26	27	42	35	25	23	24	25	24	16	23	25	8	20	19	20	22	20	19	20	22	20	19	20	22	-7	-60	-36	9
Other Surface Inflow	33	1	22	21	0	0	0	0	38	1	26	24	38	1	26	24	34	1	23	21	29	1	19	18	26	1	17	16	22	1	15	14	22	1	15	14	-38	-66	-45	-43
Total Inflow	75	400	115	61	19	9	39	96	416	126	106	96	360	113	99	80	85	57	83	59	117	57	44	64	62	42	71	52	59	36	45	52	59	36	45	52	21	6	0	10
FWMC	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ	% Δ
Cherry Creek Surface Flow	93	186	18	-4	23	-10	-11	24	119	161	10	12	119	131	3	8	99	-2	-19	7	73	33	-6	-7	79	-6	-19	9	64	1	-15	-4	64	1	-15	-8	32	-13	-20	-13
Cottonwood Creek Surface Flow	20	18	13	-8	7	-6	-5	8	27	12	8	-1	27	12	7	-1	23	0	2	0	16	6	7	-7	20	0	0	2	20	0	0	2	20	0	0	2	-7	-57	-31	17
Other Surface Inflow	33	-24	-8	-9	0	0	0	0	38	-27	-9	-10	38	-27	-9	-10	34	-25	-8	-9	29	-22	-7	-8	26	-20	-6	-7	22	-18	-6	-7	22	-18	-6	-7	-38	-45	-11	-9
Total Inflow	75	186	23	-8	19	-8	-9	17	96	164	16	5	96	135	9	2	80	3	-13	2	59	36	-1	-9	64	-1	-14	4	52	4	-11	-5	52	4	-10	-8	21	-13	-18	-9



DRAFT

January 5, 2024

Katie Seefus
Water Control and Water Quality Section
Hydrologic Engineering Branch
US Army Corps of Engineers

Subject: Cherry Creek Low-Level Release Sustainable Rivers Program Proposal

Ms. Seefus:

The Cherry Creek Basin Water Quality Authority (CCBWQA) is pleased to support the Cherry Creek Low-Level Release Sustainable Rivers Program Proposal prepared by the US Army Corps of Engineers (USACE). CCBWQA's reservoir model has shown that both watershed controls and in-reservoir management are needed to reduce nutrient loading to and within the reservoir and associated algal blooms that cause the reservoir to frequently exceed the chlorophyll-a standard assigned by Colorado Department of Public Health and Environment. We appreciate USACE's interest in exploring innovative reservoir management approaches to manage conditions that cause internal phosphorus loading from reservoir sediments.

We look forward to collaborating with you on this project. CCBWQA has long-term water quality monitoring data that we would be happy to share to support the project. Additionally, CCBWQA's reservoir model could be a tool to explore potential response to various release strategies. We would also like to coordinate sampling of the reservoir releases under these operations for nutrients and other water quality indicators (e.g., DO, pH, conductivity, temperature), particularly in the context of downstream water users.

If you have any questions regarding the Cherry Creek Basin Water Quality Authority's support or participation, please contact me.

Sincerely,

Cherry Creek Basin Water Quality Authority

By 
Jane Clary, Technical Manager
www.cherrycreekbasin.org
jclary@wrightwater.com